

APPENDIX

REVISED WATER MODELLING REPORT

Modelling report

Annexure I to Main Works Preferred Infrastructure Report and Response to Submissions

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Executive Summary

ES1 Overview

Snowy Hydro Limited (Snowy Hydro) proposes to develop Snowy 2.0, a large-scale pumped hydro-electric storage and generation project which would increase hydro-electric capacity within the existing Snowy Mountains Hydro-electric Scheme (Snowy Scheme). Snowy 2.0 is the largest committed renewable energy project in Australia and is critical to underpinning system security and reliability as Australia transitions to a decarbonised economy. Snowy 2.0 will link the existing Tantangara and Talbingo reservoirs within the Snowy Scheme through a series of underground tunnels and a new hydro-electric power station will be built underground.

A single EIS was prepared to address the requirements set out by the NSW Department of Planning, Industry and Environment (DPIE) and the Commonwealth Department of the Environment and Energy (DEE). In accordance with the EP&A Act and *Environmental Planning and Assessment Regulation 2000* (EP&A Regulation), the EIS was placed on public exhibition for a period of 42 days, between 26 September 2019 and 6 November 2019.

A total of 201 submissions were received during the public exhibition period, including 30 from special interest groups and 161 individual community submitters. In addition, ten submissions were received from State government agencies and councils. Of the 201 submissions, 5% were in support of the Main Works, 73% objected to the works, and the remaining submissions provided comments (22%).

This revised Modelling Report does not directly discuss submissions or responses but does present revised or additional information to support the responses provided in the Preferred Infrastructure Report and Response to Submissions (PIR-RTS). This report details groundwater and surface water modelling undertaken to describe the impacts of the proposed Snowy 2.0 project on:

- groundwater head and drawdown in the vicinity of the project;
- groundwater inflow rates to the various tunnels and excavations;
- the baseflow component of streamflow; and
- overall streamflow statistics within the project area.

Modelling of the groundwater flow system was undertaken using MODFLOW-USG via the Groundwater Vistas graphical user interface, while modelling of the surface water system was undertaken using eWater Source. These two models were loosely coupled, with data transfer occurring during model calibration and scenario modelling phases.

ES2 Peer review

A pre-eminent hydrogeologist, Hugh Middlemis, was engaged to peer review the numerical groundwater model and coupled surface water model.

The peer reviewer deemed that:

- the catchment model was been prepared in a manner consistent with best practice surface water modelling guidelines published by eWater (Black, et al., 2011);
- the groundwater model was developed in accordance with the principles of the best practice Australian Groundwater Modelling Guideline (Barnett, et al., 2012); and

• the coupled models are fit for the purpose of assessing catchment water balance impacts, and to inform management strategies and licensing.

ES3 Groundwater

ES3.1 Model setup

The groundwater model domain encompasses all underground excavations of the Snowy 2.0 project, Yarrangobilly Caves, all major rivers and creeks as well as all project-related groundwater monitoring sites. Hydrostratigraphic units were assigned to the model for each of the geological units mapped by drilling and geophysical surveys along the project alignment and the model design considered the water balance reporting required as per the NSW Aquifer Interference Policy (AIP), NSW DPI Water, 2012.

As the project is located within Kosciuszko National Park, with no existing suitable groundwater monitoring network or third party groundwater supply wells, no groundwater monitoring data were available within or near the model domain prior to the proposal of Snowy 2.0. As part of the water assessment for the project a network of bores was drilled, largely along the project alignment. The earliest groundwater monitoring data date from late November 2017 and many of the monitoring sites have significantly shorter records. Whilst the data do span a full set of seasons, the monitoring of long-term seasonality of groundwater behaviour is limited but is only one month short of the guideline duration for baseline of two years (DPI Water, 2014) and the response to prolonged wet or dry periods was not available to inform calibration of the model.

The combination of hydraulic head measurements, baseflow calculations and hydraulic property testing were used to inform calibration of the model in both steady state and transient modes. Calibration achieved a scaled RMS statistic of 3.6% for the steady state model, and 3.9% for the transient model.

ES3.2 Scenarios modelled

One project scenario was modelled, representing:

- the pre-construction groundwater system;
- construction of the project, with model boundary conditions added in accordance with the project tunnel design and schedule, considering wet, dry and average climate sequences;
- a 20-year operation period; and
- post-construction steady state groundwater conditions representing long-term stable conditions.

Tunnel inflow mitigation measures were not included in the model scenario presented in this document. Sensitivity of predicted impacts to modelled tunnel inflows was addressed by three additional simulations for "Low inflow", "High inflow" and one at the "Maximum capacity" of the tunnel boring machines (TBMs).

Climate change was not explicitly modelled. Tests utilising wet and dry climate sequences indicated that groundwater inflow rates to the excavations are insensitive to climate.

ES3.3 Model conservatism

The model predictions are considered conservative due to the design scenario assumptions (unmitigated) and the adoption of conservative hydraulic parameters (as per field measurements). Therefore, it is considered that the actual inflow (and subsequent impacts) will be lower than predicted due to mitigation and management measures committed to during construction (ie pre-grouting and post-grouting of key areas).

Groundwater flow into the tunnel is expected to occur primarily as a function of secondary porosity (ie via fractures and along bedding planes). The groundwater model assumes significant connection between the tunnel and the watertable in the Gooandra Volcanics and the Kellys Plains Volcanics due to the hydraulic testing undertaken throughout the unit. It is possible that additional field testing may reveal that locations with vertical connection occur only in isolated locations.

The model cannot simulate individual fractures because the locations, conductivity and connectivity of individual fractures will not be known until the tunnel intersects them. Because the exact locations and extent of inflow mitigation strategies are not yet known, the groundwater modelling adopted a conservative approach of simulating all excavations as non-mitigated/controlled. The modelling results are therefore conservative for two reasons:

- modelling does not consider mitigating activities:
 - conservative as during construction the discrete fractures that yield excess water will be grouted and will reduce the actual overall tunnel inflow volume;
- hydraulic parameter values adopted in the numerical model for the Gooandra Volcanics and the Kellys Plain Volcanics assume significant connection to the watertable based on limited pumping test data:
 - potentially conservative as the entirety of the Gooandra Volcanics and the Kellys Plain Volcanics may not behave like this, with some parts being less permeable or less connected.

Therefore, the model predictions of tunnel inflow, baseflow reduction and watertable drawdown are likely to be over estimating project impacts. The results of this conservative model approach need to be considered within this overall context to accurately assess the project on its true merits for impacts to water resources.

ES3.4 Results

The groundwater model predicted that groundwater would flow into the project excavations during construction, and into the power waterway during operation. The total inflow to excavations is expected to peak at 62 L/s (2 GL/year) in the final year of construction, and reducing to 45 L/s (1.4 GL/year) during operation (Figure ES1). Inflows to excavations will be from groundwater sourced from the fractured rock groundwater sources of the Lachlan Fold Belt Murray Darling Basin (MDB) Groundwater Source and the Lachlan Fold Belt South Coast Groundwater Source.

Groundwater flow to the excavations and power waterway will result in groundwater hydraulic head drawdown developing over time. Groundwater drawdown of the watertable is predicted to occur primarily near the Tantangara adit, and in the vicinity of the Gooandra Volcanics geological unit (near Gooandra Creek and the Snowy Mountains Highway) (Figure ES3). The model also predicts scattered pockets of watertable drawdown within the Yarrangobilly River catchment. No change in groundwater level was predicted at the Yarrangobilly Caves.

As a result of watertable drawdown, rates of groundwater discharge to surface water features (ie groundwater available for baseflow) are predicted to decline within some river and creek catchments in the vicinity of the tunnel alignment. Although streams remain gaining (continue to receive groundwater baseflow), a reduction in baseflow is expected to develop over time, with the peak impact being realised a number of decades after the completion of the project (Figure ES2). Long-term peak baseflow reductions are predicted to approximately match the long-term inflow rate to the power waterway. Upper Tumut, Murrumbidgee and Lake Eucumbene surface water sources are each predicted to receive less baseflow due to inflows to the power waterway, with the largest impact occurring within the Murrumbidgee (Gooandra Creek) catchment, followed by the Eucumbene River headwaters (Figure ES2).

The hydraulic conductivity of the rock to be excavated by the project has been estimated using appropriate hydrogeological techniques and pumping test methods. However, groundwater flow in fractured rock is highly heterogeneous and local scale and overall groundwater inflow to excavations will only be known once the project

commences and groundwater flows into the tunnel are measured. Until that time, the groundwater drawdown and baseflow reduction predictions of the groundwater model will carry a degree of uncertainty. Should the hydraulic conductivity of the rock be higher than modelled (ie there are more or larger fractures encountered than anticipated), then impacts to creeks at the surface may be larger than estimated. This could take the form of more severe impacts within creeks already predicted to be impacted, or it could take the form of impacts to creeks previously estimated to be unaffected by the project. Conversely, if fewer fractures are encountered, or if these fractures are not regionally connected, or if mitigation measures are applied (which is planned) then the estimated tunnel inflows may be significantly reduced, and reduction to baseflow in creeks would be lower.

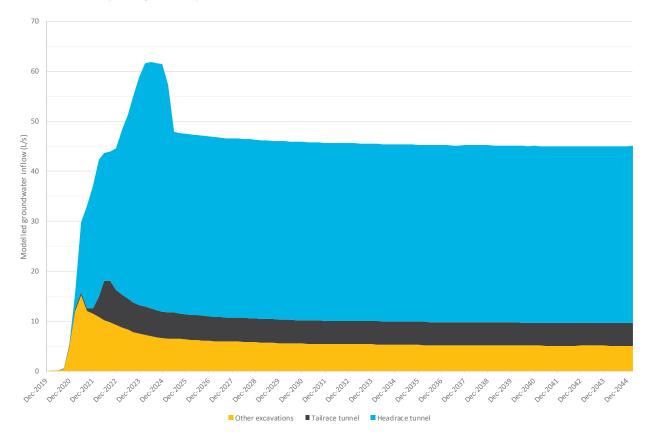


Figure ES1 Predicted inflow to excavations during construction and operation

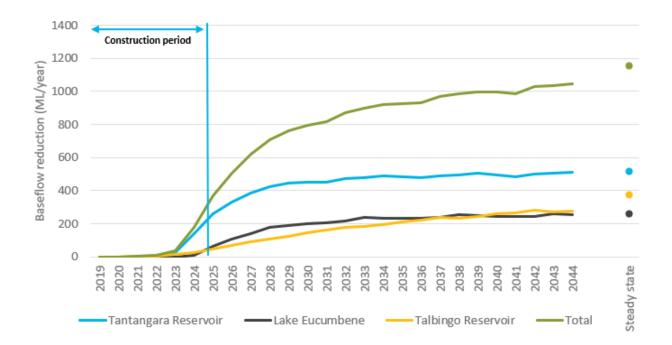


Figure ES2 Predicted baseflow reduction

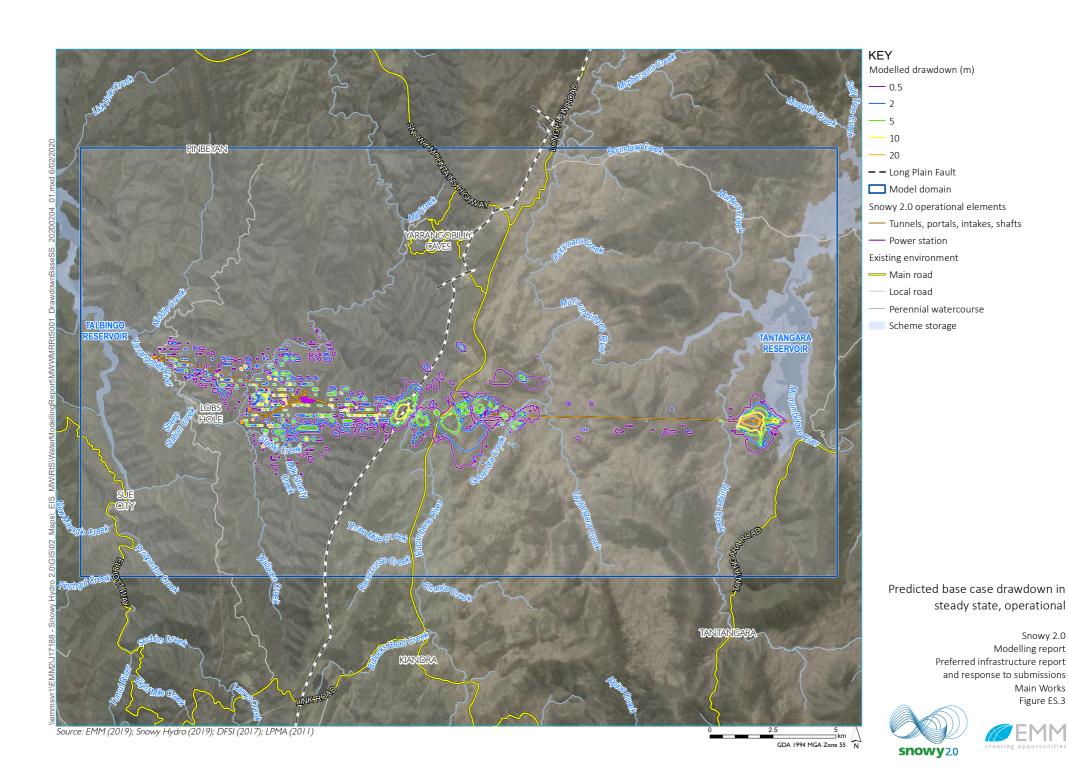


Table ES1 Predicted steady state (long term) baseflow impacts

Surface water management unit	Catchment	Baseflow Reduction (ML/yr)	Total (ML/yr)
Murrumbidgee	Murrumbidgee River	477	
	(including Gooandra Creek and Tantangara Creek)		518
	Nungar Creek	41	
Lake Eucumbene	Eucumbene River	258	258
Upper Tumut	Yarrangobilly River	372	
	(including Wallaces Creek and Stable Creek)		375
	Middle Creek	3	
Total			1,151

ES4 Surface water

ES4.1 Model setup

The surface water catchment model extent covered the Murrumbidgee River upstream of the Tantangara Reservoir, the Yarrangobilly River upstream of the Talbingo Reservoir, the Eucumbene River within the groundwater model domain extent, Nungar Creek and Middle Creek (Figure ES5). This extent included the area where groundwater drawdown was predicted to reach the surface.

The model was calibrated using approximately 40 years' daily streamflow data at gauges 410535 and 410574 located on the Murrumbidgee and Yarrangobilly rivers. Model validation was undertaken using streamflow data collected at several locations across the plateau, via manual and automated gauging.

ES4.2 Model coupling

The catchment water balance and runoff model utilised by the surface water model is illustrated in Figure ES4, in which a number of processes relating to runoff, infiltration, evapotranspiration and streamflow are illustrated. This runoff model is a modified version of the SIMHYD runoff model. Alterations to the standard SIMHYD model were made to enable better representation of interflow and groundwater recharge processes occurring at the project site and resulted in improved model calibration.

Processes modelled by the groundwater model are included within the green box labelled 'Groundwater Model' in Figure ES4. Recharge estimates produced by the surface water model were utilised by the groundwater model as an input. Each model produced independent estimates of baseflow discharge, on different time scales, and the calibration process ensured consistency.

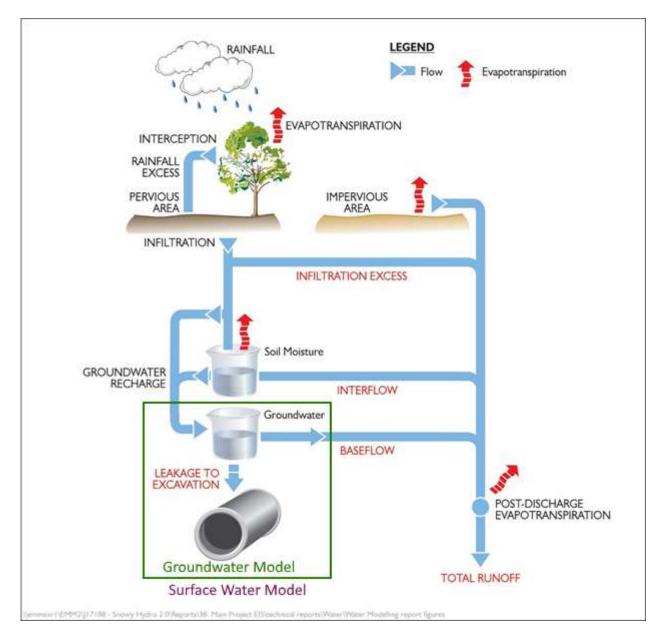
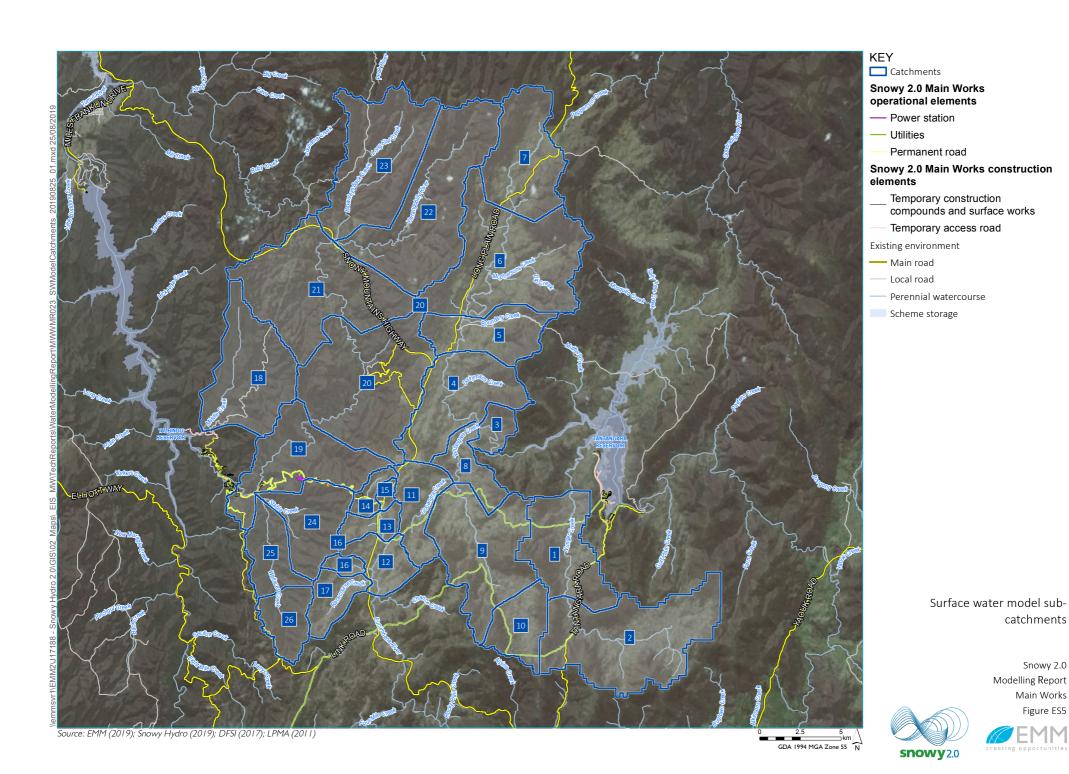


Figure ES4 Processes modelled using the surface and groundwater models

Baseflow reductions predicted by the groundwater model were applied in the surface water model using a 'leakage' term in the Modified SIMHYD rainfall runoff model. This term caused the model groundwater store to empty at a faster rate and resulted in reduced baseflow. Leakage rates were only applied within model subcatchments substantially affected by groundwater drawdown.



ES4.3 Scenarios modelled

As per the groundwater modelling, one scenario was modelled, representing the following project phases:

- the pre-construction surface water system;
- construction of the project considering wet, dry and average climate sequences; and
- operation of the project (ie post-construction steady state groundwater conditions).

Climate change was not explicitly modelled. Sensitivity analysis indicated that runoff statistics are sensitive to changes in rainfall, but that the change to runoff statistics due to project impacts is relatively insensitive to changes in rainfall or evapotranspiration.

ES4.4 Results

The groundwater model predicted that impacts to creek and river baseflows would develop over time, with the largest impacts seen after construction is complete, and showed that groundwater drawdown at the surface will mainly occur in the vicinity of Gooandra Creek and the Eucumbene River headwaters (Figure ES3). The groundwater model predicted that during construction and in the areas directly overlying the tunnel alignment:

- baseflow to Gooandra Creek may decline by up to 6%, beginning in year 4 of construction; and
- baseflow to the Eucumbene River may decline by up to 1%, beginning in year 5 of construction, with impacts centered on the uppermost 1.5 km of the Eucumbene River headwaters.

The surface water catchment model was used to investigate the effect of these baseflow reductions on the streamflow regimes downstream of the impacted catchments, and showed that:

- Gooandra Creek is likely to change from a perennial streamflow regime to ephemeral, as days with less than 0.1 ML/day streamflow at the downstream end of the creek increase from 0% to 2%; and
- north of the Snowy Highway the Eucumbene River could also become ephemeral, as days with less than 0.1 ML/day streamflow at this location increase from 0% to approximately 5–7%.

It is expected that the quickflow component of streamflow (surface runoff in response to rainfall) will not be affected by groundwater drawdown and baseflow reduction.

In each catchment, the modelled impact reduced with distance downstream as flows from catchment areas unaffected by the project entered the creek system.

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1 Introduction

1.1 Overview

Snowy Hydro Limited (Snowy Hydro) proposes to develop Snowy 2.0, a large-scale pumped hydro-electric storage and generation project which would increase hydro-electric capacity within the existing Snowy Mountains Hydro-electric Scheme (Snowy Scheme). Snowy 2.0 is the largest committed renewable energy project in Australia and is critical to underpinning system security and reliability as Australia transitions to a decarbonised economy.

Snowy 2.0 will link the existing Tantangara and Talbingo reservoirs within the Snowy Scheme through a series of underground tunnels and a new hydro-electric power station will be built underground. The major construction elements of Snowy 2.0 include permanent infrastructure, temporary construction infrastructure, management and storage of excavated rock material and establishing supporting infrastructure. Snowy 2.0 Main Works also includes the operation of Snowy 2.0.

A single EIS was prepared to address the requirements set out by the NSW Department of Planning, Industry and Environment (DPIE) and the Commonwealth Department of the Environment and Energy (DEE). In accordance with the EP&A Act and Environmental Planning and Assessment Regulation 2000 (EP&A Regulation), the EIS was placed on public exhibition for a period of 42 days, between 26 September 2019 and 6 November 2019.

