



APPENDIX F.6 UPDATED BDAR

PCT 540 - Silvertop Stringybark - Ribbon Gum - Rough-barked Apple open forest on basalt hills of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion



Vegetation formation: Dry Sclerophyll Forests (Shrub/grass formation)

Vegetation class: New England Dry Sclerophyll Forests

Conservation status:

- EPBC Act: Not listed
- BC Act: Not listed

Extent within development footprint: 67.47 hectares

No. BAM plots: 9

Recorded species:

- Canopy: *Eucalyptus laevopinea*, *Eucalyptus viminalis*, *Angophora floribunda*
- Shrub: *Bursaria spinosa*, *Cassinia laevis*.
- Ground: *Ajuga australis*, *Aristida ramos*, *Arthropodium milleflorum*, *Carex inversa*, *Brachyscome spathulata*, *Cheilanthes sieberi* subsp. *sieberi*, *Cymbopogon refractus*, *Dianella longifolia*, *Dichondra repens*, *Echinopogon mckiei*, *Glycine microphylla*, *Lomandra confertifolia*, *Rytidosperma penicillatum*, *Poa sieberiana*.

Justification of PCT: PCT commonly recorded on mid to high hills below the ridgeline, generally with a north-facing aspect. *Eucalyptus laevopinea* is dominant in the canopy with *Eucalyptus viminalis*, and *Angophora floribunda* sub-dominant. Understorey vegetation was found to be generally open, with native species most common and diverse in the ground layer.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
493	Forest Oak - Rough-barked Apple - Silvertop Stringybark shrub grass open forest on protected slopes of the Liverpool Range	<ul style="list-style-type: none"> PCT dominated by Forest Oak which was only occasionally recorded within the PCT within the subject land Notes as a very restricted community where Forest Oak dominates forests
541	Silvertop Stringybark - Rough-barked Apple grassy open forest of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion	<ul style="list-style-type: none"> <i>Eucalyptus nortonii</i> more common in PCT 541, with <i>Eucalyptus viminalis</i> more common in PCT 540 Understorey cover was generally found to be more dense in PCT 41, with tree cover higher in PCT 540
542	Stringybark - Rough-barked Apple - cypress pine shrubby open forest of the eastern Nandewar Bioregion and western New England Tableland Bioregion	<ul style="list-style-type: none"> Cypress Pine not recorded within the subject land PCT geographical extent does not is not representative of the subject land
565	Silvertop Stringybark - Mountain Gum grassy open forest of the New England Tableland Bioregion	<ul style="list-style-type: none"> Mountain Gum not recorded within the PCT within the subject land PCT usually supports a sparse to mid-dense shrub/small tree layer, not representative of the PCT within the subject land
1164	Silvertop Stringybark - Mountain Ribbon Gum open forest of the New England Tableland Bioregion	<ul style="list-style-type: none"> Characteristic canopy species are not representative of the PCT within the subject land
1223	Stringybark shrubby open forest of the north east parts of the New England Tableland Bioregion	<ul style="list-style-type: none"> Restricted to the far north-east part of the tablelands, for example Boonoo National Park and Basket Swamp Nature Reserve
1396	Rough-barked Apple - Silvertop Stringybark - Ribbon Gum shrub/grass open forest on hills of the southern Nandewar Bioregion	<ul style="list-style-type: none"> Characteristic canopy species are not representative of the PCT within the subject land

PCT 541 - Silvertop Stringybark - Rough-barked Apple grassy open forest of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion



Vegetation formation: Dry Sclerophyll Forests (Shrub/grass formation)

Vegetation class: New England Dry Sclerophyll Forests

Conservation status:

- EPBC Act: Not listed
- BC Act: Not listed

Extent within development footprint: 30.85 hectares

No. BAM plots: 8

Recorded species:

- Canopy: *Angophora floribunda*, *Eucalyptus blakelyi*, *Eucalyptus laevopinea*, *Eucalyptus nortonii*
- Shrub: *Acacia implexa*, *Cassinia laevis*, *Melicytus dentatus*, *Solanum aviculare*
- Ground: *Crassula sieberiana*, *Dichondra repens*, *Echinopogon mckiei*, *Einadia nutans*, *Einadia trigonos*, *Eustrephus latifolius*, *Galium aparine*, *Geranium solanderi* var. *solanderi*, *Glycine microphylla*, *Microlaena stipoides*, *Mentha diemenica*, *Rytidosperma leave*.