In order to assess potential groundwater and surface water related issues from the construction and operation of Snowy 2.0, a water assessment (EMM 2019) was prepared as an appendix to the Snowy 2.0 Main Works EIS. The water assessment has a number of supporting technical reports which are termed annexures. Each annexure has further supporting technical reports which are termed attachments. The EIS modelling report was an annexure to the EIS water assessment. The document structure of the technical reports and assessments which supported the overall EIS water assessment are shown in Figure 1.1.

It is noted that due to project refinements and submissions, the Modelling Report was required to be updated. This revised Modelling report is therefore Appendix I to the Preferred Infrastructure Report and Response to Submissions (PIR-RTS), but still fits into the broad structure of the EIS water assessment detailed in Figure 1.1.

1.1.1 Key project refinements since public exhibition

Snowy Hydro and its appointed contractor, Future Generation Joint Venture (FGJV), continue to refine and improve the design for Snowy 2.0 as information is obtained from the geotechnical investigation program and Exploratory Works. In addition, matters raised by agencies and stakeholders during public exhibition of the Main Works EIS has necessitated refinements to key elements of the project. These are described in more detail in the PIR-RTS, however include:

- considerable refinement of the disturbance area;
- reduced traffic volumes;
- refinement of the groundwater model to better represent the inflow mitigation that will occur from the segmental concrete lining of the power waterway; and
- alternative options for management of excavated rock.

1.1.2 Key submissions related to water

Of the 201 submissions received, there were 4 key water related submissions from government agencies and 6 submissions from special interest groups with water related themes. A number of public submissions were also received with issues raised related to water related impacts as a result of Snowy 2.0 Main Works.

The PIR-RTS details the key themes of the submissions and directly addresses those themes. In addition, detailed responses to key government agency submissions are included as appendices to the PIR-RTS.

This revised Modelling Report does not directly discuss submissions or responses but does present revised or additional information to support the responses provided in the PIR-RTS.

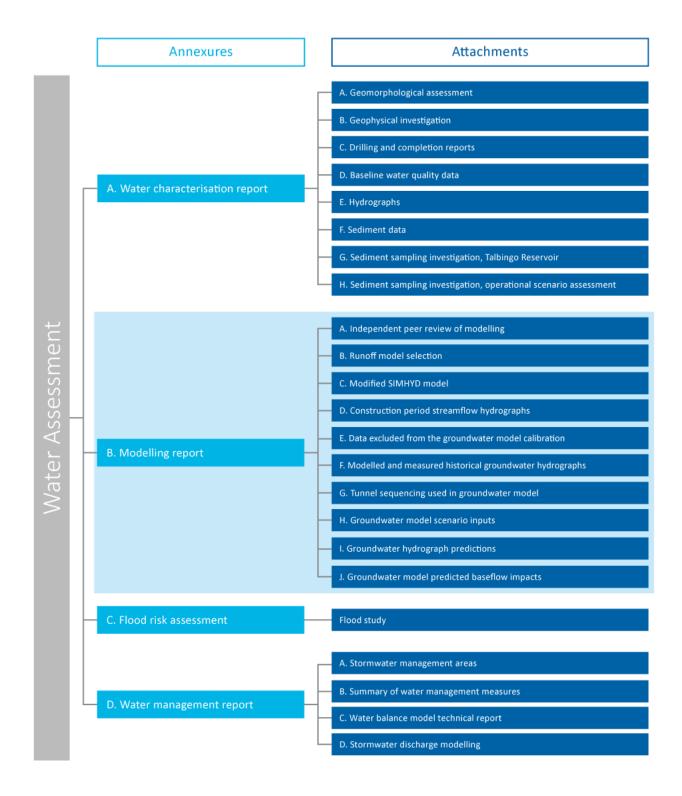


Figure 1.1 EIS Water Assessment document structure

1.2 Scope of this report

This report describes groundwater and surface water modelling undertaken to describe the impacts of the proposed Snowy 2.0 project on:

- groundwater head and drawdown in the vicinity of the project;
- groundwater inflow rates to the various tunnels and excavations;
- the baseflow component of streamflow; and
- overall streamflow statistics within the project area.

1.3 Modelling approach

Modelling of the groundwater flow system was undertaken using MODFLOW-USG via the Groundwater Vistas graphical user interface, while modelling of the surface water system was undertaken using eWater Source. These two models were loosely coupled (ie using a methodology similar to that applied to the Murray-Darling Basin Sustainable Yields project; CSIRO 2007, 2008), with data transfer occurring during model calibration and scenario modelling phases (Figure 1.2).

The surface water model (see chapter 2) was developed using a rainfall runoff model that explicitly described the movement of water through the soil unsaturated zone into the aquifer, and discharge of groundwater as baseflow to streams. Seasonal groundwater recharge rates estimated by the calibrated surface water model were provided as input to the groundwater model (see chapter 3) such that the two models utilised a consistent catchment water balance.

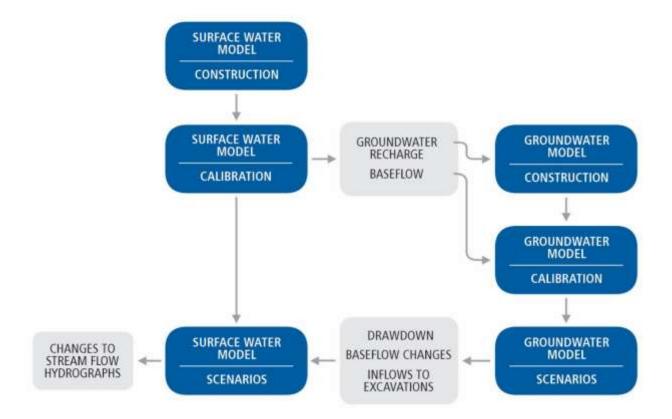


Figure 1.2 Data transfer between model stages

Rainfall interception, evapotranspiration prior to groundwater recharge, and evapotranspiration post baseflow discharge were modelled within the surface water model. The total evapotranspiration flux extracted from within the surface water model is consistent with the calibrated whole of catchment water balance. The recharge rates provided to the groundwater model were a net rate post evapotranspiration within the soil and root zone.

To avoid double accounting for evapotranspiration, the groundwater model was not used to directly estimate additional evapotranspiration from saturated groundwater (although the evapotranspiration package was applied in the groundwater model to estimate distributed/diffuse groundwater discharge to the surface and near surface; see chapter 3 for details). The groundwater model was used to model the movements of water underground from recharge to baseflow discharge locations, with a focus on estimating flow into the excavated tunnels and caverns.

The catchment water balance utilised by the surface water model is illustrated in Figure 1.3, in which a number of processes relating to runoff, infiltration, evapotranspiration and streamflow are illustrated. Processes modelled by the groundwater model are included within the green box labelled 'Groundwater Model'. The recharge and flow to excavation rates were each the result of one model, used as an input to the other (Table 1.1). Each model produced independent estimates of baseflow discharge, on different time scales, and the calibration process ensured consistency. The quarterly sum of baseflow discharges predicted by the surface water model was compared to the baseflow predictions made by the groundwater model during the groundwater model calibration process (see section 3.2.7). The daily baseflow predictions produced by the surface water model were used for developing descriptions of changes to streamflow and related statistical measures.

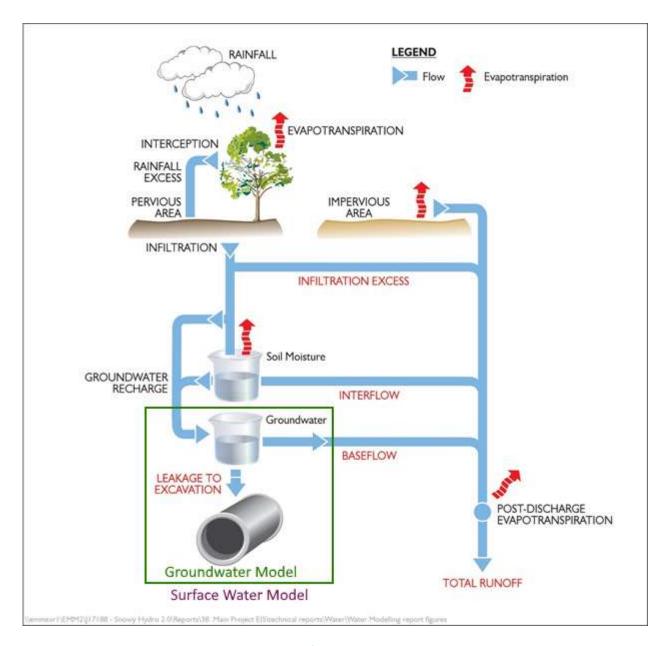


Figure 1.3 Processes modelled using the surface and groundwater models

Table 1.1 Processes modelled by the coupled groundwater and surface water models

Catchment Process	Status in surface water model	Status in groundwater model
Groundwater recharge	Result	Input
Leakage to excavation	Input	Result
Baseflow discharge	Result	Result

6

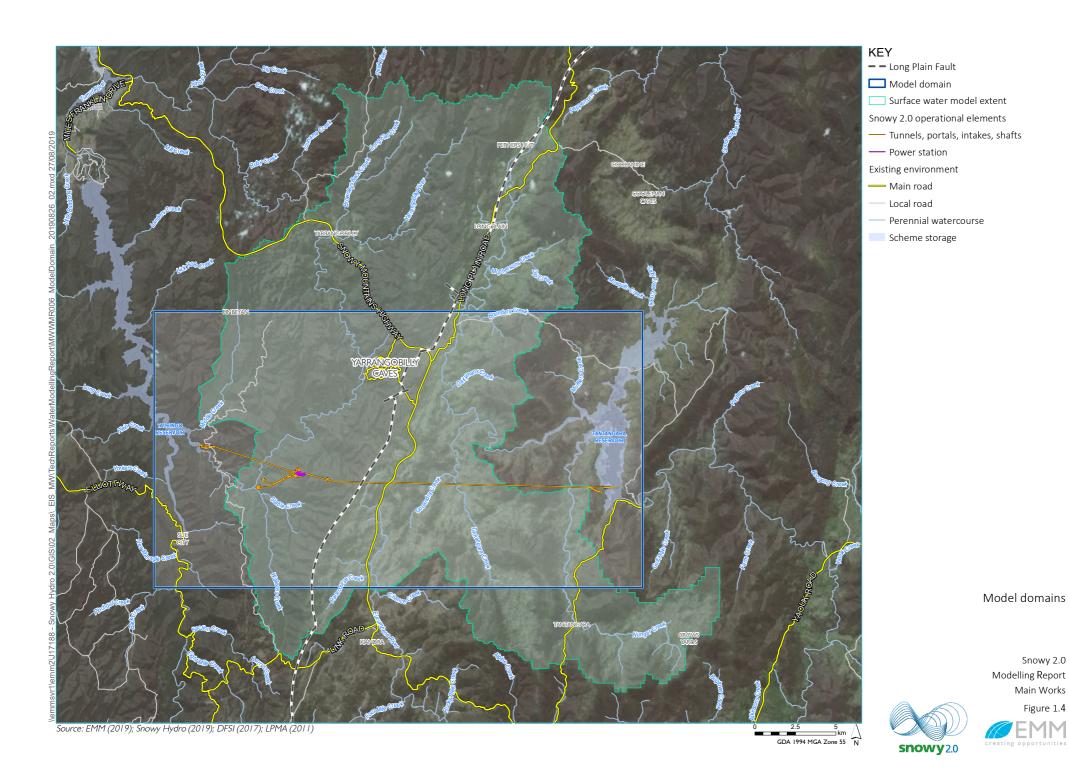
Baseflow discharge from the groundwater model was estimated via two methods:

- discharge directly to the base of creeks and rivers; and
- distributed groundwater discharge in catchment headwaters.

Discharge directly to the base of creeks and rivers was modelled using river boundary conditions along mapped creek alignments in the groundwater model. The concept of distributed baseflow discharge was utilised to model discharges at springs located away from the creek and river alignment, for example on flood plains or at break of slope on hill sides. Distributed baseflow discharge was modelled using the built-in evapotranspiration package within MODFLOW, as this allows water to be removed from the model wherever it nears or reaches the surface.

The groundwater model extent was a 30 km by 17 km rectangle centred on the power waterway location (section 3.2.2). The surface water model extent encompassed the headwaters of the Murrumbidgee River, Yarrangobilly River, Eucumbene River, and the entirety of the Nungar Creek and Middle Creek catchments, extending further north and south than the groundwater model extent. A comparison of model extents is provided in Figure 1.4.

The groundwater model produced predictions of inflow to the excavations during construction and flow exchanges during operations, as well as the induced impact on surface water features (section 3.4) using a quarterly (seasonal) time period. These impacts were temporally disaggregated using the surface water model, which ran with a daily time step. Changes to streamflow hydrographs were assessed using the surface water model and are described in section 2.7.



2 Surface Water

2.1 Surface water modelling overview

2.1.1 Catchment model purpose

Catchment scale water balance and rainfall-runoff modelling was undertaken for the surface water catchments in the vicinity of the tunnel alignment for two purposes.

Firstly, to develop a catchment scale daily water balance consistent with measured streamflow data and the hydrological concept of the area that includes surface runoff, groundwater recharge, and discharge flow processes. Quarterly groundwater recharge rates post evapotranspiration from soil were taken from this water balance and utilised by the groundwater model as an input.

Secondly, to develop a framework in which project impacts to surface water flows might be assessed. The model was set up such that changes to baseflow due to tunnelling (an output from the groundwater model), or discharges of excess water, might be modelled and compared to unaffected runoff. Streamflow hydrographs and changes to streamflow statistics were then provided to project ecologists as an input to the ecological impact assessment.

The development of recharge estimates alleviated the potential for non-uniqueness within the groundwater model parameter set, narrowing the range of possible aquifer property values.

2.1.2 Peer review

A pre-eminent hydrogeologist, Hugh Middlemis, was engaged to peer review the numerical groundwater model and coupled surface water model.

The peer reviewer deemed that:

- the catchment model has been prepared in a manner consistent with best practice surface water modelling guidelines; and
- the coupled models are fit for the purpose of assessing catchment water balance impacts, and to inform management strategies and licensing.

The peer review report (Middlemis, August 2019) is included in Attachment A.

2.2 Baseflow component of streamflow

The contribution of groundwater discharge (baseflow) to the Murrumbidgee River and Yarrangobilly River was estimated by several methods prior to commencing catchment modelling:

- 1. automated baseflow separation using the Lyne and Hollick digital filter;
- 2. analysis of groundwater monitoring bore data during streamflow recession; and
- 3. analysis of stream and groundwater salinity.

Each of these are discussed below.

2.2.1 Recession analysis

As a precursor to applying automated methods to separate the baseflow component from the streamflow records, recession analysis of the streamflow data was undertaken. It was assumed that baseflow recession would follow the relationship shown in Equation 2.1.

Equation 2.1 Recession equation

$$Q = Q_0 k^t$$

Where

- Q is the flow on a particular day t days after the recession began
- Q_0 is the initial flow when t=0
- *k* is the recession constant

Results of the analysis are presented in Figure 2.1 to Figure 2.3 for the Murrumbidgee (gauge site 410535), Eucumbene (gauge site 22522) and Yarrangobilly (gauge site 410574) rivers, in which it can be seen that:

- when the streamflow in each river is above 1 kL/s, the recession constant k is around 0.925; and
- when the streamflow in each river is below 1 kL/s a higher k value fits the data better, with k = 0.95 0.98 fitting well for each river during the end of summer low flow period.

Note that the Eucumbene River gauge 22522 is outside the domain of the surface water model described in this report, and streamflow at this gauge was not used for model calibration.

The low flow recession rate at the Wallaces Creek gauge was around 0.925 as presented in Figure 2.4.

It is likely that during the end of summer low flow period baseflow occurs only from groundwater discharge sources. During wetter months when river flows are higher, the streamflow recession may be influenced by other factors in addition to groundwater discharge, such as snow melt and interflow processes.



Figure 2.1 Murrumbidgee River recession



Figure 2.2 Eucumbene River recession

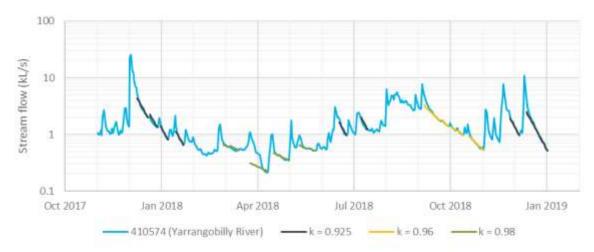


Figure 2.3 Yarrangobilly River recession

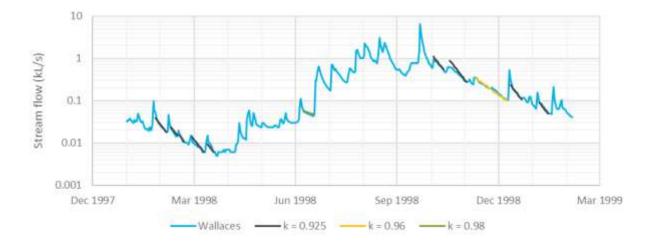


Figure 2.4 Wallaces Creek recession

2.2.2 Baseflow separation using the Lyne and Hollick filter

Digital filtering is a repeatable automated mathematical method of separating quickflow and baseflow using streamflow hydrograph data alone. Rapid rises and subsequent recessions in the hydrograph are located and apportioned as quickflow, while the remainder of the flow is apportioned as baseflow. A number of digital filtering methods were assessed for use in Australia by the *Australian Rainfall and Runoff Revision Project 7* team, as discussed in Murphy et al (2009), with the Lyne and Hollick filter (Equation 2.2) recommended for use across Australia in the release of *Australian Rainfall and Runoff* (2016) as it produces plausible splits between quickflow and baseflow for all daily streamflow datasets.

Equation 2.2 Lyne and Hollick filter equation

$$Q_f(i) = k.\,Q_f(i-1) + \frac{(Q(i)-Q(i-1)).\,(1+k)}{2}$$

Where

- $Q_f(i) \ge 0$
- $Q_b(i)=Q(i)-Q_f(i)$
- Q_b(i) is the filtered baseflow at day i
- Q_f(i) is the filtered quickflow at day i
- Q(i) is the original streamflow at day i
- k is the filter parameter, sometimes labelled α , equivalent to the recession constant

As the Lyne and Hollick filter uses 'blind' frequency filtering mathematics and takes no catchment specific or climatic inputs (it uses streamflow records alone), it cannot distinguish between groundwater discharge, snow melt, and other forms of slow water release. In locations such as the project site where alluvial materials adjacent to creeks, bogs, and snow may each detain runoff, the rate of groundwater discharge may be lower than the 'baseflow' reported by this method.

The Lyne and Hollick filter requires the filter parameter k to be specified. This parameter influences the nature of attenuation of the streamflow hydrograph and thus the percentage of streamflow predicted to be from baseflow. A parameter value of 0.925 is generally accepted as appropriate (Nathan & McMahon, 1990), providing similar results to manual baseflow separation methods for catchments in NSW and Victoria, though a value of 0.98 has been found in some cases to provide a better result for some Murray-Darling Basin catchments (Ladson, Brown, Neal, & Nathan, 2013).

Baseflow analysis results for the range of k values between 0.92 and 0.98 are presented in Table 2.1 for streamflow recorded at the Murrumbidgee River gauge (410535) and Table 2.2 for the Yarrangobilly River gauge (410574) following analysis using streamflow data from the beginning of the data record until the end of 2018, and for the Wallaces Creek gauge (Table 2.3) using streamflow data from 1982 to 1999.

Regression analysis using rainfall and the filtered baseflow was utilised to select the optimum k values for each catchment. It was found that the best fit parameter value for the Murrumbidgee River was in the range 0.94–0.95, giving a baseflow index estimate of around **53%.** For the Yarrangobilly River the best fit k value was 0.935, giving a baseflow index estimate of **56%**. The best fit k value for the Wallaces Creek catchment was 0.97, giving a baseflow index estimate of **38%**.

The recession analysis (section 2.2.1) indicated that a k value of 0.98 may be appropriate for separating the groundwater discharge component of baseflow for both the Murrumbidgee River and Yarrangobilly River. When utilising a k of 0.98, the baseflow index for the Murrumbidgee River reduces to **38%.** Similarly, the baseflow index for the Yarrangobilly River reduces to **41%** when using a k of 0.98.

Within Wallaces Creek, the summer recession k of 0.925 results in a baseflow index of 53%.

Table 2.1 Regression results for Murrumbidgee River Lyne and Hollick baseflow separation

k	R ²	Trend	BFI
0.925	82.3%	-0.1%	58%
0.93	82.4%	-0.1%	57%
0.935	82.6%	-0.1%	56%
0.94	82.7%	0%	55%
0.945	82.7%	0%	54%
0.95	82.7%	0%	52%
0.955	82.4%	0%	51%
0.96	81.6%	0%	49%
0.965	80.7%	0.1%	47%
0.97	78.6%	0.1%	44%
0.975	75.0%	0.1%	41%
0.98	68.7%	0.2%	38%

 Table 2.2
 Regression results for Yarrangobilly River Lyne and Hollick baseflow separation

k	R ²	Trend	BFI
0.925	72.60%	-1.0%	58%
0.93	72.60%	-1.0%	57%
0.935	73%	-1.0%	56%
0.94	72.70%	-0.9%	55%
0.945	72.70%	-0.9%	54%
0.95	72.60%	-0.9%	53%
0.955	72.3%	-0.9%	52%
0.96	69.0%	-0.8%	50%
0.965	70.7%	-0.8%	48%
0.97	69.0%	-0.8%	46%
0.975	66.1%	-0.8%	44%
0.98	61%	-0.8%	41%

Table 2.3 Regression results for Wallaces Creek Lyne and Hollick baseflow separation

k	R ²	Trend	BFI
0.925	68.2%	8.3%	53%
0.93	69.0%	7.6%	52%
0.935	70.0%	6.9%	51%
0.94	70.8%	6.3%	49%
0.945	71.8%	5.6%	48%
0.95	72.8%	5.1%	47%
0.955	73.8%	4.5%	45%
0.96	74.8%	3.9%	43%
0.965	75.4%	3.4%	41%
0.97	75.6%	2.8%	38%
0.975	75.1%	2.2%	36%
0.98	74.0%	1.6%	32%

2.2.3 Manual baseflow separation using monitoring bore data

Manual baseflow separation was performed for the Murrumbidgee River gauge 410535 and Eucumbene River gauge 222522 utilising analysis of baseflow recession curves and nearby groundwater well hydrographs.

The streamflow during times assumed to be dominated by baseflow discharge was correlated with the groundwater level recorded in nearby groundwater monitoring bores. This correlation was then used to predict the contribution of groundwater during higher flow periods, and thus provide an estimate of the contribution of groundwater to total streamflow. Approximately one year of groundwater monitoring data was utilised for this assessment at bores within the Murrumbidgee River and Eucumbene River catchments. See Annexure A of the water assessment section 7.2.2 for further description of the method.

This analysis showed that during 2018, the baseflow index for the Murrumbidgee River at gauge 410535 was around 43%, and the baseflow index for the Eucumbene River at gauge 222522 was also around 43%.

2.2.4 Baseflow separation using salinity as an environmental tracer

Snowmelt and rainfall have a freshening effect on streamflow, while groundwater discharges (baseflow) tend to be more saline. This relationship was used to separate the groundwater discharge component of streamflow from surface flows, using salinity (EC) as a chemical tracer, as per the method described in Miller et al (2014), summarised by Equation 2.3.

Equation 2.3 Salinity mass balance baseflow separation (Miller et al, 2014)

$$BFI = \left(\frac{EC - EC_P}{EC_{GW} - EC_P}\right)$$

Where BFI = the groundwater baseflow component (percent) of streamflow

EC = the salinity of the mean daily flow

EC_P = the salinity of precipitation (rainfall and snow)

EC_{GW} = the salinity of groundwater

Between May 2007 and December 2011, the Bureau of Meteorology and CSIRO (Crosbie, et al., 2012) analysed rainfall chemistry at 21 sites across Australia. The average rainfall salinity at each of the study sites is presented in Figure 2.5, in which it can be seen that there is a general trend of decreasing salinity with distance from the coast. Rainfall measurements at Woomera appear to be an outlier (potentially due to the proximity of salt lakes) and were not included in the analysis presented here. The project site lies approximately 128 km from the coast and was assumed to have an average rainfall salinity of 5.5 mg/L based on the Australia-wide trend.

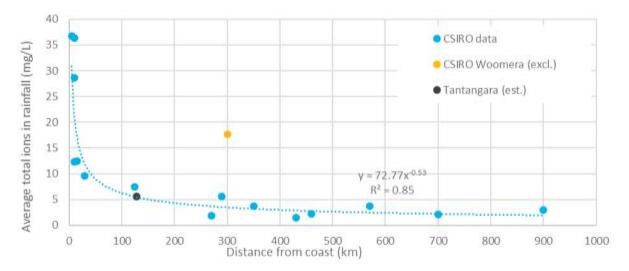


Figure 2.5 Rainfall salinity as a function of distance from the coast

Surface water and groundwater sampling was undertaken monthly during February - November 2018 within the Murrumbidgee River catchment.

The average salinity of the Murrumbidgee River (and tributaries) and local groundwater varied through the year, with samples generally fresher during winter months and more saline during summer months (Figure 2.6). The average salinity in surface water samples was 26 mg/L; for groundwater sampled within 15 m of the surface was 64 mg/L; and for groundwater sampled within 50 m of the surface was 89 mg/L. Due to the increase of groundwater salinity with depth, the analysis was limited to sites at which groundwater was sampled at less than 15 m depth as it is likely that shallow groundwater contributes a greater proportion of baseflow than deep groundwater.

The application of Equation 2.3 produced results shown in Figure 2.7, in which it can be seen that groundwater discharge accounted for between 5 and 125 ML/d of the flow measured at the Murrumbidgee gauge. This analysis shows that over the course of the analysis period, groundwater discharge accounted for **33%** of total streamflow. This estimate may be lower than the actual groundwater discharge component if very shallow groundwater from recent rainfall recharge containing very low salt concentrations was actually a significant portion of the groundwater discharge volume.

2.2.5 Summary of baseflow separation

A summary of the baseflow separation results is provided in Table 2.4. From these data it is apparent that the methods applied each support a yearly average baseflow index of around 40% for each of the three rivers assessed, with low and high estimates giving a range of around 30–55 %. This is consistent with previous work completed by van Tol (2016), who found that the mean baseflow index for Snowy Mountain rivers was 41%, with a range of 29–55%.

Table 2.4 Summary of baseflow separation methods

	Murrumbidgee	Eucumbene	Yarrangobilly	Wallaces
Baseflow Separation method	Gauge 410535	Gauge 222522	Gauge 410574	
Lyne and Hollick (best R ²)	53% ↑		56% ↑	38%
Lyne and Hollick (summer recession)	38%		41%	53%
Manual separation	43%	43%		
Chemical tracer (salinity)	33% [↓]			

Note: 1

↑: Likely to be an overestimate

 ψ : Likely to be an underestimate

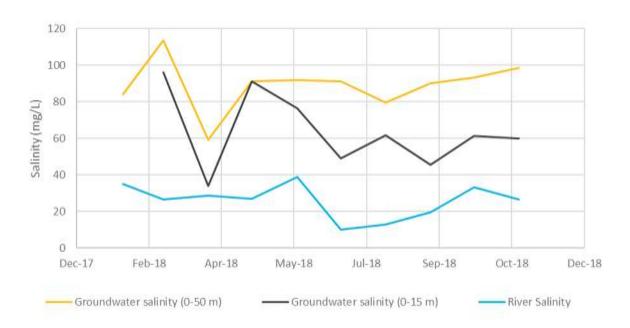


Figure 2.6 Groundwater and surface water salinity within the Murrumbidgee River catchment

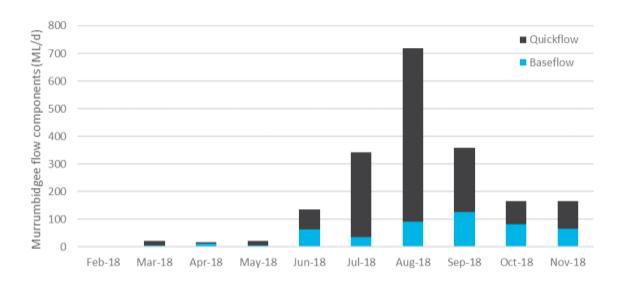


Figure 2.7 Murrumbidgee flow components when spilt using salinity as a tracer for groundwater discharge (utilising groundwater salinity measured in the upper 15 m of the aquifer)

2.3 Baseflow discharge lag analysis

The relationship between rainfall and baseflow discharge was investigated using the Q-Lag method (Brodie, Hostetler, & Slatter, 2007) as a tool to provide understanding about groundwater pathways. The key steps in the application of the Q-Lag analysis were:

- separating each daily streamflow record by day of year (0-365);
- deriving flow statistics for each daily flow population of streamflow data; and
- cross correlating daily flow percentiles with rainfall data.

Low flow (greater than 50% exceedance probability) percentiles for the Murrumbidgee River at gauge 410535 are presented in Figure 2.8 with the 25%ile 14 day average daily rainfall (14R25¹) and 25%ile 14 day average net daily rainfall (rain minus evaporation). This plot shows that net rainfall tends to increase during April–June, then decline September–December. Low flows tend to increase in June–August (two months after the net rainfall increase), then decline September–December (matching the timing of the net rainfall decline).

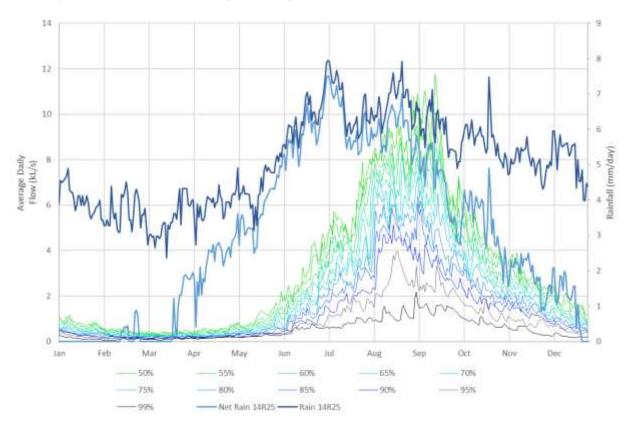


Figure 2.8 Daily discharge for various low flow percentiles of Murrumbidgee River flow at gauge site 410535 compared with 14R25 SILO rainfall and net rainfall (SILO rainfall minus SILO Morton's PET)

This chapter has followed the convention of Brodie, Hostetler & Slatter (2007) in referring to rainfall data in terms of the averaging period in days, and the exceedance probability. 25% ile rainfall was used as approximately 50% of days had no rain, and 25% ile provided a data series for analysis not dominated by either dry days or extreme events.

Rainfall was cross correlated with August–December streamflow for time lags of 0–180 days using Equation 2.4. The results of the cross correlation analysis are presented in Figure 2.9, which shows that the decline in low (Q99² and Q90) and high flows (Q50 and Q10) fitted the decline in net rainfall best when a lag of around 10–30 days was applied. The extreme flow (Q1) data appeared anomalous, likely due to the intermittent nature of extreme flow events and is not relevant to discussion of baseflow.

Equation 2.4 Cross correlation equation

$$r_m = \frac{COV_{1,2}}{S_1. S_2}$$

Where

- r_m is the cross-correlation statistic (closer to 1.0 indicates a better fit);
- $COV_{1,2}$ is the covariance of the two data sequences; and
- s_z and s_2 are the standard deviations of the two data sequences.

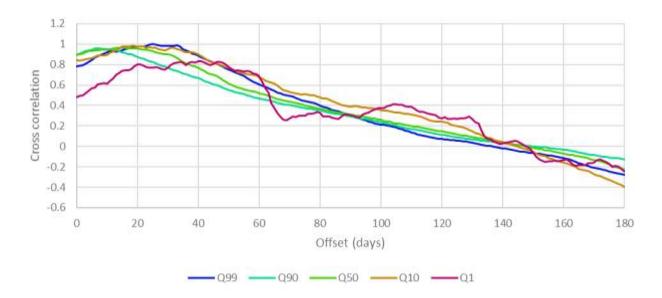


Figure 2.9 Cross correlation statistics for offset net 14R25 rainfall data with various percentiles of Murrumbidgee gauge 410535 streamflow data

The analysis was repeated for the Yarrangobilly gauge 410574, with results again indicating a lag of about two months between increases in net rainfall and increases in streamflow (Figure 2.10), and maximised cross correlation between streamflow and 14R25 net rainfall for lags of <30 days (Figure 2.11).

² Qx in this chapter refers to the flow exceedance probability. Q99 indicates a 99% probability of exceedance, i.,e., a low flow likely to be baseflow. Q50 indicates median flow. Q10 indicates a flow with a 10% exceedance probability, a higher flow rate likely to be dominated by quickflow

The Q-Lag analysis showed:

- there is a lag of approximately two months between net rainfall increases and streamflow increases, likely caused by a need to wet the catchment following summer prior to significant runoff occurring; and
- both quickflow³ and baseflow⁴ decline within weeks of net rainfall declines, indicating that baseflow may be discharging after only a short residence time within the groundwater system.

A short groundwater residence time implies that a large portion of the groundwater discharge travels through relatively short pathways underground, likely remaining at a shallow depth and discharging close to the point of infiltration. A significant component of infiltration occurring along ridge lines thus may be discharging in nearby gullies high in the catchment, with deeper and longer flow paths to the larger rivers and creeks contributing a smaller volume of water.

This conceptualisation is consistent with the two-month discharge lag following the onset of positive net rainfall, as groundwater levels fluctuate with a greater magnitude higher in the catchment and are more stable in the river valleys. During summer the groundwater level high in the catchment may fall to a level such that short pathway discharge doesn't occur. During this time, groundwater discharge may still occur in the large river valleys where groundwater remains close to the surface, likely supported by longer/slower flow paths. Following net rainfall increase, the groundwater level in the upper catchment must rise prior to discharge commencing discharge, leading to the two-month lag.

The groundwater model does not include all upper catchment discharge features as river boundary conditions as there are an innumerable number of small features, many of which may discharge only intermittently. Groundwater approaching the surface in locations not served by a river boundary condition was removed from the groundwater model as distributed baseflow using the evapotranspiration model package (section 1.3).

These Q-Lag results show that the groundwater model estimates of discharge to cells with river boundary conditions may match the baseflow data in the surface water model during drier times, but during wet times when discharge is occurring higher in the catchment to minor stream features distributed baseflow discharge estimation will form an important part of the groundwater model discharge water balance.

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³ The component of streamflow that has travelled through the catchment as interflow or across the surface as overland flow or is released from bank storage during the recession from a flood peak.

⁴ The component of streamflow supplied by groundwater discharge. Baseflow is characterised by an exponential decay curve following the cessation of surface runoff.

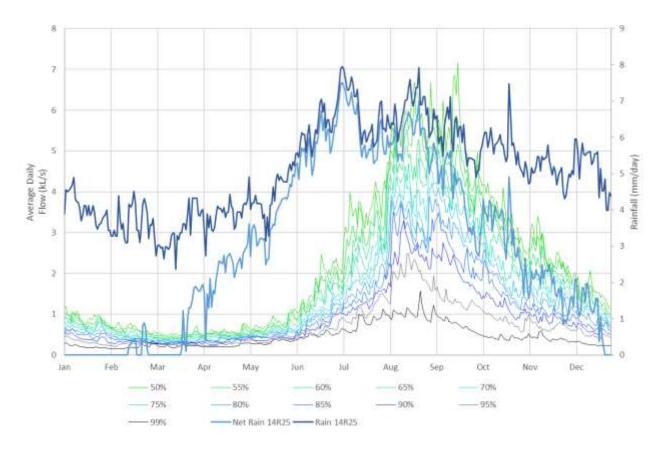


Figure 2.10 Daily discharge for various low flow percentiles of Yarrangobilly River flow at gauge site 410574 compared with 14R25 SILO rainfall and net rainfall (SILO rainfall minus SILO Morton's PET)

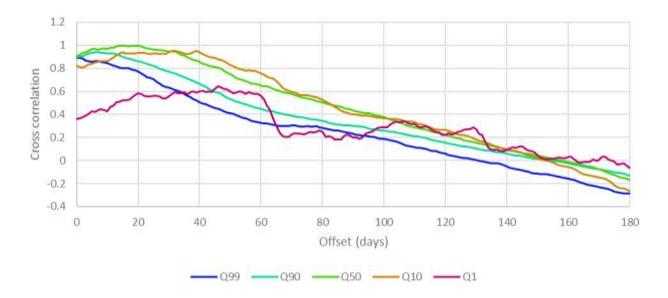


Figure 2.11 Cross correlation statistics for offset net 14R25 rainfall data with various percentiles of Yarrangobilly gauge 410574 streamflow data

2.4 Model design

2.4.1 Model software

The catchment model was developed using the eWater Source software (Source)⁵. Source is a hydrological modelling platform originally developed by the eWater Cooperative Research Centre (CRC) until June 2012, with development post June 2012 undertaken by eWater Limited. The software supports planning of water resource systems at catchment scales by providing a flexible framework in which to integrate spatial, climate and hydrological data with published rainfall runoff models and a plugin interface for customisation (Carr & Podger, 2012).

Within the Source framework, a number of runoff models were tested to determine the numerical approach best suited to the project.

2.4.2 Data utilised in model construction

Data use to create the Source model were:

- rainfall;
- potential evaporation/evapotranspiration (PET);
- digital elevation model (DEM);
- recorded streamflow data; and
- aerial imagery.

The nearest precipitation gauges maintained by the Bureau of Meteorology and Snowy Hydro are described in Annexure A of the water assessment section 4.2. Other than the Bureau of Meteorology gauge 71000 (Adaminaby Tourist Park), the rainfall records at these stations are significantly shorter than the recorded streamflow records. To enable modelling to utilise the entire streamflow record period for calibration, rainfall data were sourced from the SILO (Scientific Information for Land Owners) Data Drill website. SILO is hosted by the Science Division of the Queensland Government Department of Environment and Science (DES). The datasets are interpolated from observed climate data obtained from the nearby Australian Bureau of Meteorology stations. SILO data are available nationally at a 0.05 degree (approximately 4.5 km) grid resolution. Thirteen SILO grid points were utilised in the preparation of the Source model, with each sub-catchment utilising data from the closest SILO data drill grid point location (Figure 2.13).

Potential evapotranspiration data were obtained from the same 13 SILO grid points utilised for rainfall data. As per the Source User Guide 4.5 (eWater, n.d.) Morton's areal potential evapotranspiration was used in the model, as this data set is developed for the purpose of estimating evapotranspiration from vegetated landscapes.

A LiDAR derived DEM was utilised when developing sub-catchments, such that sub-catchment boundaries aligned with geographical features.

The model was calibrated using recorded streamflow data for Murrumbidgee River gauge 410535 and Yarrangobilly River gauge 410574, with data obtained from the Bureau of Meteorology website *Water Data Online* (Australian Government Bureau of Meteorology, n.d.).

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⁵ eWater Source 4.7.0.b.8947

Manual streamflow readings were taken during 2018–2019 using a handheld propeller meter at a number of sites within the Murrumbidgee River catchment. Streamflow velocity measurements were taken at monitoring points in an approximate grid pattern (across the stream, and at various depths) with the total streamflow estimated via integration of the measurements. These manual streamflow readings were not used for calibrating the model, but were visually compared to model hydrographs developed at the monitoring locations to provide confidence that the calibrated model did not contain gross errors at the sub-catchment scale.

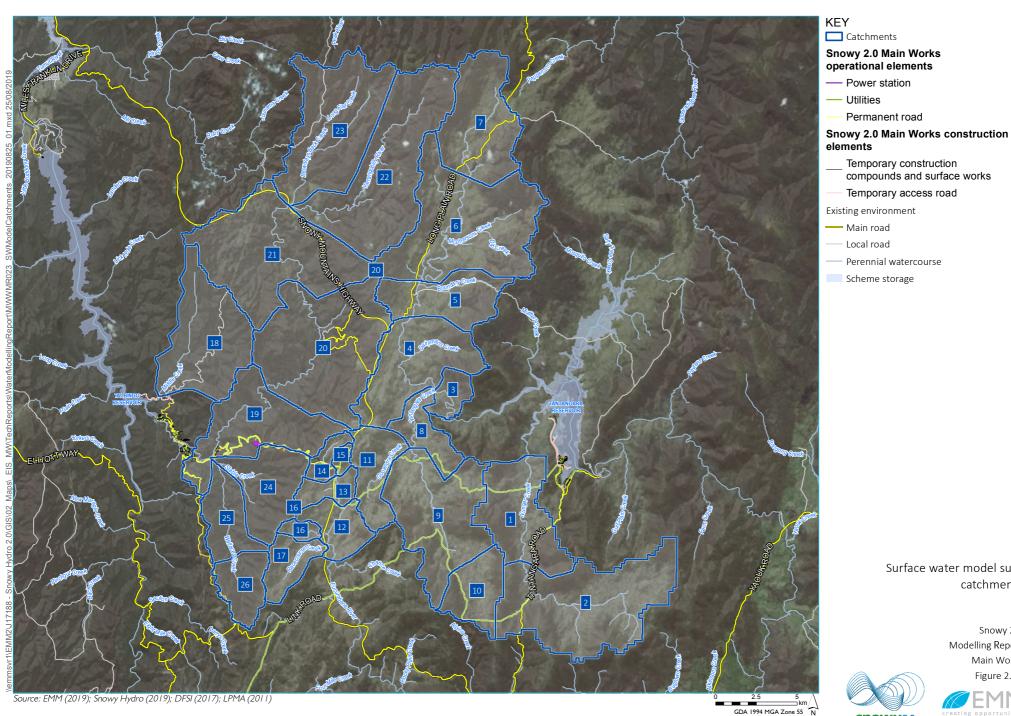
Aerial imagery was used together with GIS analysis to confirm pertinent dimensions such as creek alignments, and vegetation cover.

2.4.3 Sub-catchment delineation

The Source model domain included the Murrumbidgee River upstream of gauge 410535, the Yarrangobilly River upstream of gauge 410574, Middle Creek upstream from Talbingo Reservoir, Nungar Creek upstream from Tantangara Reservoir, and the Eucumbene River upstream (north) of 35.85° S. Each catchment was split into subcatchments based on the following rules:

- the upper limit for sub-catchment area was approximately 50 km²;
- the minimum size was determined by:
 - sub-catchment boundaries aligned with watershed divides; and
 - reporting locations at:
 - the downstream end of groundwater model stream reaches;
 - locations at which manual streamflow readings had been recorded; and
 - established gauge sites.

The model has 26 sub-catchments, illustrated in Figure 2.12, along with stream links and flow measurement points. The SILO grid data were applied to the model sub-catchments as per Figure 2.13.

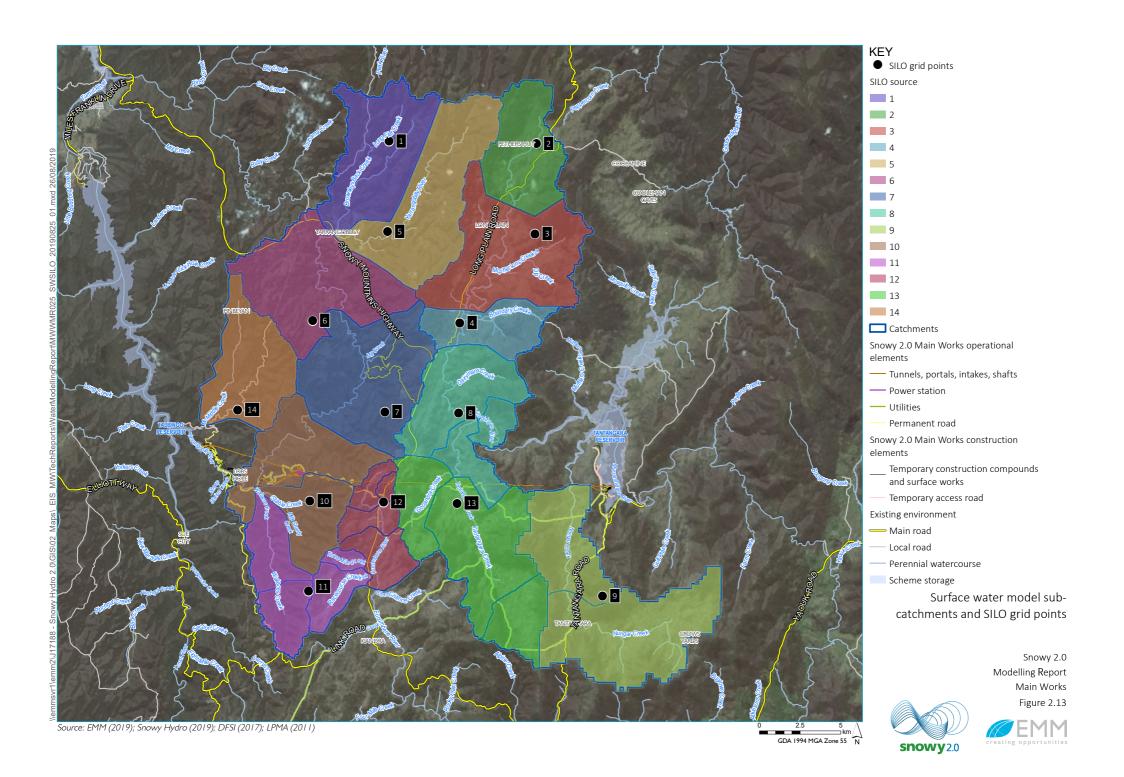


Surface water model subcatchments

> Snowy 2.0 Modelling Report Main Works Figure 2.12







2.4.4 Functional Units

A key concept within the eWater Source modelling framework is the application of runoff response (or 'functional') units at a sub-catchment scale, each with different runoff models or parameter sets. Within the model domain, the following potential functional unit sources were identified:

- land use (ie forested areas, grassland, exposed rock faces, etc);
- catchment slope;
- soil type; and
- underlying geology.