Justification of PCT: PCT commonly recorded as a mosaic with PCT 540 on mid to high hills below the ridgeline, generally with a north-facing aspect. *Eucalyptus laevopinea* is dominant in the canopy with *Eucalyptus nortonii*, and *Angophora floribunda* sub-dominant. Understory is grassy with shrub cover generally found to be quite low.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
493	Forest Oak - Rough-barked Apple - Silvertop Stringybark shrub grass open forest on protected slopes of the Liverpool Range	<ul style="list-style-type: none"> PCT dominated by Forest Oak which was only occasionally recorded within the PCT within the subject land Notes as a very restricted community where Forest Oak dominates forests
540	Silvertop Stringybark - Ribbon Gum - Rough-barked Apple open forest on basalt hills of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion	<ul style="list-style-type: none"> <i>Eucalyptus nortonii</i> more common in PCT 541, with <i>Eucalyptus viminalis</i> more common in PCT 540 Understorey cover was generally found to be more dense in PCT 541, with tree cover higher in PCT 540
542	Stringybark - Rough-barked Apple - cypress pine shrubby open forest of the eastern Nandewar Bioregion and western New England Tableland Bioregion	<ul style="list-style-type: none"> Cypress Pine not recorded within the subject land PCT geographical extent does not is not representative of the subject land
565	Silvertop Stringybark - Mountain Gum grassy open forest of the New England Tableland Bioregion	<ul style="list-style-type: none"> Mountain Gum not recorded within the PCT within the subject land PCT usually supports a sparse to mid-dense shrub/small tree layer, not representative of the PCT within the subject land
1164	Silvertop Stringybark - Mountain Ribbon Gum open forest of the New England Tableland Bioregion	<ul style="list-style-type: none"> Characteristic canopy species are not representative of the PCT within the subject land
1223	Stringybark shrubby open forest of the north east parts of the New England Tableland Bioregion	<ul style="list-style-type: none"> Restricted to the far north-east part of the tablelands, for example Boonoo National Park and Basket Swamp Nature Reserve
1396	Rough-barked Apple - Silvertop Stringybark - Ribbon Gum shrub/grass open forest on hills of the southern Nandewar Bioregion	<ul style="list-style-type: none"> Characteristic canopy species are not representative of the PCT within the subject land

PCT 586 - Snow Grass - Swamp Foxtail tussock grassland sedgeland of cold air drainage valleys of the New England Tableland Bioregion



Vegetation formation: Grasslands

Vegetation class: Temperate Montane Grasslands

Conservation status:

- EPBC Act: Not listed
- BC Act: Not listed

Extent within development footprint: 2.56 hectares

No BAM plots: 3

Recorded species:

- Canopy: n/a
- Shrub: n/a
- Ground: *Pennisetum alopecuroides*, *Poa sieberiana*, *Carex inversa*, *Rumex brownii*, *Persicaria decipiens*, *Haloragis heterophylla*, *Juncus subsecundus*

Justification of PCT: PCT was found to occur as a grassland / sedgeland / swamp meadow dominated by *Pennisetum alopecuroides* and *Poa sieberiana*. PCT was found to occur along flats, between slopes, at mid elevation along the transverse track section of the development footprint.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
497	Tea tree shrubland / sedgeland / forbland swamp wetland on the Liverpool Range, mainly Brigalow Belt South Bioregion	<ul style="list-style-type: none"> PCT represents a tall shrubland over a low forbland or sedgeland wetland, which is not representative of the PCT within the subject land
569	Derived Snow Grass +/- Kangaroo Grass +/- Wild Sorghum tussock grassland of the NSW Northern Tablelands	<ul style="list-style-type: none"> A broadly defined community covering high altitude derived snow grass pastures, is not representative of the PCT within the subject land

PCT 599 - Blakely's Red Gum - Yellow Box grassy tall woodland on flats and hills in the Brigalow Belt South Bioregion and Nandewar Bioregion



Vegetation formation: Grassy Woodlands

Vegetation class: Western Slopes Grassy Woodlands

Conservation status:

- EPBC Act: White Box-Yellow Box-Blakely's Red Gum Grassy Woodland and Derived Native Grassland
- BC Act: White Box Yellow Box Blakely's Red Gum Woodland

Extent within development footprint: 4.96 hectares

No BAM plot: 0

Characteristic species (NSW BioNet):