While Crosbie and Jolly et al (2010) report that groundwater recharge is controlled by soil type and vegetation rather than the underlying geology, it was considered possible that geology could affect runoff model terms relating to baseflow discharge. During calibration this theory was tested by applying a functional unit split based on the presence or absence of Gooandra Volcanics, as this geological unit appears to be distinctly different to other geological units in the project area in terms of its higher hydraulic conductivity. Calibration performance was not improved by creating functional units based on the presence or absence of Gooandra Volcanics, and so the presented model does not use functional units created on this basis.

The Digital Atlas of Australian Soils (Northcote, et al., 1960-1968) describes soils across the project area predominantly as kurosols. As the soil type was largely consistent across the model domain, soil type was not used to define separate functional units.

Catchment slope varies markedly between the plateau region containing the Murrumbidgee and Eucumbene rivers, and the ravine region containing the Yarrangobilly River. The land use also varies significantly between these two regions, with the plateau area dominated by low grasses, and the ravine area heavily forested. On this basis, the model domain was split into ravine and plateau functional units.

While the plateau is dominated by grasses, there are stands of trees throughout. The plateau region was not split into grassland and treed functional units as the distribution of forested areas appears to be relatively even between plateau sub-catchments, and so a runoff model parameter set representing a grassland/tree mix is broadly applicable across each plateau sub-catchment.

The Wallaces Creek catchment within the ravine region has a steeper slope than the rest of the Yarrangobilly River catchment, and could have a different runoff relationship. This catchment was gauged from 1969-1999. A separate set of runoff model parameters was used within the Wallaces Creek catchment to attempt calibration to the Wallaces Creek gauged flows, but this produced runoff model parameters outside believable ranges. This was ultimately abandoned and the Wallaces Creek gauge data was used instead for comparison/validation. The presented model utilises the same parameter sets within the Wallaces Creek catchment as within the rest of the Yarrangobilly River catchment.

2.4.5 Runoff model

Source facilitates the use of a number of rainfall runoff models commonly used by hydrologists to describe catchment processes. SIMHYD, AWBM and GR4J were trialled (Attachment B), and it was found that each contained weaknesses that made their use less than ideal for meeting the modelling objectives:

- the SIMHYD model did not allow interflow processes to occur on days following rain;
- the AWBM model utilised a fixed baseflow index that did not respond to seasonality or catchment wetness;
 and
- the GR4J model did not provide an explicit groundwater recharge and discharge pathway.

A custom runoff model was created to address these weaknesses. This custom model was conceptualised as a modification of SIMHYD, utilising the explicit groundwater recharge and discharge pathways but altering the recharge and discharge equations to align with the conceptual model of the site hydrology. A number of alterations were tested (see Attachment B.5), with those contributing to improved calibration retained in the final model.

To allow groundwater recharge and interflow to occur on non-rain days, the equations relating to flow out of the soil moisture store were altered. A new variable was introduced describing the minimum soil saturation threshold required for recharge to the groundwater store or interflow. A 'saturation variable' utilised for scaling recharge and interflow rates was then calculated as per the example in Figure 2.14. When the soil moisture store was saturated, groundwater recharge and interflow occurred at the maximum allowed rate, while when the soil moisture was at or below the saturation threshold no recharge or interflow was allowed. At intermediate soil moisture saturation the rates were linearly interpolated.

The groundwater recharge and interflow equations were altered to take the form shown in Equation 2.5. The soil moisture minimum threshold, interflow coefficient, and recharge coefficient were each varied through the model calibration process.

Evapotranspiration equations were not altered and were not subject to the minimum soil moisture threshold.

Equation 2.5 Altered interflow and recharge equations

$$Interflow \left(\frac{mm}{d}\right) = Interflow Coefficient \left(\frac{mm}{d}\right). Saturation Variable (\%)$$

$$Recharge(\frac{mm}{d}) = Recharge\ Coefficient(\frac{mm}{d}).Saturation\ Variable(\%)$$

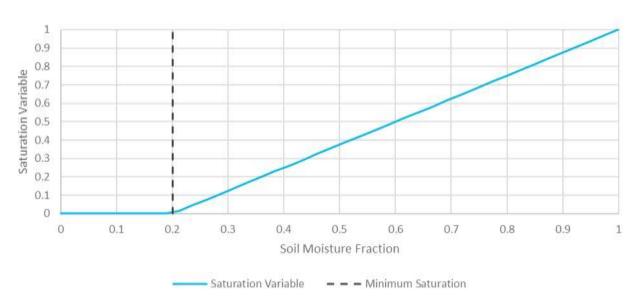


Figure 2.14 Example saturation variable using a minimum saturation for drainage of 20%

During field investigations within the plateau region, it was noted that:

- downstream from groundwater discharge locations there was often a significant area of saturated soil, with vegetation utilising the discharged water prior to it reaching the larger creeks. In many cases, saturated alluvial material; and
- adjacent to some creeks was a significant width of saturated alluvial material supporting dense populations of grasses (see Photograph 2.1).

An analysis of the project stream network spatial dataset was undertaken to determine the approximate portion of the plateau catchments that might contribute to post-discharge/runoff evapotranspiration. While some creeks appeared to support saturated alluvial systems in the order of 100 m wide, tributaries to these creeks and vegetation supported by groundwater seeps covered a smaller area. Spatial analysis showed that that approximately 5% of the plateau lies within 10 m of a mapped creek or creek tributary, which was taken as a reasonable estimate of the area that might contribute to post-discharge/runoff evapotranspiration. The estimated post-discharge evapotranspiration flux was calculated as the daily potential evapotranspiration rate multiplied by the affected area, and was removed from the model at the downstream end of the runoff calculations.



Photograph 2.1 Grasses rooted in saturated soil adjacent to a small creek (not visible)

The runoff model was modified to allow the baseflow loss predicted by the groundwater model to be simulated by adding a 'leakage' term to the groundwater store. This 'leakage' was varied seasonally in accordance with the groundwater model stress periods, and spatially so that the impacts to baseflow predicted by the groundwater model were applied within the appropriate sub-catchments. The leakage rate for each season and sub-catchment was adjusted manually until the baseflow loss matched the scaled impact predicted by the groundwater model. Through the application of a leakage rate, the Modified SIMHYD groundwater store emptied at a faster rate than it would have if no leakage were applied.

A schematic of the modified SIMHYD model is presented in Figure 2.15. The equation set for the modified SIMHYD model is supplied in Attachment C.

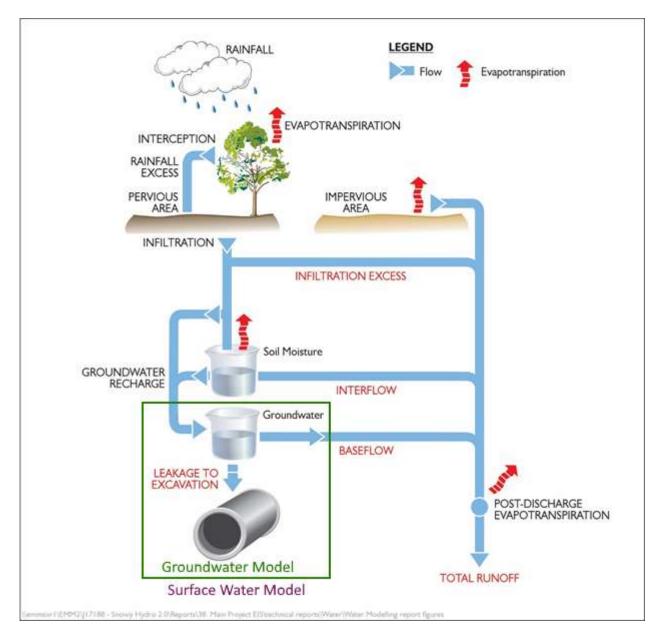


Figure 2.15 Modified SIMHYD runoff model

2.5 Calibration and sensitivity analysis

2.5.1 Calibration method

Optimal model parameters were obtained through the use of the Rosenbrock optimiser contained in Source.

The final parameter set was selected after considering the following metrics:

- Nash-Sutcliffe Efficiency (NSE) for daily flow, log daily flow, daily flow and log flow duration, and monthly flow at:
 - Murrumbidgee River gauge 410535; and
 - Yarrangobilly River gauge 410574.

The Nash-Sutcliffe Efficiency (NSE) is a normalised statistic that determines the relative magnitude of the residual variance ('noise') compared to the measured data variance ('signal' or 'information'). Values of NSE between 0.0 and 1.0 generally indicate acceptable levels of performance, with a value of NSE greater than 0.5 indicating nominally satisfactory performance. Values of NSE less than zero indicate that the mean observed value is a better predictor than the simulated value (ie NSE<0 indicates unacceptable model performance).

- flow exceedance goodness of fit, particularly the low flow (baseflow) portion of the curve, at:
 - Murrumbidgee River gauge 410535; and
 - Yarrangobilly River gauge 410574.
- average split between baseflow and surface flow discharges from:
 - model sub-catchment 6, located in the plateau region, within the Murrumbidgee catchment; and
 - model sub-catchment 30, located in the ravine region, within the Yarrangobilly catchment.

Model calibration considered the full data record at gauges 410535 and 410574 (approximately 40 years of daily streamflow data). Validation was achieved through comparison of model results to streamflow data collected at other points in the model catchments.

2.5.2 Calibration statistics

Moriasi et al (2007) recommended that watershed model calibration should be assessed against monthly NSE, volume bias, and the ratio of the ratio of the monthly root mean square error to the monthly standard deviation of the measured data (RSR), and suggested calibration statistic target ranges should be provided (Table 2.5). In addition to these statistics the calibration considered the log daily NSE, a combined daily and log flow duration NSE statistic, and the baseflow index.

When assessed against the target ranges provided by Moriasi et al (2007), the modified SIMHYD model calibration achieved a good to very good calibration at the Murrumbidgee and Yarrangobilly gauges (Table 2.6).

Table 2.5 General performance ratings for recommended statistics

performance Rating	RSR	NSE	Volume Bias
Very Good	0% ≤ RSR ≤ 50%	75% < NSE ≤ 100%	Bias < ±10%
Good	50% < RSR ≤ 60%	65% < NSE ≤ 75%	±10% ≤ Bias < ±15%
Satisfactory	60% < RSR ≤ 70%	50% < NSE ≤ 65%	±15% ≤ Bias < ±25%
Unsatisfactory	RSR > 70%	NSE ≤ 50%	Bias ≥ ±25%

Source: Moriasi et al (2007)

Table 2.6 Calibration performance

	Murrumbidgee		Yarrangobilly	
	Calibration Statistic	Interpretation ¹	Calibration Statistic	Interpretation ¹
RSR Monthly	53%	Good	41%	Very Good
NSE Monthly	72%	Good	83%	Very Good
NSE Log Daily	75%	Good	79%	Very Good
NSE Daily and log flow duration	80%	Very Good	69%	Good
Volume Bias	9%	Very Good	1%	Very Good
Baseflow Index	39%	Very Good ²	39%	Very Good ²

Note:

- 1. As per Moriasi et al (2007)
- 2. Baseflow index within pre-determined range (see section 2.2.5)

2.5.3 Selected runoff model parameters

The calibrated model parameters are listed in Table 2.7. The application of these parameters may be seen in Attachment C.

While the groundwater model framework was set up to utilise rainfall and evapotranspiration multipliers, it was not necessary to scale either rainfall or evapotranspiration datasets to obtain a good model calibration, and these scaling parameters remained at 1.0.

The calibrated soil moisture store capacity within the plateau area is notably lower than within the ravine area. Due to the vegetation types present, it is likely that the average root depth on the plateau is shallower than in the ravine, and so a shallower soil moisture store is conceptually appropriate on the plateau. Soil moisture infiltrating below the root depth will no longer be subject to evapotranspiration, and is appropriately modelled via the groundwater store.

The calibrated infiltration coefficient for the plateau is higher than for the ravine. As the ravine area contains steeper slopes, it is conceptually appropriate that, given similar soils, infiltration would be lower in the ravine area. It is also likely that exposed rock faces exist within the ravine, which would lead to lower infiltration.

The interflow and recharge coefficients are also higher in the plateau area than in the ravine. The processes represented by the interflow and recharge equations are related to both the saturated and unsaturated hydraulic permeability of the upper soils, about which little is known. Consequently, it is difficult to make firm statements regarding the appropriate ranges of these parameter values.

Table 2.7 Calibrated model parameters

Parameter Name	Parameter Value within plateau area	Parameter Value within ravine area	
Rainfall multiplier	1	1	
Evapotranspiration multiplier	1	1	
Soil Moisture Store Capacity	100 mm	225 mm	
Impervious Store Capacity	NA – no impervious fraction modelled	NA – no impervious fraction modelled	
Interception Store Capacity	5 mm	2.7 mm	
Impervious Fraction	0%	0%	
Alluvial Fraction	5%	1%	
Pervious Fraction	95%	99%	
Infiltration Coefficient	155 mm/day	144 mm/day	
SQ	1.39	1.76	
Interflow Coefficient	5 mm/day	1.5 mm/day	
Recharge Coefficient	2.89 mm/day	0.61 mm/day	
Minimum Saturation required for soil drainage	20%	18.4%	
Groundwater Store Recession	0.02 (k=0.98)	0.02 (k=0.98)	

2.5.4 Calibrated model catchment water balance

The catchment water balance for the model sub-catchment 4 (located near the centre of the Murrumbidgee catchment, just upstream of the Murrumbidgee River and Tantangara Creek confluence) and model sub-catchment 19 (located near the Yarrangobilly River streamflow gauge, upstream of Lobs Hole) are presented in Figure 2.16 and Figure 2.17. While the individual components of the catchment water balance cannot be compared to measured data, the comparison of the water balance for the plateau and ravine areas reveals contrasts consistent with expectations:

- within the ravine area the model produced greater infiltration excess than in the plateau, consistent with the increased relief in the ravine;
- total evapotranspiration was higher in the ravine area, consistent with the extent of deep rooted vegetation; and
- post-discharge evapotranspiration was higher in the plateau area, consistent with the presence of saturated soil supporting vegetation adjacent to creeks and downstream of groundwater discharge zones observed within the plateau.

The data presented in this section represents average data from single sub-catchments. Individual sub-catchments have slightly different water balances due to spatial variation in the SILO precipitation and potential evapotranspiration data. Water balances will also vary year to year with wetter or dryer conditions affecting the proportion of precipitation evaporating or becoming streamflow.

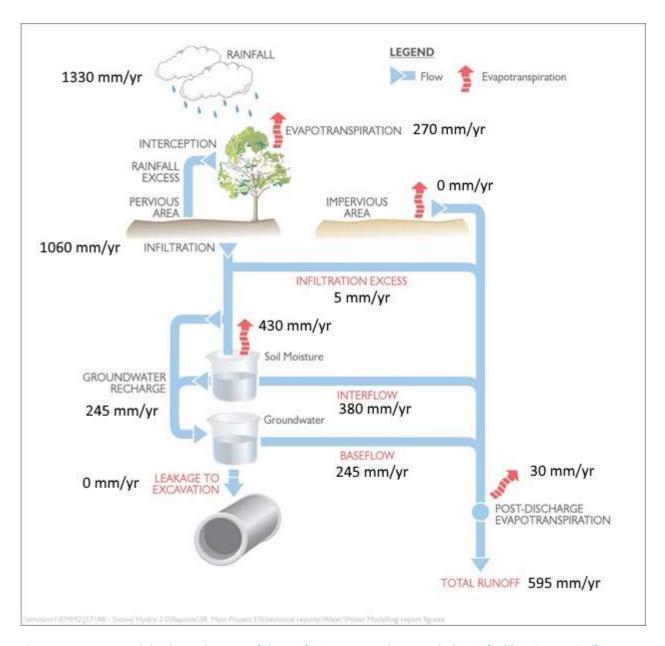


Figure 2.16 Model sub-catchment 4 (plateau) average yearly water balance (calibration period)

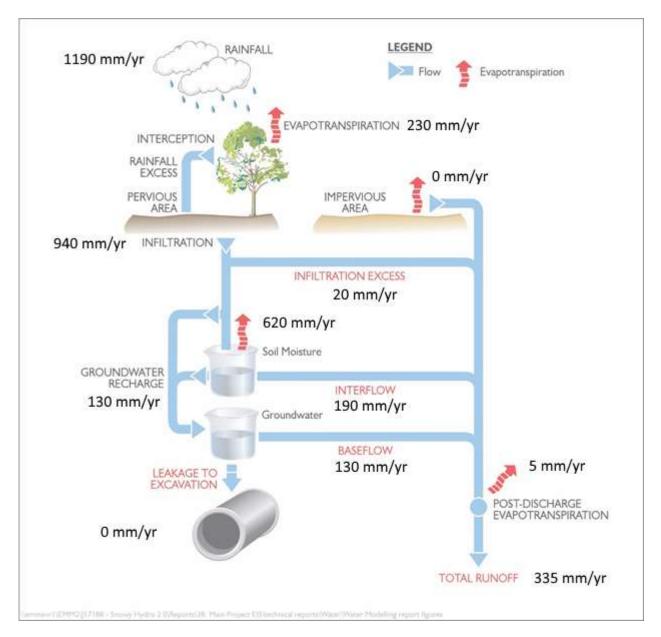


Figure 2.17 Model sub-catchment 19 (ravine) average yearly water balance (calibration period)

2.5.5 Data provided to the groundwater model

Following calibration of the catchment model, groundwater recharge time series data were exported for each sub-catchment for use in the groundwater model. Prior to inclusion in the groundwater model these datasets were spatially aggregated into ravine and plateau regions and temporally aggregated into seasons (Figure 2.18). This data is discussed further in section 3.2.6ii.

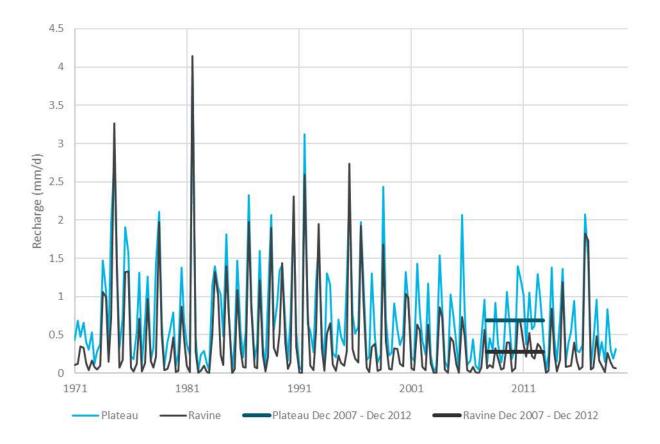


Figure 2.18 Historical quarterly average groundwater recharge predicted by the Source model

2.5.6 Calibration sensitivity

The sensitivity of the model calibration to model parameter selection was evaluated by altering each model parameter through its appropriate range and recording the resulting impact on the combined daily and log flow duration NSE statistic and the baseflow index. Each parameter was varied individually, and the parameters not being tested in each sensitivity run were not altered from the calibrated values. Though not tested, it is likely that in many cases the impacts of altering of one parameter could be offset by recalibration of other parameters.

The results of the parameter sensitivity analysis are presented in Figure 2.19 through to Figure 2.29 with shaded bands indicating the good to very good target range (Moriasi, et al., 2007) for each statistic. These plots show that the calibrated parameter values in each case achieve the best combined fit to the target baseflow index and maximised NSE, indicating that the calibration process reached a local calibration maxima within both the plateau and ravine.

When rainfall and evapotranspiration factors were varied, significant volume bias changes were observed (Figure 2.19 and Figure 2.20). Variation of other parameters had minimal impact on the volume bias, and so volume bias has been reported for rainfall and evapotranspiration sensitivity only.

When reading Figure 2.19 through to Figure 2.28, the calibrated range for each parameter has been taken as the range for which each reported statistic lies within the shaded target ranges. For example, when assessing sensitivity of the calibration at the Yarrangobilly gauge to the recharge coefficient, good NSE statistics were obtained through the parameter range 0.2–5.0 mm/day. Appropriate baseflow statistics were obtained through the parameter range 0.3–1.2 mm/day. The parameter range with good NSE and baseflow statistics is thus 0.3–1.2 mm/day; the appropriate range for this parameter is controlled by the baseflow response.

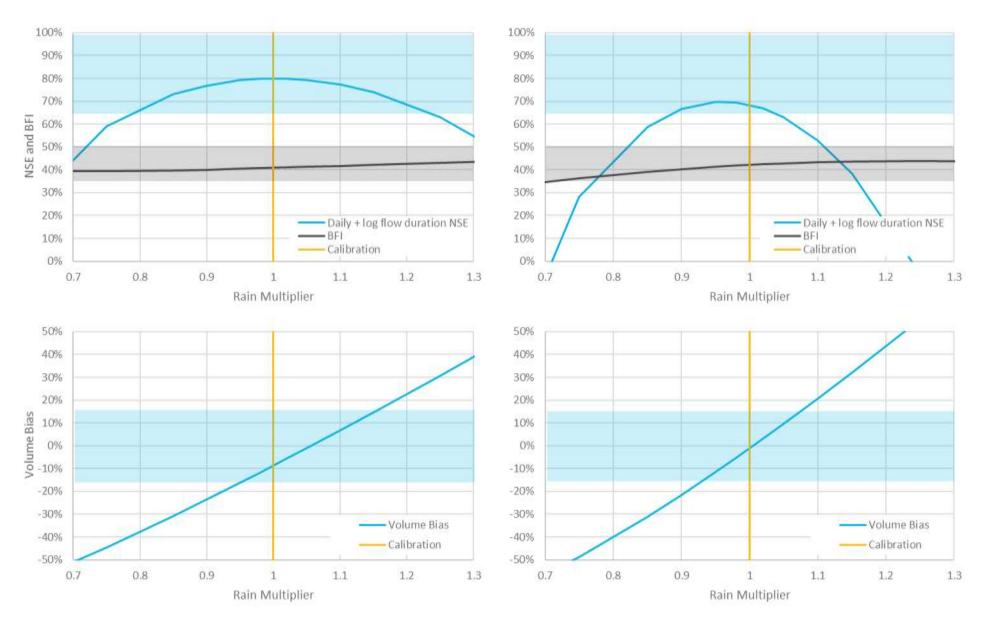


Figure 2.19 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Rainfall Multiplier

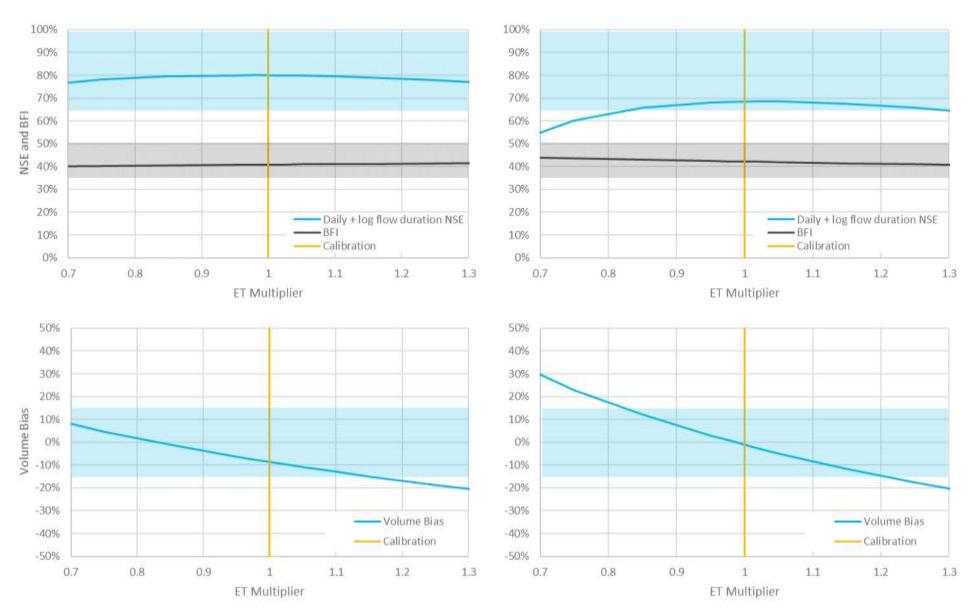


Figure 2.20 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the ET Multiplier

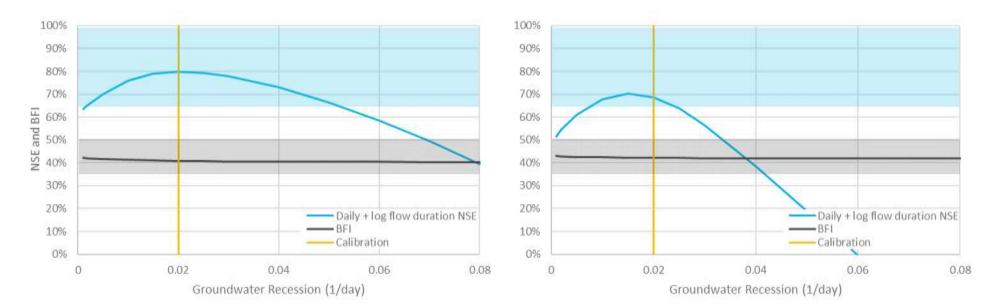


Figure 2.21 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Groundwater Recession parameter

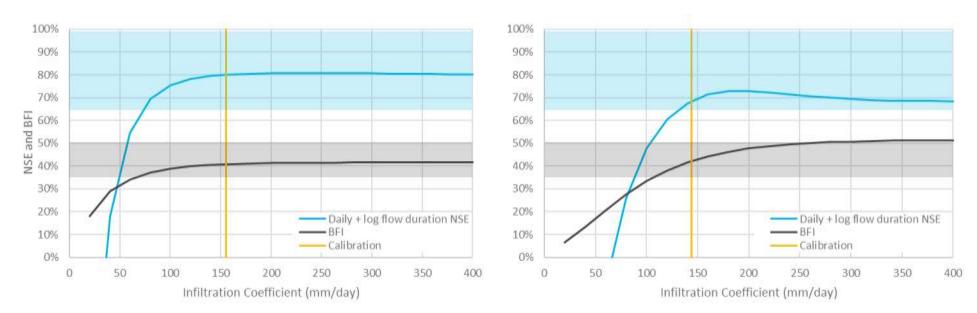


Figure 2.22 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Infiltration coefficient

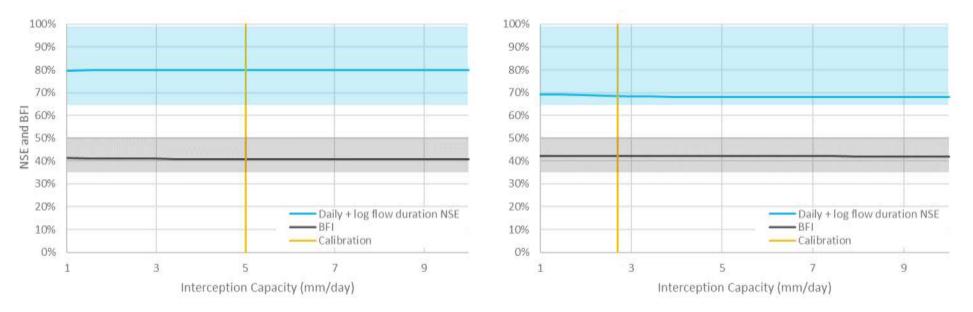


Figure 2.23 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Interception Capacity parameter

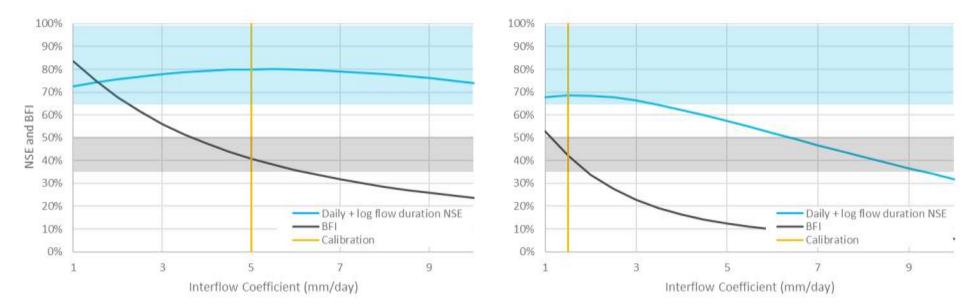


Figure 2.24 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Interflow coefficient

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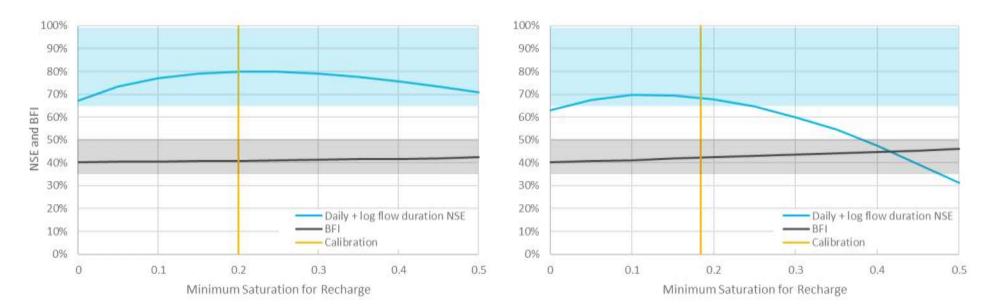


Figure 2.25 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Minimum Saturation parameter

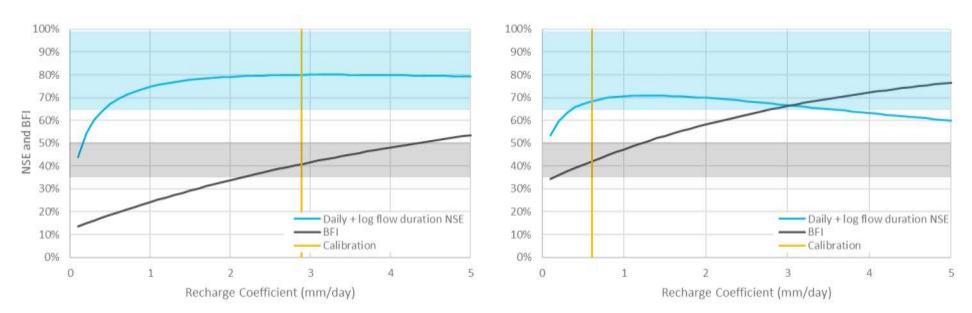


Figure 2.26 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Recharge coefficient

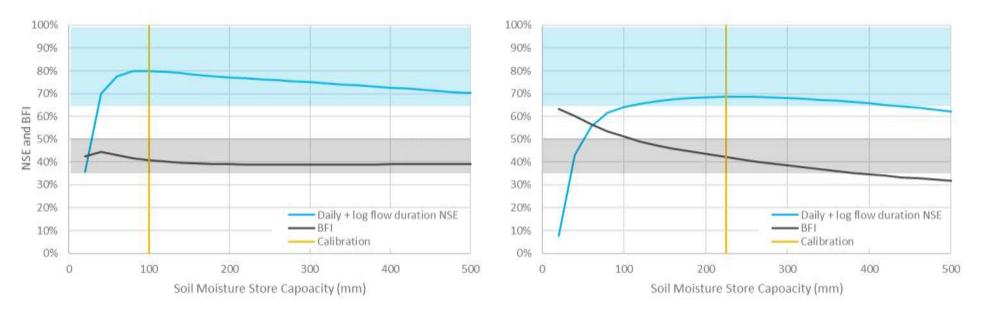


Figure 2.27 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Soil Moisture Store Capacity

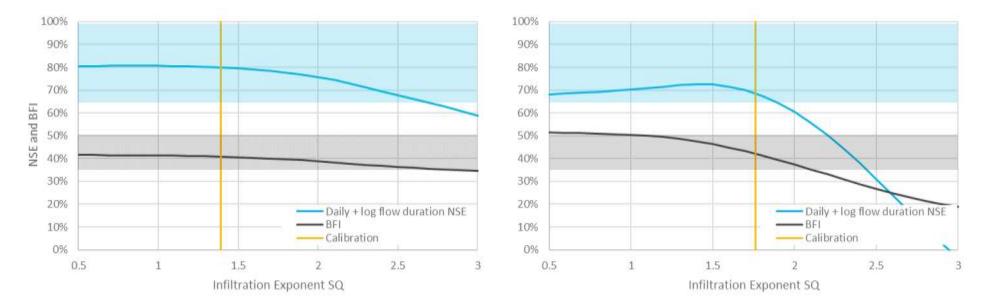


Figure 2.28 Calibration sensitivity at the Murrumbidgee gauge (left) and Yarrangobilly gauge (right) to the Infiltration exponent

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The relative sensitivity of the calibration to the runoff model parameters (Figure 2.29) was estimated by assessing the proportion of the reasonable parameter range (Figure 2.19 through to Figure 2.28) which enabled a good calibration for each parameter. This assessment showed that the calibration to the Yarrangobilly River gauge (Ravine area parameters) was much more sensitive than the calibration to the Murrumbidgee River gauge. The model was most sensitive to the rainfall multiplier and the interflow coefficient, and least sensitive to the interception capacity.

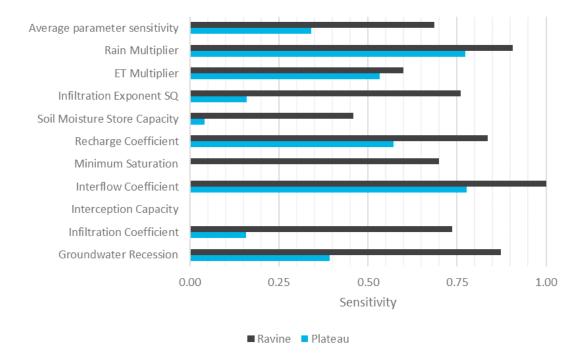


Figure 2.29 Calibration sensitivity to runoff model parameters

2.6 Validation

2.6.1 Method

The predicted runoff was qualitatively compared with manual spot flow measurements recorded at 11 sites utilising a hand held propeller flow meter at an approximately monthly frequency between October 2018 and July 2019. The manual flow readings were obtained by sampling the stream velocity across the width of the stream at regular intervals at multiple depths, with the average flow obtained via a really weighted integration of the individual velocity readings.

Data from a depth logger installed at a culvert across the Eucumbene River became available after the catchment model calibration process was complete. A rating curve for the culvert was developed using the 1D streamflow hydraulic modelling software HEC-RAS with measured culvert dimensions and creek cross sections extracted from LiDAR. This rating curve was used to convert the logged depth data into an estimated streamflow hydrograph for comparison with the catchment model results.

Streamflow data recorded at the streamflow gauge on Wallaces Creek within the Yarrangobilly River catchment was compared to the model predictions of flow past this point.

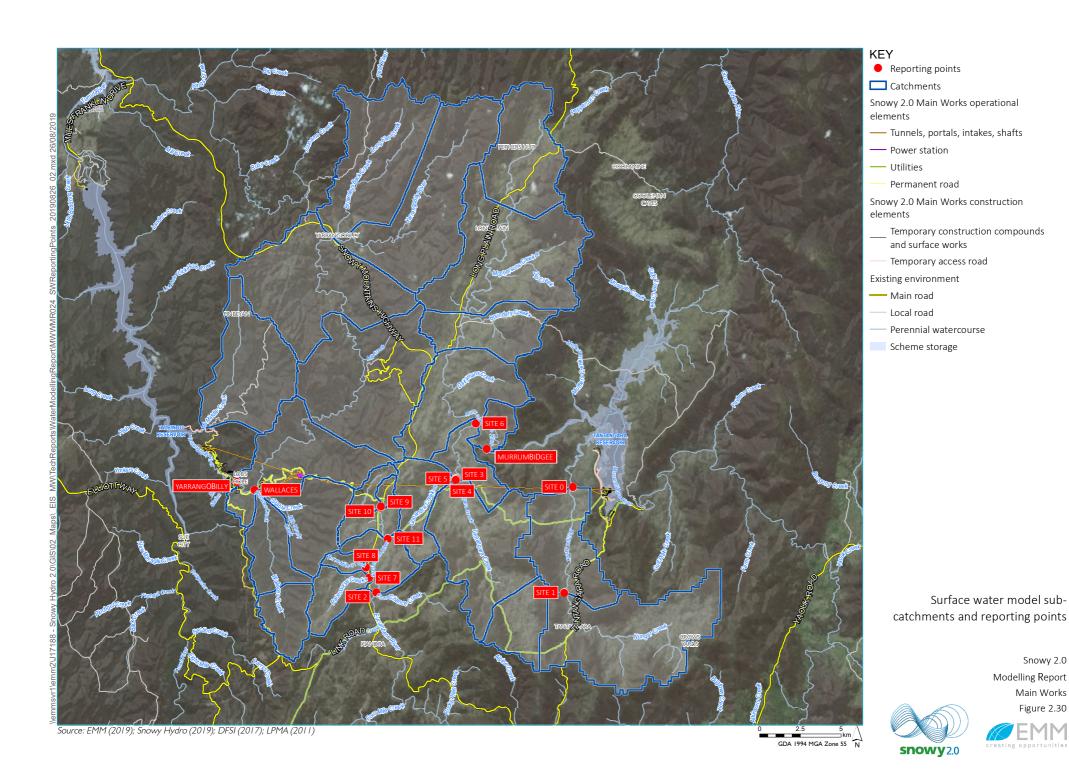
The Q-Lag analysis method (Brodie, Hostetler, & Slatter, 2007) was used to assess the timing and magnitude of median and baseflow components of streamflow, with comparisons made between the modelled and gauged datasets.

2.6.2 Manual flow measurements

Manual flow measurements were recorded at the 11 locations listed in Table 2.8 (coordinates in GDA94 MGA Zone 55) and illustrated in (Figure 2.30). These locations were included within the catchment model at the locations shown in Figure 2.30. Flow measurements were made using a ThermoFisher GLFWP211 protected water turbo prop positive displacement sensor, with a measurement range of 0.1–6.1 m/s. Velocity measurements were taken using a grid sampling approach across the stream channel, at regular depths and distances from bank, with the resulting velocity data integrated to provide a total flow estimate.

Table 2.8 Manual streamflow sites

Site name	Site description	Easting	Northing
Site 0	Nungar Creek	646640	6031810
	Near Tantangara Reservoir		
Site 1	Nungar Creek	641210	6042210
	At groundwater model southern boundary		
Site 2	Eucumbene River	635330	6037070
	At groundwater model southern boundary		
Site 3	Gooandra Creek	635380	6037120
	Upstream of confluence with Tantangara Creek		
Site 4	Tantangara Creek	634630	6032710
	Upstream of confluence with Gooandra Creek		
Site 5	Tantangara Creek	635080	6031830
	Downstream of confluence with Gooandra Creek		
Site 6	Tantangara Creek	639960	6038810
	Upstream of confluence with Murrumbidgee River		
Site 7	Racecourse Creek	640010	6038820
Site 8	Three Mile Creek	639960	6038750
Site 9	Un-named creek	647160	6038300
Site 10	Eucumbene River	634450	6033370



Due to the small sample of manual flow measurements, NSE statistics were not calculated at the manual flow reading sites. Qualitative comparisons were made between the modelled and measured data.

The model predictions at Site 0 and Site 1 on Nungar Creek were a good fit to the measured data (Figure 2.31 and Figure 2.32). The model predicted flows of around 50 ML/day at Site 0 in October–November 2018, similar to the measured data, and flows close to zero in early 2019.

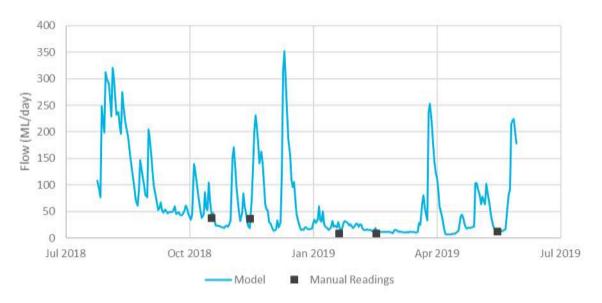


Figure 2.31 Comparison of modelled and measured flow at Site 0 (Nungar Creek)

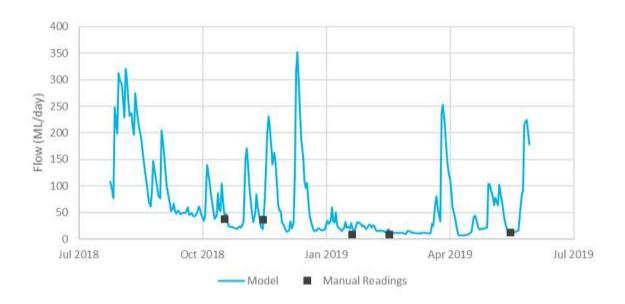


Figure 2.32 Comparison of modelled and measured flow at Site 1 (Nungar Creek)

The model gave a good estimate of flow at Site 2 (Figure 2.33) on the Eucumbene River for low flow conditions in January–February 2019 and a subsequent flow event in response to rainfall in March. There was some disparity between modelled and measured flow in October–November 2018, likely due to the timing and spatial extent of rainfall events occurring during those months not being perfectly represented in the SILO data.

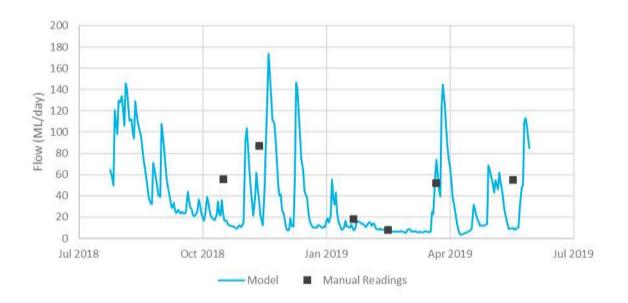


Figure 2.33 Comparison of modelled and measured flow at Site 2 (Eucumbene River)

At the downstream end of Gooandra Creek (Figure 2.34) the model under-predicted creek flow rates at Site 3. In particular, the January–February 2019 manual flow readings are well predicted at other sites but are under-predicted at this site.

This result may indicate that the Gooandra Creek catchment has characteristics that differ from the rest of the plateau catchments. Higher modelled infiltration rates could foreseeably result in higher modelled flows in January due to a larger volume of water in the groundwater store, and would be consistent with the fractured surface geology visible within the Gooandra Creek catchment which occurs along the ridge but not throughout the rest of the plateau.

One alternate plausible explanation is that due to the small dimensions of Gooandra Creek and its upstream catchment, measurement errors of a small magnitude resulted in a large percentage error. The difference between the January 2019 flow measurement and the model prediction is approximately 0.73 L/s, which is of a magnitude that it could be explained as being due to mis-measurement of creek width or depth dimensions.

Another possible explanation is that the Gooandra catchment received higher rainfall than the surrounding catchments in January 2019, and that the spatial rainfall heterogeneity was not represented accurately in the SILO rainfall data.

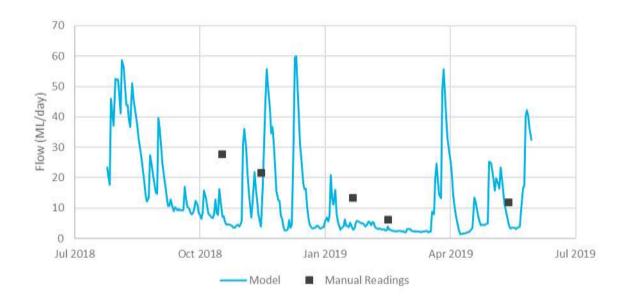


Figure 2.34 Comparison of modelled and measured flow at Site 3 (Gooandra Creek)

The model provided predictions of flow at Site 4, Site 5 and Site 6 in the Tantangara Creek which closely matched the measured flow rates (Figure 2.35, Figure 2.36, and Figure 2.37), other than the October 2018 sampling when it has been presumed there was non-recorded rainfall.