- Canopy: *Eucalyptus blakelyi*, *Eucalyptus melliodora*, *Angophora floribunda*, *Eucalyptus albens*, *Eucalyptus melanophloia*, *Brachychiton populneus* subsp. *populneus*, *Eucalyptus conica*, *Callitris glaucophylla*,
- Shrub: *Acacia implexa*, *Geijera parviflora*, *Acacia decora*, *Myoporum montanum*, *Olearia elliptica* subsp. *elliptica*, *Pimelea neo-anglica*, *Maireana microphylla*, *Hibbertia riparia*, *Pimelea curviflora* var. *curviflora*, *Notelaea microcarpa*, *Acacia deanei* subsp. *deanei*,
- Ground: *Aristida personata*, *Glycine tabacina*, *Carex inversa*, *Themeda triandra*, *Cyperus gracilis*, *Austrostipa verticillata*, *Bothriochloa macra*, *Microlaena stipoides* var. *stipoides*, *Aristida ramosa*, *Geranium solanderi* var. *solanderi*, *Hydrocotyle laxiflora*, *Ajuga australis*, *Dichelachne micrantha*, *Lomandra longifolia*, *Lomandra filiformis* subsp. *coriacea*, *Desmodium brachypodum*.

Justification of PCT: PCT commonly mapped in the locality (DPIE 2019, DPIE 2015) with *Eucalyptus albens*, *Eucalyptus blakelyi*, *Eucalyptus melliodora*, *Angophora floribunda* commonly noted as the dominant species in the canopy, *Acacia implexa* and *Notelaea macrocarpa* dominant in the generally sparse midstorey, with *Themeda triandra*, *Microlaena stipoides* var. *stipoides*, *Dichelachne micrantha* and *Lomandra filiformis* subsp. *coriacea* dominating parts of the ground layer.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
510	Blakely's Red Gum - Yellow Box grassy woodland of the New England Tableland Bioregion	<ul style="list-style-type: none"> PCT noted as similar to PCT 599, however the locality is noted as surrounding the Mt Kaputar, Horton and Inverell areas.
704	Blakely's Red Gum - Yellow Box grassy open forest or woodland of the New England Tableland Bioregion	<ul style="list-style-type: none"> Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land
1329	Yellow Box - Blakely's Red Gum grassy woodland of the Nandewar Bioregion	<ul style="list-style-type: none"> Landscape position of fertile loamy-clay soils on slopes, drainage lines and alluvial plains is not repetitive of the PCT within the subject land Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land
1332	Yellow Box - Grey Box - Red Gum woodland of the central eastern parts of the New England Tableland Bioregion	<ul style="list-style-type: none"> PCT locality noted as east of Armidale

PCT 931 - Messmate - Mountain Gum tall moist forest of the far southern New England Tableland Bioregion



Vegetation formation: Wet Sclerophyll Forests (Shrubby sub-formation)

Vegetation class: Northern Escarpment Wet Sclerophyll Forests

Conservation status:

- EPBC Act: Not listed
- BC Act: Not listed

Extent within development footprint: 4.45 hectares

No. BAM plots: 5

Recorded species:

- Canopy: *Eucalyptus obliqua*, *Eucalyptus dalrympleana* subsp. *heptantha*, *Eucalyptus viminalis*, *Acacia melanoxylon*
- Shrub: *Lomatia arborescens*, *Melicytus dentatus*,
- Ground: *Calochlaena dubia*, *Cotula australis*, *Dichondra repens*, *Geranium potentilloides*, *Lobelia concolo*, *Microlaena stipoides*, *Poa sieberiana* var. *sieberiana*, *Pteridium esculentum*, *Solanum prinophyllum*

Justification of PCT: PCT commonly recorded on high hills, on or just below the ridgeline moving down into deep gullies, generally with a south-facing aspect. *Eucalyptus obliqua* is dominant in the canopy with *Eucalyptus dalrympleana* subsp. *heptantha*, and *Acacia melanoxylon* sub-dominant. Understorey vegetation was found to be generally dense and closed (compared to the similar PCT934) where the PCT occurred in high condition, with native species common and diverse in the understorey.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
934	Messmate open forest of the tableland edge of the NSW North Coast Bioregion and New England Tableland Bioregion	<ul style="list-style-type: none"> • <i>Eucalyptus nobillis</i> more common in PCT 934, with <i>Eucalyptus dalrympleana</i> more common in PCT 931 • Understorey cover was generally found to be more dense in PCT 934, however tree cover was more sparse, when compared to PCT 931
935	Messmate tall moist open forest of the escarpment ranges, southern New England Tableland Bioregion and NSW North Coast Bioregion	<ul style="list-style-type: none"> • Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land
954	Mountain Ribbon Gum - Messmate open forest of escarpment ranges of the NSW North Coast Bioregion and New England Tableland Bioregion	<ul style="list-style-type: none"> • Characteristic canopy species assemblage are representative of the PCT within the subject land
1555	Mountain Gum - Messmate - Snow Gum grassy open forest of the New England Tablelands	<ul style="list-style-type: none"> • PCT noted as Open forests with a canopy dominated by <i>Eucalyptus dalrympleana</i> which is not representative of the PCT within the subject land
1574	Messmate grassy tall open forest on Barrington and Northern Tablelands escarpment	<ul style="list-style-type: none"> • Characteristic canopy species assemblage are representative of the PCT within the subject land