It might be expected that the under-estimation of flow from Gooandra Creek (Site 3; Figure 2.34) would lead to under-prediction of flow downstream of the confluence of Gooandra Creek and Tantangara Creek at Site 5 (Figure 2.36). However, flow prediction at Site 5 appears to have the same level of accuracy as the other sites on Tantangara Creek. This is consistent with the suggestion that the measured flows seen at Site 3 in Figure 2.34 may be higher than the modelled flow due to measurement error.

However, Sites 8–10 (Figure 2.39, Figure 2.40, Figure 2.41) similarly were located on small creeks with catchments of comparable size to Gooandra Creek, yet the modelled and measured flows were closely aligned January–March 2019. This leaves the source of the departure between modelled and measured flows from Gooandra Creek uncertain at the time of writing.

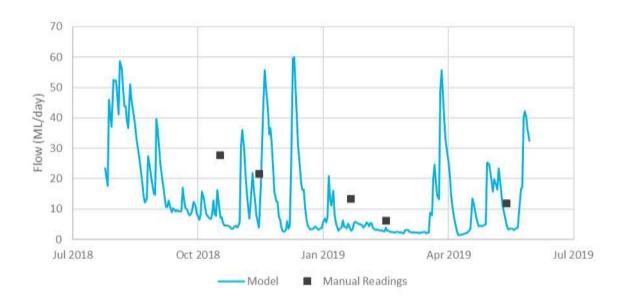


Figure 2.35 Comparison of modelled and measured flow at Site 4 (Tantangara Creek)

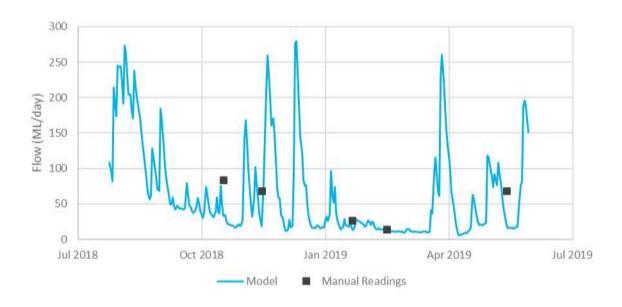


Figure 2.36 Comparison of modelled and measured flow at Site 5 (Tantangara Creek)

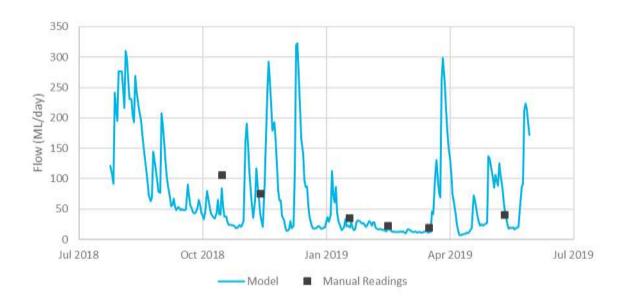


Figure 2.37 Comparison of modelled and measured flow at Site 6 (Tantangara Creek)

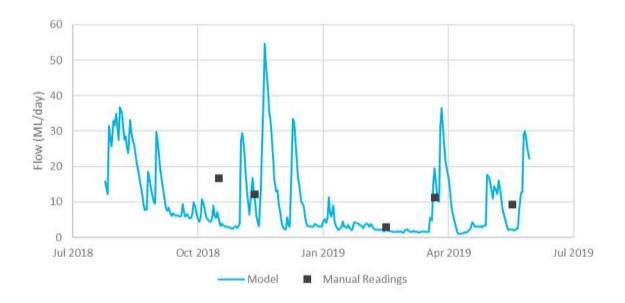


Figure 2.38 Comparison of modelled and measured flow at Site 7 (Racecourse Creek)

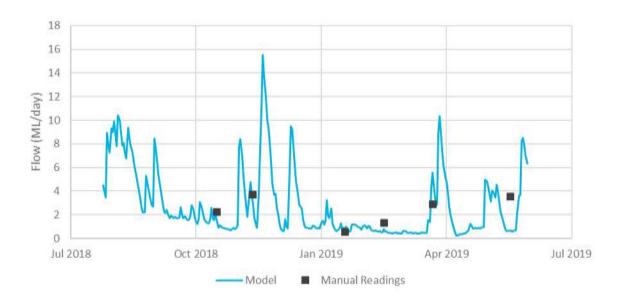


Figure 2.39 Comparison of modelled and measured flow at Site 8 (Three Mile Creek)

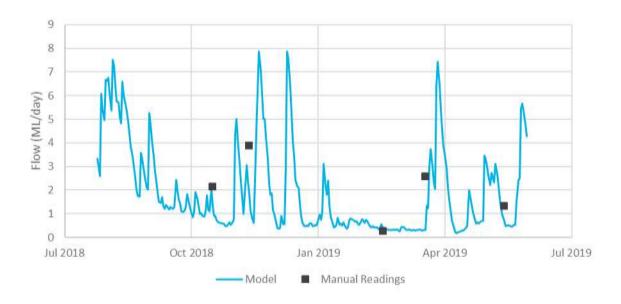


Figure 2.40 Comparison of modelled and measured flow at Site 9 (Un-named creek)

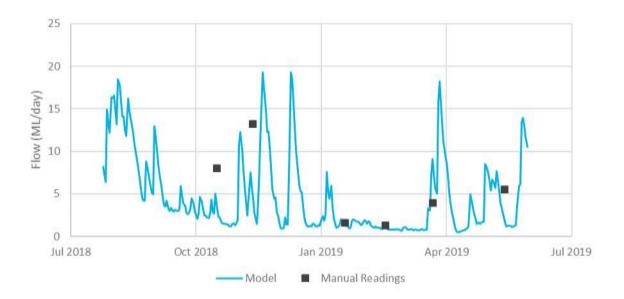


Figure 2.41 Comparison of modelled and measured flow at Site 10 (Eucumbene River)

Overall, the model provided a good fit to manual streamflow measurements, particularly during summer months January–March 2019. It has been hypothesised that unrecorded rainfall in October–November 2018 affected the accuracy of model predictions at a number of the measurement sites in those months.

The predicted flows were systematically lower than the recorded flows at the downstream end of Gooandra Creek. It is currently not certain whether the under-prediction is due to measurement errors (eg in the geometry of the creek at the measurement point) or whether the Gooandra Creek catchment requires an alternate parameter set. Other sites within the plateau have similar elevation and vegetation as Gooandra Creek, and the upper reaches of the Eucumbene River feature the same geology as Gooandra Creek, so there is no clear physical basis for Gooandra Creek to require an alternate parameter set.

2.6.3 Eucumbene River depth logger

A depth logger was installed on the headwall of the box culvert where the Eucumbene River crosses the Gooandra Trail. The logger recorded pressure data at the creek bed at five-minute intervals for the period 16 April 2019 to 17 May 2019, and was corrected for barometric pressure via comparison to a project barometric gauge. Photos of the installation are shown in Figure 2.42.



Figure 2.42 Installation of pressure logger – Eucumbene River at the culvert on the Gooandra Trail (Site 11)

The location of the pressure logger is listed in Table 2.9 (coordinates in GDA94 MGA Zone 55) and was included as a gauge node in the catchment model, see Figure 2.30.

Table 2.9 Pressure logger site

Site name	Site description	Easting	Northing
Site 11	Eucumbene River	635815	6035070
	Culvert at Gooandra Trail		

A 1D hydraulic model of Site 11 was developed using the modelling software HEC-RAS⁶ with measured culvert dimensions (see Figure 2.43), aerial imagery and stream line and cross sections extracted from the project digital elevation model (1 m resolution LiDAR data). This model was used to develop a rating curve for the culvert which was then used to convert the logged pressure data to an estimated streamflow hydrograph for comparison with the catchment model results.

The HEC-RAS model, illustrated in Figure 2.44, extends approximately 35 m upstream and 85 m downstream of the culvert at Gooandra Trail and includes 20 cross sections. The culvert was modelled as a rectangular box culvert having a '90 degree headwall', as the headwall has rock gabions and is not tapered.

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⁶ HEC-RAS version 5.0.5

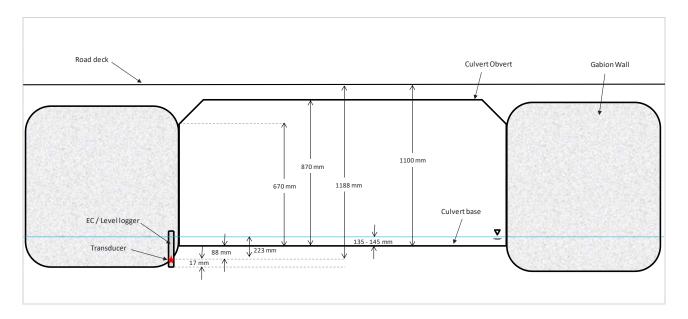


Figure 2.43 Schematic with measured dimensions for the culvert at Site 11

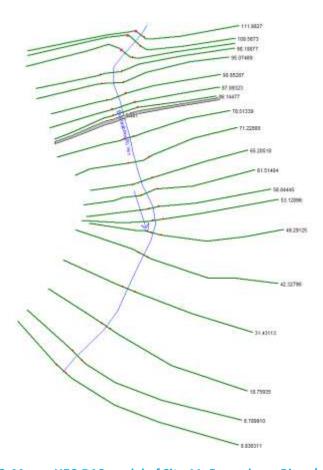


Figure 2.44 HEC-RAS model of Site 11, Eucumbene River (culvert at the Gooandra Trail)

Selection of an appropriate Manning's n value is important as it affects the accuracy of computed water surface elevations in HEC-RAS. The value of Manning's n is highly variable and depends on a number of factors including but not limited to: surface roughness, vegetation and channel alignment (Brunner, 2016). The HEC-RAS Reference Manual includes guidelines for selecting appropriate Manning's n values. For the model of Site 11, the following Manning's n values were applied:

- within the culvert a Manning's n value of 0.015, consistent with unfinished concrete; and
- main channel and floodplain a Manning's *n* value of 0.05, assuming the following descriptors:
 - main channel: Clean, winding, some weeds, stones; and
 - flood plains: Scattered brush, heavy weeds (NOTE: same Manning's n as in channel).

The rating curve was developed by modelling a range of flows through the culvert from 0.01–2.5 m³/s, extracting and plotting the water surface elevation upstream of the culvert (at the logger location) against the modelled flow. Normal depth was assumed at the downstream end of the model.

In the absence of detailed data/information about conditions at the site for calibration, a higher (0.07) and lower (0.03) creek bed Manning's *n* value were applied in the HEC-RAS model to provide an indicative upper and lower bound to the rating curve. The rating curve is shown in Figure 2.45. Higher and lower Manning's n value for within the culvert were applied in the model and found not to significantly affect results - the results of this analysis are not shown.

The project digital elevation model shows a low point in Gooandra Trail on the eastern bank of the Eucumbene River such that when water levels rise above 1382.2 m AHD, water flows both through the culvert and over the top of the track. This can be seen in see the HEC-RAS cross section immediately upstream of the culvert (Figure 2.46) and is responsible for the increase in flow relative to water surface elevation relative in the ratings curve at this elevation.

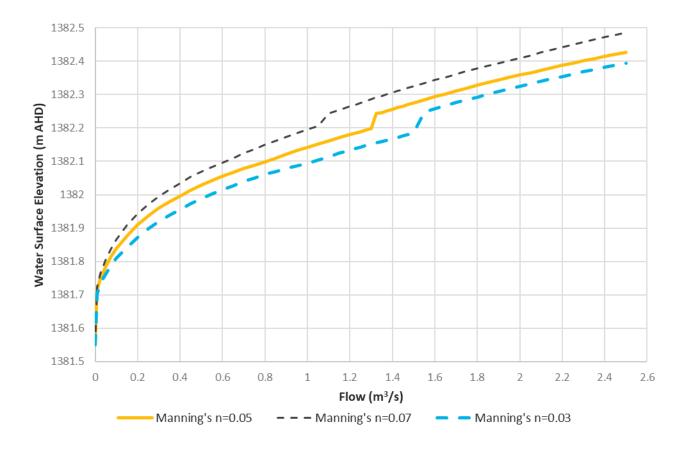


Figure 2.45 Rating curve for Site 11, Eucumbene River (culvert at the Gooandra Trail)

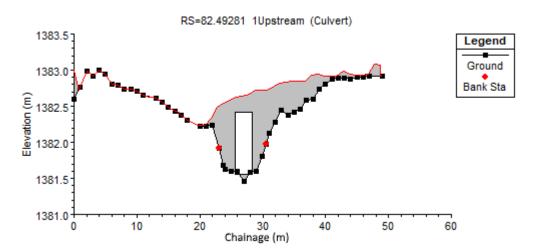


Figure 2.46 HEC-RAS cross section of the Eucumbene River immediately upstream of the Gooandra Trail culvert at Site 11

The estimated streamflow at this location based on the modelled rating curve are presented in Figure 2.47 for the monitoring period (16 April 2019–1 June 2019), together with streamflow estimates from the calibrated catchment model. The flow comparison shows that the model did not predict the peak flow recorded in May or June, but that flow predictions during recession and low flow periods closely matched the recorded data. Daily NSE and RSR statistics showed that the Source model produced a good to very good match to the recorded data (Table 2.10).

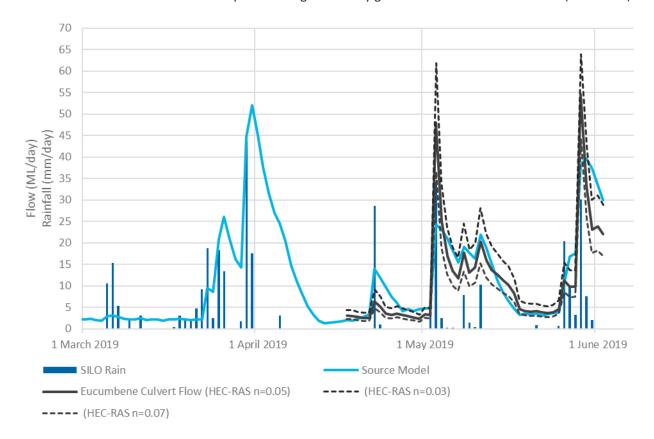


Figure 2.47 Comparison of rainfall data, catchment model flow and flow converted from logger data at Site 11

Table 2.10 Eucumbene River validation statistics

	Calibration Statistic ¹	Interpretation ²
RSR Daily	53%	Good
NSE Daily	72%	Good
NSE Log Daily	84%	Very Good
NSE Daily and log flow duration	84%	Very Good
Volume Bias	10%	Good

Note:

- $1. \ Comparison \ of \ Source \ model \ with \ logger \ data \ converted \ to \ flow \ using \ the \ rating \ developed \ with \ the \ Hec-Ras \ model \ using \ n=0.05$
- 2. As per Moriasi et al (2007)

2.6.4 Wallaces Creek streamflow gauge

From 1969 to 1999 a streamflow gauge was operated on Wallaces Creek in the ravine area.

The Wallaces Creek gauge recorded lower runoff per square kilometre than the Yarrangobilly River gauge through much of the flow duration curve, and in particular recorded low flows approximately an order of magnitude lower than the Yarrangobilly River gauge. This could be due to the Wallaces Creek catchment being steeper than the remainder of the Yarrangobilly River catchment, such that runoff in Wallaces Creek is 'flashier' with less sustained interflow and baseflow.

A number of difficulties were encountered during calibration when including data from the Wallaces Creek gauge:

- automated calibration procedures preferred unrealistic parameter sets within Wallaces Creek and Stable Creek which resulted in routing all runoff via the groundwater store with no surface runoff; and
- when utilising separate parameter sets in Wallaces Creek and Stable Creek based on calibration of runoff from those catchments to the Wallaces Creek gauge, the calibration statistics at the downstream Yarrangobilly River gauge were poorer.

As the gauge was decommissioned 20 years ago, the condition of the gauge and the accuracy of the gauge during the record period are not known, and it is possible that the calibration problems experienced when using data recorded at this gauge were caused by a rating curve that was poor at representing either low or high flows.

Due to the described calibration difficulties and the possibility of unreliable gauge data, the data from this gauge was excluded from the model calibration.

The Wallaces Creek catchment was modelled using the same parameter set as the Yarrangobilly River, resulting in the model over estimating runoff within this catchment through much of the flow duration curve. Statistics describing the comparison of the gauged and modelled flow (Table 2.11) at this gauge were poorer than the calibration statistics for the downstream Yarrangobilly gauge (cf Table 2.6).

Rainfall contours (Annexure A of the water assessment Figure 4.1) indicate that average yearly precipitation upstream of Wallaces Creek gauge varies from 950 mm/yr near the gauge to 1300 mm/yr at the Wallaces Creek headwaters. The yearly average precipitation provided by the SILO grid data for Wallaces Creek is 1145 mm/yr, providing a good representation of conditions likely to exist at the catchment centroid.

Preliminary tests undertaken during the model validation phase of the project indicated that reducing the rainfall multiplier within the Wallaces and Stable Creek catchments to 90% reduces the Wallaces Creek volume bias to near zero, while maintaining or improving other calibration statistics at the Wallaces Creek gauge. This change has a minor negative influence on Yarrangobilly River gauge calibration statistics that could likely be ameliorated through calibration optimisation. Changing the rainfall multiplier for the Wallaces Creek and Stable Creek catchments and reintroduction of the Wallaces Creek gauge data in the calibration process was not undertaken prior to EIS submission because:

- the SILO data appears to a provide a good representation of precipitation within the catchment (ie there is no immediate justification for altering the precipitation input other than calibration improvement); and
- the groundwater modelling predicted that impacts to streamflow within the Wallaces Creek catchment due to baseflow reduction are likely to be minor or insignificant (see section 2.7.6), and as such inaccuracy in streamflow predictions at the Wallaces Creek gauge are likely to represent a low risk.

If monitoring during construction and operation indicates that larger impacts to baseflow may occur within Wallaces Creek, local recalibration of the catchment model to improve the volume bias will be required before utilising the model to describe the changes to streamflow. The preliminary rainfall reduction test is documented to

illustrate that a pathway exists for prediction improvement if monitoring during construction indicates a departure from the model results.

Table 2.11 Wallaces Creek validation statistics

	Calibration Statistic	Interpretation ¹
RSR Monthly	53%	Good
NSE Monthly	72%	Good
NSE Log Daily	72%	Good
NSE Daily and log flow duration	75%	Good
Volume Bias	28%	Unsatisfactory

Note: 1. Moriasi et al (2007)

2.6.5 Q-Lag analysis

The modelled streamflow results were analysed with the Q-Lag method described in 2.3 and compared to the analysis completed with gauge data. Comparison of Q-Lag data relating to the Murrumbidgee gauge (see Figure 2.48) showed that the modelled 90th percentile exceedance streamflow in summer months (January–May) is very similar between modelled and gauged data, indicating that baseflow discharges are well represented in these months.

The modelled 50th percentile exceedance Murrumbidgee streamflow begins to increase during April, while the gauged data begins to increase in May, indicating that the model produces excess runoff in autumn and early winter. In August to October the trend is reversed, with the model producing less runoff for both the 50th and 90th percentile exceedance hydrographs.

One possible explanation for this result is that snowfall in early winter is retained in the catchment, melting several months later and contributing to both quick flow and groundwater recharge. As the model does not model a snow pack, this storage and release process does not occur in the model.

The modelled Yarrangobilly streamflow 50th and 90th exceedance percentiles were generally a good match to the gauged data (Figure 2.49). The reduction in streamflow lag seen in the analysis of Murrumbidgee flow data is not apparent in the Yarrangobilly data. The Yarrangobilly River catchment is lower than the Murrumbidgee catchment, and experiences less snowfall, consistent with the possibility that the departures between measured and modelled data in the Murrumbidgee catchment are driven by snowfall.

The effect of storage of precipitation as snow was investigated, as described in B.5.2, with the conclusion that the predictive power of the trialled model was not improved when simulating a snow pack but that a more complex snow pack model may have produced a different result.

As the impacts of the project on streamflow are primarily limited to summer months (section 2.7.6), the weakness of the catchment model to predict storage of snow through the winter is unlikely to affect impact predictions.

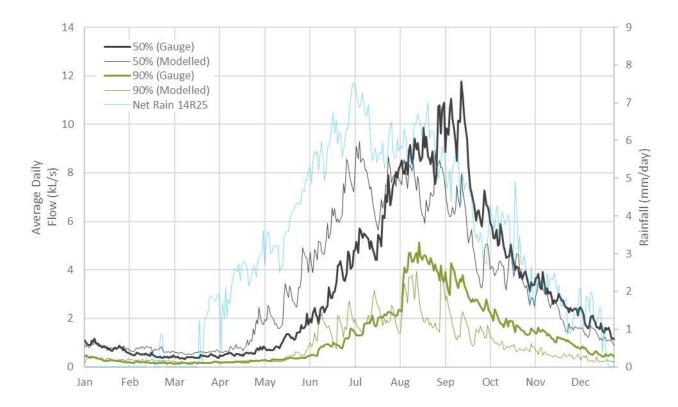


Figure 2.48 Comparison of Murrumbidgee Q-Lag analyses

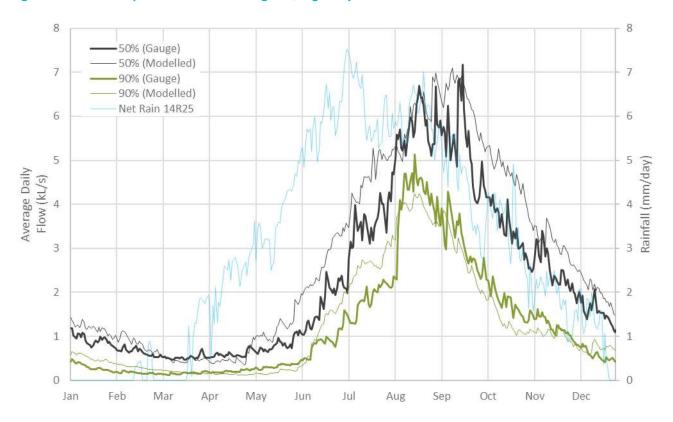


Figure 2.49 Comparison of Yarrangobilly Q-Lag analyses

2.7 Predictive modelling

The calibrated catchment model was used to investigate streamflow regimes for creek and river reaches within Tantangara Creek, Gooandra Creek, Nungar Creek, the Murrumbidgee River, Wallaces Creek, and the Yarrangobilly River at predetermined reporting sites (section 2.7.2).

Streamflow regimes were investigated for the following phases of the project:

- Pre-construction a pre-construction simulation using the historical climate record from 1 Jun 1970—30 May 2019. This simulation describes the flow regime prior to any project impacts and provides a baseline against which other simulation can be compared;
- Construction a set of simulations which considered the impacts of tunnel excavation during the 5.5 year construction period using average, wet and dry climate sequences from within the historical record from 1 Jun 1970–30 May 2019. These simulations were developed using transient baseflow discharge results from the groundwater model (section 2.7.5); and
- Operating a post-construction simulation considering the long-term impact of operating the power waterway. This simulation uses the historical climate record from 1 Jun 1970–30 May 2019 and was developed using steady state baseflow discharge results from the groundwater model (section 2.7.6).

One construction/operation scenario was assessed reflecting the current project construction schedule, and using historical climate data inputs. This scenario is the "revised groundwater modelling scenario" and represents a lined but unmitigated tunnel scenario.

The groundwater model run identifiers, catchment model filenames and catchment model run identifiers used to simulate each phase of the project, are recorded in Table 2.12.

Table 2.12 Groundwater model run identifiers and catchment model filenames for each project phase

Project phase	Groundwater model run	Catchment model filename	Catchment model run	
Pre-construction		pacts are given in square brackets unde c conditions and simulation period)	erneath corresponding model	
Construction (average)	SH4.0_tpred31		Construction – Average	
	[SH4.0_tpred10]		[Preconstruction – Average]	
Construction (wet)	SH4.0_tpred33	Snowy 2.0 M05 2020 01 28 Constru	Construction – Wet	
	[SH4.0_tpred16]	ctionScenario_RTS.rsproj	[Preconstruction – Wet]	
Construction (dry)	SH4.0_tpred34		Construction – Dry	
	[SH4.0_tpred17]		[Preconstruction – Dry]	
Operating	SH4.0 sspred24	Snowy	Operating	
	[SH4.0_sspred10b]	2.0_M05_2020_01_28_PreCons truction_Operating_RTS.rsproj	[Preconstruction]	

2.7.1 Method for applying predicted baseflow impacts to the catchment model

In some catchments the groundwater drawdown was predicted to occur (ie reductions in the water table as a result of the project. In areas where streams were present, this also resulted in reductions to the available groundwater for baseflow also being predicted by the groundwater model. These catchments are illustrated in Figure 2.50, with groundwater drawdown seen to occur primarily in the Gooandra and Eucumbene catchments.

The impacts to baseflow predicted by the groundwater model on a quarterly (seasonal) time step were incorporated in the catchment model as a leakage rate from the groundwater store. The leakage rate was applied as a fixed rate through each season in temporal alignment with the groundwater model results. Leakage rates were only applied within model subcatchments substantially affected by groundwater drawdown as illustrated in Figure 2.50. The magnitude of the leakage rate was adjusted until the baseflow reduction achieved in the catchment model matched that predicted in the groundwater model.

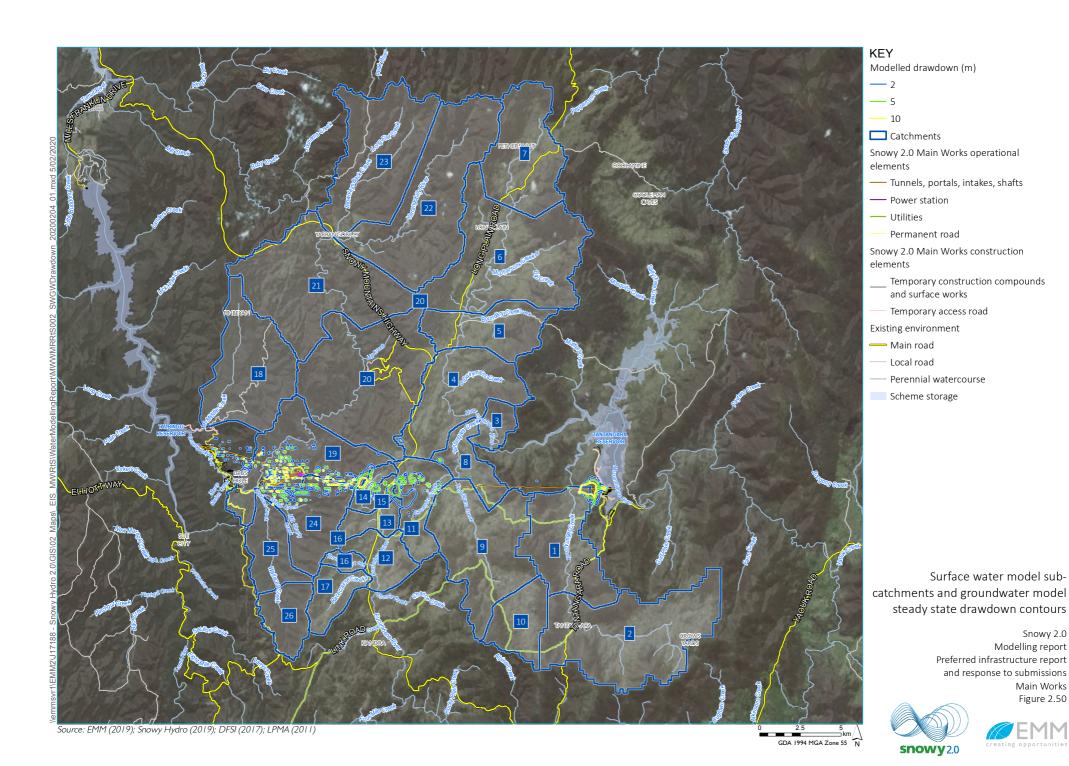
Baseflow reductions predicted by the groundwater model were applied in the catchment model using a 'leakage' term in the Modified SIMHYD rainfall runoff model (see Figure 2.15). This leakage term causes the Modified SIMHYD groundwater store to empty at a faster rate, which results in reduced baseflow.

The leakage rate was varied:

- spatially, so that the impacts to baseflow predicted by the groundwater model are applied within the appropriate sub catchments; and
- seasonally, in accordance with the groundwater model stress periods (construction phase only).

The leakage rate for each season and sub catchment was adjusted iteratively until the baseflow impact in the catchment model matched the baseflow impact predicted by the groundwater model. Where the groundwater model predicted very small reductions in baseflow (<5%); these impacts were not applied in the catchment model due to the levels of uncertainty inherent in the groundwater modelling and because very small reductions in baseflow did not produce measurable reductions in streamflow.

Further discussion of the baseflow loss method is supplied in Attachment B.5.



2.7.2 Reporting sites

Sub-catchments in the catchment model were delineated to allow reporting of streamflow results at locations (section 2.4.3):

- which coincided with the downstream end of groundwater model stream reaches;
- where manual flow measurements were taken; and
- at established streamflow gauge sites.

These locations are referred to as reporting sites and are shown in Figure 2.30, with coordinates given for these sites in Table 2.8.

Streamflow results are only presented for reporting sites with >5% reduction in baseflow predicted by the groundwater model. Groundwater modelling showed that key drawdown impacts were in the area within the Gooandra Volcanics, in the western section of the plateau and to a lesser extent in the Kellys Plain Volcanics in the eastern section of the plateau (Figure 2.50). Therefore, based on the groundwater model results, baseflow reductions were applied to sub-catchments in Gooandra Creek and the Eucumbene River and results are presented for reporting sites downstream of these sub catchments. These sites are listed and briefly described in the following sections.

i Gooandra Creek

Gooandra Creek is a located on the plateau in the upper reaches of the Murrumbidgee River catchment and is a tributary to Tantangara Creek. Within the catchment model, Gooandra Creek is represented by one sub catchment to which predicted baseflow reductions were applied.

Streamflow results are shown for the reporting site at the outlet of this sub catchment and for the reporting sites in the river reaches downstream of the Gooandra Creek sub catchment:

- Site 3 Gooandra Creek upstream of the confluence with Tantangara Creek;
- Site 5 Tantangara Creek downstream of the confluence with Gooandra Creek;
- Site 6 Tantangara Creek upstream of the Murrumbidgee River confluence; and
- Murrumbidgee Gauge on the Murrumbidgee River downstream of the confluence with Tantangara Creek.

ii Eucumbene River

The Eucumbene River is located on the plateau.

Based on results from the groundwater model, particularly drawdown contours produced using the groundwater model results (Figure 2.50), only the headwater catchments of the upper reach of the Eucumbene are predicted to be impacted by the tunnel excavation.

The upper reach of the Eucumbene refers to the river reach above the confluence with Racecourse Creek and is represented in the catchment model by four small sub catchments. Baseflow reductions predicted by the groundwater model were applied to the upstream two of these four sub catchments.

Streamflow results are shown for the reporting sites at the outlet of each of these four sub catchments:

- Site 10 Eucumbene River upstream of Snowy Mountains Highway. This is a headwater catchment;
- Site 9 Unnamed watercourse downstream of Snowy Mountains Highway, upstream of its confluence with the Eucumbene River proper. This is a headwater catchment;
- Site 11 Eucumbene River Culvert Logger where the Eucumbene River crosses the Gooandra Trail; and
- Site 2 Eucumbene River at Garden Gully confluence. This site is downstream of the confluences with Three Mile Creek and Racecourse Creek.

The catchment model domain does not include any reporting sites downstream of the upper reach of the Eucumbene River.

2.7.3 Result types

i No flow threshold

A 'no flow' threshold was applied to aid interpretation of modelled streamflow. Flows below the threshold are presented on hydrographs; however, zero flow was assumed to occur for modelled flows less than 0.1 ML/day (corresponding to approximately 1 L/s) in the calculation of statistics describing the streamflow regime.

The purpose of the no flow threshold is not to state definitively that streams cease to flow below this level, but rather to indicate that flows are very small and to reflect lower confidence in the ability of the catchment model to predict streamflow below this level.

Periods where there is no flow are an important ecological metric. Use of a no flow threshold provides clarity around the assessment and reporting of this metric.

ii Flow categories

River flow objectives are used by the NSW Government in the management of environmental flows and set out aspects of flow considered to be critical for the protection or restoration of river health, ecology and biodiversity. The aim of the objectives is to aid in improving river health by recognising the importance of natural river flow patterns (NSW Department of Environment, Climate Change and Water, 2006). Several flow categories are defined within the descriptions of the river flow objectives. The flow categories used in this analysis are:

- Very low flows: flows below the level naturally exceeded on 95% of all days with flow;
- Low flows: flows below the level naturally exceeded on 80% of all days with flow; and
- High flows: flows that are greater than the level naturally exceeded on 30% of all days with flow.

Flows falling between low flows and high flows are termed "medium flows" for reporting purposes.

For each reporting site, the flow category thresholds were calculated using the modelled pre-construction flow, over the full modelled period.

In addition to the flow categories listed, a "no flow" category was also assessed such that zero flow was assumed to occur for modelled flows less than 0.1 ML/day.

Flow categories relating to river height (eg freshes and floods) have not been used as river height is not a result produced by the surface water model. Freshes and floods are primarily driven by quickflow response to heavy rain, and so the frequency of freshes and floods is not likely to be affected by the predicted changes to baseflow.

iii Flow duration curves

Flow duration curves, also called probability of exceedance curves, are provided for each reporting site. They show the probability that a given streamflow will be exceeded on any given day and, conversely, they show the streamflow corresponding to a given probability of exceedance (ie median streamflow). It is important to note that statistics relating to streamflow must be treated with caution as they only relate to the modelled (or measured) period and this does not reflect the full range of stream flows that could potentially occur at a location.

Flow duration curves do not have a time dimension. Therefore, seasonal flow duration curves are useful in providing information about the flow regime on a seasonal basis (eg to tie in with temporal ecological requirements).

iv Hydrographs

Daily hydrograph samples are provided for each reporting site to illustrate the impact of the tunnel excavation and operation of the power waterway on streamflow over time.

Through the construction period, hydrographs are shown for the chosen wet, dry and average climate sequences modelled as coinciding with the final two years of construction.

When reporting the impacts of the ongoing operation of the project, hydrographs are shown using 2006 climate data as the lowest yearly flow at the Murrumbidgee gauge was recorded in 2006, and this year thus highlights the impact of baseflow reduction on streamflow.

2.7.4 Pre-construction simulation

i Description

The pre-construction simulation represents the streamflow regime prior to any project impacts, with no changes to baseflow due to the tunnel excavation or operation of the power waterway. This simulation used the calibrated catchment model with no leakage term applied. The modelled period was 1 Jun 1970–30 May 2019, which coincides with the calibration period adjusted to include an equal number of each season.

The pre-construction simulation was used to produce inputs to the groundwater modelling process (Figure 1.2):

- a groundwater recharge dataset for input to the groundwater model; and
- a baseflow discharge dataset which was used to validate the calibration of the groundwater model.

The pre-construction simulation provides a baseline against which the operating and construction simulation were compared.

ii Results

a Gooandra Creek catchment

Annual flows (July to June) for the reporting sites in and downstream of Gooandra Creek are shown in Figure 2.51.

Annual flow varies significantly from year to year; at all sites, 2006 was the lowest flow year and 1974 the highest. Modelled results indicate lower than average flow conditions from 1997 to the present (2019), with only 5 out of 21 years having above average flow.

Site 3 has a modelled average flow of approximately 10 GL/year.

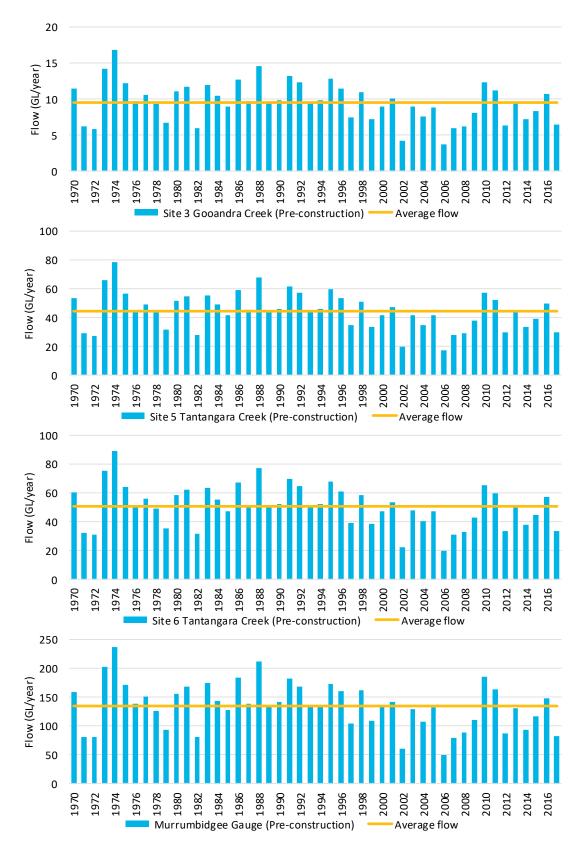


Figure 2.51 Annual stream flows (Gooandra Creek reporting sites)

Site 5 and Site 6 are relatively close together and have modelled average flows 4–5 times higher than those at Site 3–45 GL/year and 51 GL/year respectively. These sites are on Tantangara Creek downstream of the confluence with Gooandra Creek and include flows from the upstream Tantangara Creek catchment area.

The Murrumbidgee Gauge has a modelled average annual flow of 135 GL/year, an order of magnitude higher than flows in Gooandra Creek. Flows at this location include Tantangara Creek flows as well as flows from a large catchment area to the north of Tantangara Creek.

Seasonal streamflow is shown in Figure 2.52. For all sites, flows are higher and more variable in winter and spring, when localised precipitation peaks and accumulated snow melts. Flows are lower and less variable in summer and autumn, when climate conditions are drier and baseflows predominate.

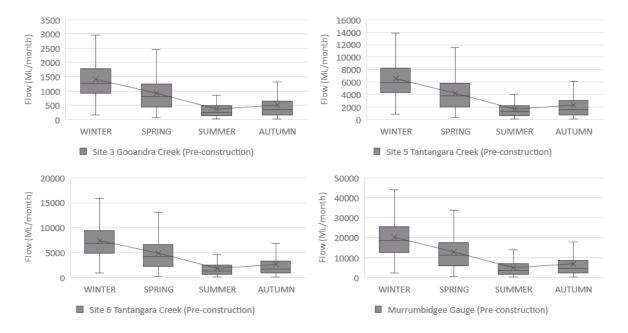


Figure 2.52 Seasonal stream flows (Gooandra Creek reporting sites)

Flow duration curves showing daily flows over the full modelled period for the Gooandra Creek reporting sites are shown in Figure 2.53.

Although Gooandra Creek is characterised as having a perennial flow regime (Annexure A of the water assessment), Figure 2.53 indicates that flows in Gooandra Creek (reporting site 3) fall below the no flow threshold of 0.1 ML/day on approximately 3% of modelled days. This indicates that the flow regime in at Site 3 is vulnerable to reductions in baseflow.

The shape of the flow duration curve is the same for these reporting sites (ie the curves are parallel) because the sub catchments use the same rainfall-runoff model and model parameters and rainfall does not vary significantly between them.

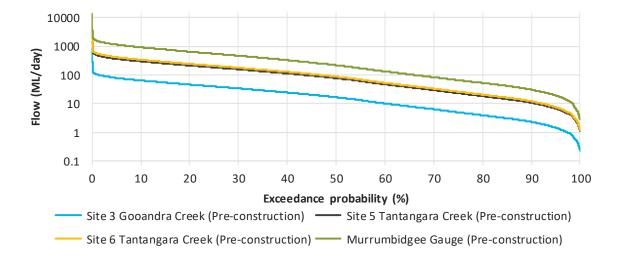


Figure 2.53 Modelled flow duration curves (Gooandra Creek reporting sites)

Flows corresponding to the flow categories described in section 2.7.3ii are given in Table 2.13 for the Gooandra Creek reporting sites.

These were determined using results over the full modelled period and were used to assess changes in the flow regime due to the tunnel excavation and operation of the power waterway.

Table 2.13 Flow categories for Gooandra Creek reporting sites (ML/day)

Flow category	Site 3	Site 5	Site 6 Murrumbidgee		
	Gooandra Creek	Tantangara Creek	Tantangara Creek	Gauge	
Very low flows ¹	1.4	6.7	7.6	19.4	
Low flows ²	3.9	18.3	20.7	53.1	
High flows ³	34.1	158.8	180.3	471.1	

Note:

- 1. Flows below the level naturally exceeded on 95% of all days with flow
- 2. Flows below the level naturally exceeded on 80% of all days with flow
- 3. Flows that are greater than the level naturally exceeded on 30% of all days with flow

b Eucumbene River catchment

Annual flows (July to June) for the reporting sites in the Eucumbene River are shown in Figure 2.54.

Site 10 and Site 9 have modelled average flows of approximately 3 GL/year and 1.2 GL/year respectfully. These sites are at the outlet of headwater catchments with very small catchment areas (4 km² and 1.6 km²).

Site 11 is on the Eucumbene River where the it crosses the Gooandra Trail. It has a modelled average flow of approximately 8.5 GL/year, which includes flows from Site 10 and Site 9.

Site 2 has a modelled average annual flow of 25 GL/year, including upstream flows in the Eucumbene and from Three Mile Creek and Racecourse Creek.

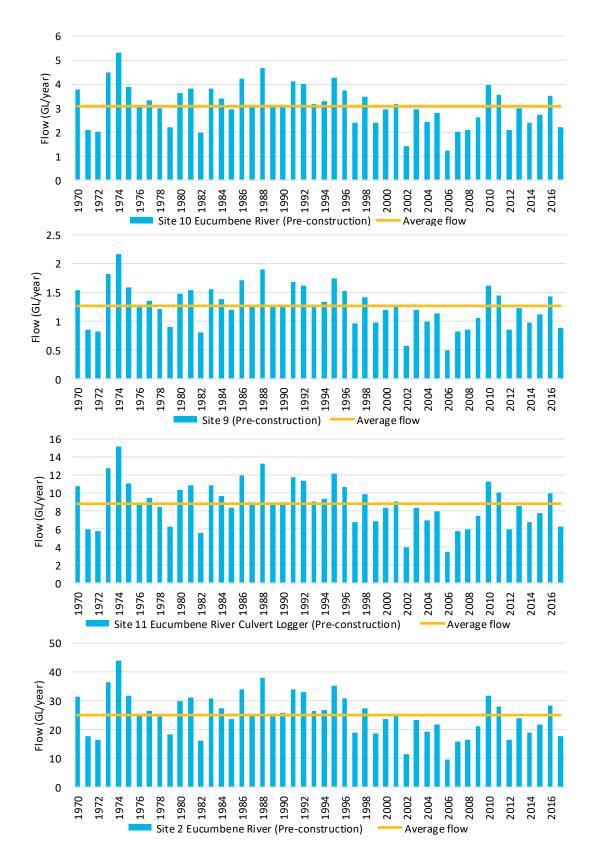


Figure 2.54 Annual stream flows (Eucumbene River reporting sites)

Seasonal streamflow are shown in Figure 2.55. For all sites, modelled flows are higher and more variable in winter and spring, when localised precipitation peaks and accumulated snow melts. Flows are lower and less variable in summer and autumn, when climate conditions are drier and baseflows predominate.

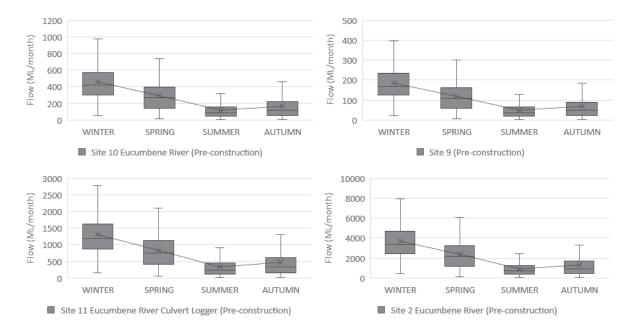


Figure 2.55 Seasonal stream flows (Eucumbene River reporting sites)

Flow duration curves showing daily flows over the full modelled period for the Eucumbene River reporting sites are shown in Figure 2.56.

Creeks in headwater catchments in the Eucumbene River upper reaches, which includes Site 10 and Site 9, are characterised as having a non-perennial flow regime (ie ephemeral). This is validated by the flow duration curves for these sites, which indicate that flows at Site 10 and Site 9 fall below the no flow threshold of 0.1 ML/day on approximately 14% and 32% of modelled days respectively.

Although the main channel of the Eucumbene River is characterised as having a perennial flow regime (Annexure A of the water assessment), Figure 2.56 indicates that modelled flows at Site 11 fall below the no flow threshold of 0.1 ML/day on approximately 3% of modelled days.

Modelled flows at Site 2 do not fall below the no flow threshold.

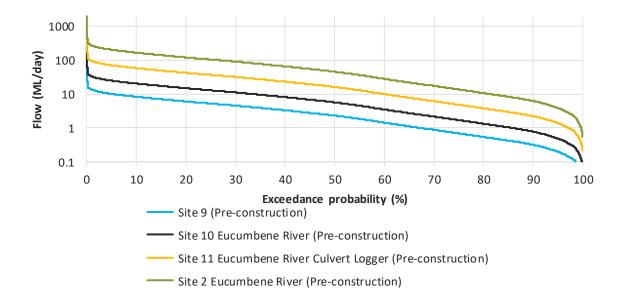


Figure 2.56 Modelled flow duration curves (Eucumbene River reporting sites)

Flows corresponding to the flow categories, described in section 2.7.3ii, are given in Table 2.14 for the Eucumbene River reporting sites.

These were determined using results over the full modelled period and were used to assess changes in the flow regime due to the tunnel excavation and operation of the power waterway.

Table 2.14 Flow categories for Gooandra Creek reporting sites (ML/day)

Flow category	Site 10 Eucumbene	Site 9	Site 11	Site 2 Eucumbene River	
	River		Eucumbene River	Eucumbene River	
Very low flows ¹	0.5	0.2	1.4	3.9	
Low flows ²	1.3	0.6	3.8	10.7	
High flows ³	11.2	4.6	31.9	90.9	

Note:

- 1. Flows below the level naturally exceeded on 95% of all days with flow
- 2. Flows below the level naturally exceeded on 80% of all days with flow
- 3. Flows that are greater than the level naturally exceeded on 30% of all days with flow

2.7.5 Streamflow changes during construction

i Description

The power waterway, power station, and associated tunnels and shafts will experience groundwater inflow during construction (section 3.4.4i). As the actual climate that will occur during construction is unknown, possible impacts were assessed with the groundwater and catchment models using wet, average and dry climate sequences. The catchment model utilised the same climate sequences as the groundwater model.

The average, wet and dry climate sequences were chosen through an assessment of streamflow data recorded in the Murrumbidgee River (gauge 410535) (Figure 2.57).

The Victorian Government reports that recorded climate data throughout south-eastern Australia indicates that there may have been a 'climate step-change' around 1997 (Victrorian Department of Environment, Land, Water and Planning, 2016, p. 6), and that utilising post-1997 climate data averages is appropriate for planning studies. Through the record period, the average yearly flow recorded at the gauge was 142 GL/year, while the post-1997 average yearly flow was 131 GL/year. The post-1997 average flow was used for this assessment. 'Average' rainfall conditions were simulated using climate data from December 2012 to December 2017. Through this period the average Murrumbidgee River flow was 127 GL/year, close to the post 1997 average of 131 ML/year. During this climate sequence there were high flows in 2015, and lower than average flows through the other years. Other possible 'average' historical climate sequences similarly feature a mixture of wet and dry years (Figure 2.57).

'Wet' rainfall conditions were simulated using climate data from December 1988 to December 1993. Through this period the average Murrumbidgee River flow was 188 GL/year, and each year experienced above average flow. This climate sequence includes 1990, during which year extensive flooding occurred in NSW.

'Dry' climate conditions were simulated using climate data from December 2001 to December 2006. Through this period the average Murrumbidgee River flow was 96 GL/year. Four of these years experienced lower than average streamflow, and one year experienced above average streamflow. This climate sequence includes the driest years of the Millennium Drought.

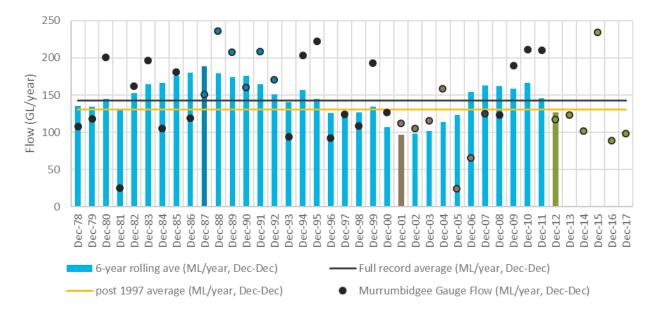


Figure 2.57 Murrumbidgee River (410535) total yearly and average 6 yearly flow (average of following years)

The quarterly (seasonal) baseflow losses predicted by the groundwater model were used to select 'groundwater leakage' rates for each catchment model subcatchment for each quarter. Leakage rates were selected such that the resulting baseflow reduction percent in the catchment model closely matched the baseflow reduction percent predicted by the groundwater model.

The groundwater model predicted that impacts to creek and river baseflow would develop over time, with the largest impacts to stream baseflow seen after construction is complete. Baseflow reduction due to tunnelling and excavation works during the construction period was predicted in Gooandra Creek and the headwaters of the Eucumbene River. The timing of the baseflow reduction will depend on the project schedule, as drawdown impacts are predicted to peak after the tunnel excavation reaches the Gooandra Volcanics, which occur in the vicinity of Gooandra Creek and the Eucumbene River headwaters. If no delays to schedule occur, Gooandra Creek baseflow

reduction could begin during year 4 of construction, and Eucumbene River baseflow reductions could begin in year 5 of construction (Figure 2.58).

The catchment model contains several Eucumbene River subcatchments. A groundwater leakage term was applied only within the Eucumbene River headwater subcatchments (model sub catchments 14 and 15) as these are located within the extent of predicted groundwater drawdown.

The groundwater model predicted that, during construction, Gooandra Creek baseflow may decline by up to approximately 6%. Baseflow for the portion of the Eucumbene River contained within the model domain was predicted to decline by up to approximately 1%. Impacts were predicted to be still developing at the end of the construction period (Figure 2.58). Long-term (steady state) impacts are presented in section 2.7.6.

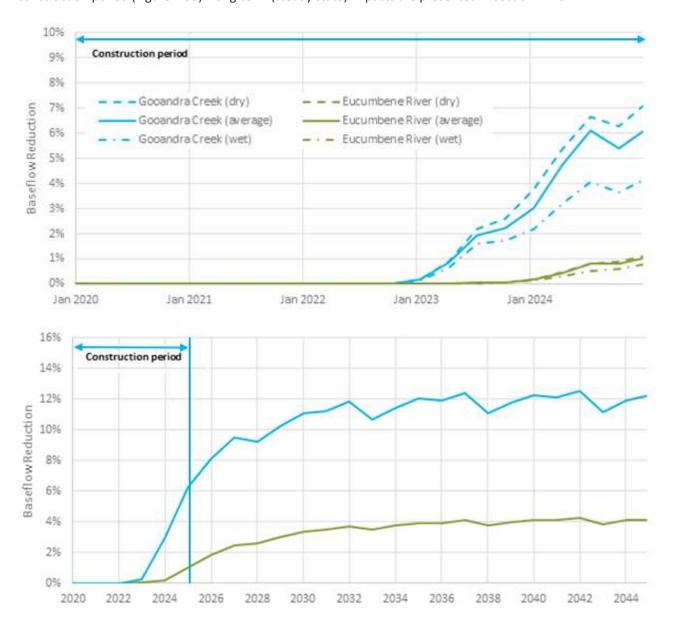


Figure 2.58 Baseflow reduction predicted by the groundwater model during project construction

The predicted declines in baseflow were modelled in the catchment model using the groundwater leakage term, with leakage rates at the end of the construction period approaching 0.025 mm/day in both the Gooandra Creek and Eucumbene River headwater catchments (Figure 2.59). A steadily increasing leakage rate was utilised as the impacts of tunnelling works will increase as the excavation progresses.

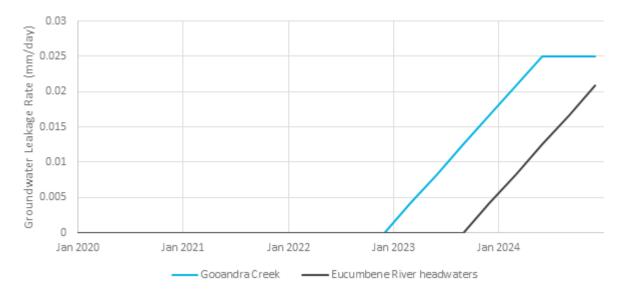


Figure 2.59 Groundwater leakage rate applied during construction years in the catchment model

ii Results

No impacts to baseflow due to tunnel excavation were predicted within creek catchments other than Gooandra Creek and the Eucumbene River north of the Snowy Highway.

The baseflow reduction in Gooandra Creek during the excavation of the power waterway is expected to cause no discernible changes to streamflow through winter months (Figure 2.60 to Figure 2.62). During March—April in the final two years of excavation baseflow reduction may result in reduced flow within the Gooandra Creek catchment if the those construction years coincide with dry climate conditions (Figure 2.62).

Within the Eucumbene River, baseflow reduction during the construction period is expected to cause no discernible changes to streamflow. Streamflow hydrographs for the modelled climate sequences are presented for sites 9, 10, and 11 within the Eucumbene River catchment in Attachment D.

Inflows to the tunnel excavation are predicted to increase markedly during the groundwater model year 2023, rising to approximately 60 L/s when the tunnel encounters the Gooandra Volcanics and then stabilising in 2024 (Figure **3.52**). Through the final quarter of construction, the baseflow impacts within the Gooandra and Eucumbene catchments were estimated to be in the order of 10 L/s, significantly less than the tunnel inflows. Impacts to baseflow within the Gooandra Creek catchment and within the Eucumbene River catchment upstream of Gooandra Track were predicted to increase over the final years of the construction period Figure 2.58), indicating a lag between the greatest tunnel inflow occurring and the greatest baseflow impacts. The peak change in baseflow is expected to occur following completion of the project, discussed in section 2.7.6.



Figure 2.60 Modelled Gooandra Creek flow during the final two years of the construction period using a 'wet' climate sequence

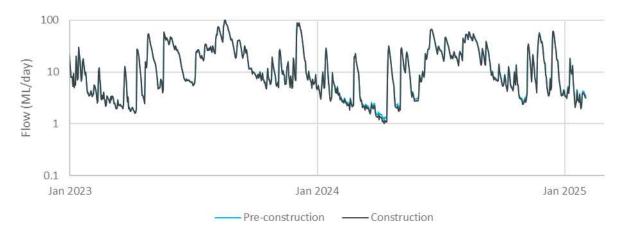


Figure 2.61 Modelled Gooandra Creek flow during the final two years of the construction period using an 'average' climate sequence

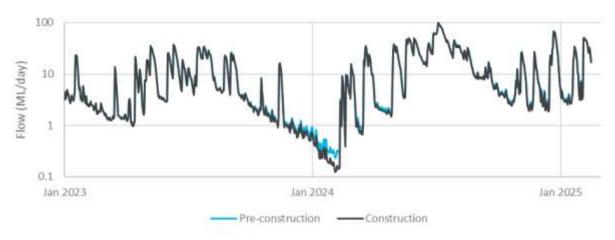


Figure 2.62 Modelled Gooandra Creek flow during the final two years of the construction period using a 'dry' climate sequence

2.7.6 Streamflow changes during operation

i Description

The long-term impact of tunnel excavation and operation of the power waterway was modelled using the results of the steady state groundwater model (groundwater model run: 'SH4.0_sspred24'). This simulation represents the streamflow regime with permanent reductions to baseflow, described as the 'medium' inflow scenario. This is a realistic worst case scenario, as the scenario still represents a lined but unmitigated tunnel (ie concrete segmental lining but no grouting). Snowy Hydro consider that once grouting is applied (which is part of the project but not modelled), the inflow numbers will be less than what the "medium" scenario inflow numbers are.

The model period was 1 Jun 1970–30 May 2019, which is the same modelled period as for the pre-construction simulation (to allow for comparison).

Baseflow reductions predicted by the groundwater model for the river reaches named in Figure 2.63 are given in Table 2.15. With the exception of the Eucumbene River, reductions of less than 5% were not modelled in the catchment model; therefore, impacts were only applied to Gooandra Creek and Eucumbene River.

Based on inspection of the drawdown contours produced using the groundwater model results (need to update this figure, or put new one in for "medium" scenario Figure 2.50), only the two headwater catchments of the upper reach of the Eucumbene are predicted to be impacted by the tunnel excavation.

Table 2.15 Baseflow reductions during operation of the power waterway

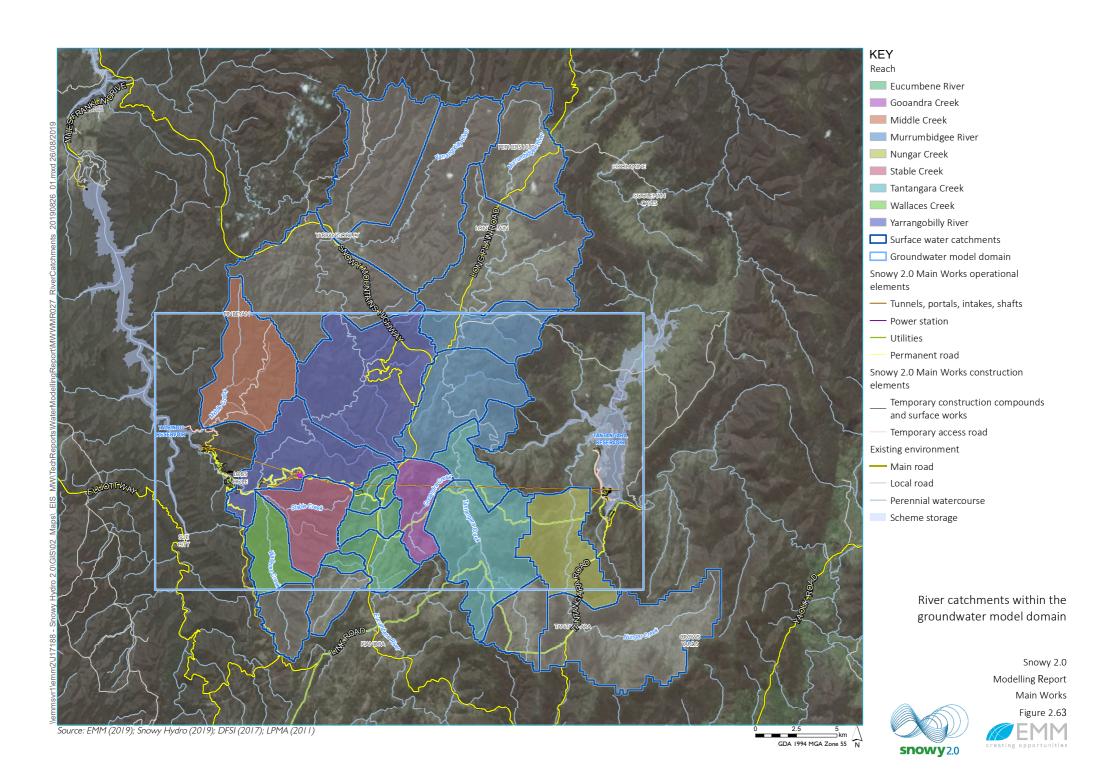
Middle	Yarrangobilly	Wallaces	Stable Creek	Eucumbene	Murrumbidgee	Tantangara	Gooandra	Nungar
Creek	River	Creek		River	River	Creek	Creek	Creek
0.1%	2.5%	0.4%	4.9%	3.9%	0.1%	1.0%	11.1%	0.7%

A constant rate was used for the leakage rate in each sub-catchment as the groundwater model water budget showed relatively constant rates of flow to the tunnel (update this reference Table 3.7).

Leakage rates for each sub catchment were determined by iterative adjustment until the baseflow impact in the catchment model over the modelled period 2007–2012 matched the baseflow impact predicted by the groundwater model for each region. The period 2007–2012 was used because it was the same period as climate inputs used in the steady state groundwater model run. The leakage rates applied are recorded in Table 2.16.

Table 2.16 Leakage rates during operation of the power waterway

	Gooandra Creek	Eucumbe	ne River
	Upstream of Site 3	Upstream of Site 9	Upstream of Site 10
	Catchment 11	Catchment 14	Catchment 15
Leakage rate (mm/day):	0.078	0.127	0.127



ii Results

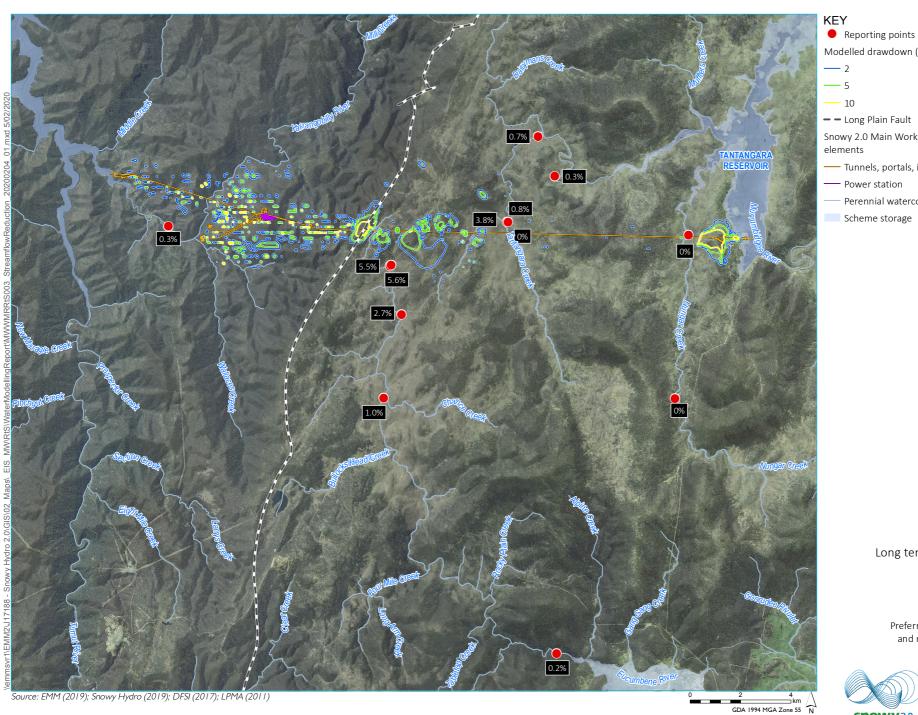
Baseflow reductions caused by the tunnel excavation and operation of the power waterway are expected to have a noticeable impact on the streamflow regime:

- in smaller headwater catchments, as:
 - reduced baseflow may lead to cease-to-flow conditions during dry periods; and
 - these catchments do not benefit from flow from unaffected catchment areas further upstream;
- during summer and autumn, when climate conditions are drier and baseflows are a large part of the total flow.

Predicted reductions in total and seasonal streamflow over the full modelled period (1 Jun 1970–30 May 2019) are given in Table 2.17 for the reporting sites. The average yearly streamflow reductions are illustrated in Figure 2.64. Figure 2.64 also includes the estimated total streamflow reduction at the Eucumbene River gauge 222522 which lies outside the model domain based on the streamflow recorded at that gauge and the magnitude of baseflow loss the models predicted within the catchment upstream.

Table 2.17 Predicted reduction in streamflow (yearly average and by season)

	Gooandra Creek reporting sites				E	ucumbene Riv	er reporting sit	es
	Site 3	Site 5	Site 6	Murrumbidgee Gauge	Site 10	Site 9	Site 11	Site 2
Average	-3.8%	-0.8%	-0.7%	-0.3%	-5.5%	-5.6%	-2.7%	-1.0%
Summer	-7.2%	-1.6%	-1.4%	-0.5%	-10.6%	-10.6%	-5.2%	-1.9%
Autumn	-5.9%	-1.3%	-1.1%	-0.4%	-8.5%	-8.6%	-4.2%	-1.5%
Winter	-2.3%	-0.5%	-0.4%	-0.2%	-3.4%	-3.4%	-1.7%	-0.6%
Spring	-3.4%	-0.7%	-0.6%	-0.2%	-5.2%	-5.2%	-2.6%	-0.9%



Modelled drawdown (m)

___ 10

─ Long Plain Fault

Snowy 2.0 Main Works operational elements

— Tunnels, portals, intakes, shafts

Power station

Perennial watercourse

Scheme storage

Long term total streamflow reduction

> Snowy 2.0 Modelling report Preferred infrastructure report and response to submissions Main Works Figure 2.64





a Gooandra Creek catchment

Hydrographs for the reporting sites in Gooandra Creek catchment (for 2006, the lowest flow year on record) are shown in Figure 2.65.

Site 3 on Gooandra Creek shows the largest predicted impact on streamflow. Peak flows are reduced, and long periods of "no flow" can be observed in the operating phase hydrograph.

For Site 5 and Site 6 on Tantangara Creek, the predicted impact of the operation and the reduction of baseflows in the upstream Gooandra Creek catchment is much less pronounced and is barely discernible on the hydrograph as the flows from the upstream Tantangara Creek catchment area were not predicted to experience baseflow reduction.

At the Murrumbidgee gauge, the predicted impact is further reduced due to the flows from the large catchment area of the Murrumbidgee to the north. The impact of operation on stream flows is barely discernible on the hydrograph at the Murrumbidgee gauge.

Approximately 4.5 km of Gooandra Creek immediately upstream of the confluence with Tantangara creek is expected to be impacted by baseflow reduction due to groundwater drawdown (check and update reference Figure 2.50).

Total and seasonal flow duration curves for Gooandra Creek reporting site 3 are show in Figure 2.66 and Figure 2.67. Flow duration curves are not shown for the other reporting sites (Site 5, Site 6 and Murrumbidgee Gauge) because impacts are too small to be visualised in this format (indicating that impacts would be effectively impossible to discern from the measured streamflow data).

The flow duration curves for Site 3 show that the low flow regime is most affected by the reduction in baseflow. Where the curve for the pre-construction case showed a perennial flow regime, the curve for the operating case shows the low flow portion of the curve dropping off sharply, with 10% of modelled days below the no flow threshold. This indicates a shift to a more ephemeral flow regime for this site; particularly in Summer and Autumn. Seasonal flow duration curves for Winter and Spring showed a much less pronounced increase in no flow days.