PCT 934 - Messmate open forest of the tableland edge of the NSW North Coast Bioregion and New England Tableland Bioregion



Vegetation formation: Wet Sclerophyll Forests (Shrubby sub-formation)

Vegetation class: Northern Escarpment Wet Sclerophyll Forests

Conservation status:

- EPBC Act: Not listed
- BC Act: Not listed

Extent within development footprint: 24.60 hectares

No. BAM plots: 4

Recorded species:

- Canopy: *Eucalyptus obliqua*, *Eucalyptus nobilis*, *Eucalyptus pauciflora*.
- Shrub: *Dicksonia antarctica*, *Coprosma quadrifida*, *Daviesia ulicifolia*, *Melicytus dentatus*
- Ground: *Smilax australis*, *Asperula conferta*, *Desmodium varians*, *Dichondra repens*, *Ehrharta calycina*, *Einadia trigonos*, *Galium gaudichaudii*, *Glycine microphylla*, *Glycine tabacina*, *Gonocarpus teucroides*, *Hydrocotyle laxiflora*, *Microlaena stipoides*, *Oxalis perennans*, *Plantago debilis*, *Poa labillardierei*, *Poa sieberiana* var. *sieberiana*, *Pseuderanthemum variabile*, *Pteridium esculentum*, *Rytidosperma carphoides*, *Rytidosperma laeve*, *Urtica incisa*, *Veronica plebeian*, *Viola betonicifolia*, *Wahlenbergia gracilis*.

Justification of PCT: PCT commonly recorded on high hills, on or just below the ridgeline generally with a south-facing aspect. *Eucalyptus obliqua* is dominant in the canopy with *Eucalyptus nobilis* and *Eucalyptus pauciflora* sub-dominant. Understorey vegetation was found to be generally more open (compared to the similar PCT931) where the PCT occurred in high condition, with native species common and diverse in the understorey. A large area of this PCT occurs as a derived native grassland/shrubland in the central portion of the development footprint following historical clearing and ongoing low intensity farms landuse.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
931	Messmate - Mountain Gum tall moist forest of the far southern New England Tableland Bioregion	<ul style="list-style-type: none"> • <i>Eucalyptus nobillis</i> more common in PCT 934, with <i>Eucalyptus dalrympleana</i> more common in PCT 931 • Understorey cover was generally found to be more dense in PCT 934, however tree cover was more sparse, when compared to PCT 931
935	Messmate tall moist open forest of the escarpment ranges, southern New England Tableland Bioregion and NSW North Coast Bioregion	<ul style="list-style-type: none"> • Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land
954	Mountain Ribbon Gum - Messmate open forest of escarpment ranges of the NSW North Coast Bioregion and New England Tableland Bioregion	<ul style="list-style-type: none"> • Characteristic canopy and understorey species assemblages are representative of the PCT within the subject land
1555	Mountain Gum - Messmate - Snow Gum grassy open forest of the New England Tablelands	<ul style="list-style-type: none"> • PCT noted as Open forests with a canopy dominated by <i>Eucalyptus dalrympleana</i> which is not representative of the PCT within the subject land
1574	Messmate grassy tall open forest on Barrington and Northern Tablelands escarpment	<ul style="list-style-type: none"> • Characteristic canopy species and ground-layer assemblages are representative of the PCT within the subject land

PCT 954 - Mountain Ribbon Gum - Messmate open forest of escarpment ranges of the NSW North Coast Bioregion and New England Tableland Bioregion

Vegetation formation: Wet Sclerophyll Forests (Grassy sub-formation)

Vegetation class: Northern Escarpment Wet Sclerophyll Forests

Conservation status:

- EPBC Act: Not listed
- BC Act: Not listed

Extent within development footprint: 1.23 hectares

No. BAM plots: 0

Characteristic species (NSW BioNet):

- Canopy: *Eucalyptus nobilis*, *Eucalyptus obliqua*, *Eucalyptus campanulata*, *Eucalyptus viminalis*.
- Shrub: *Acacia falciformis*, *Acacia melanoxylon*, *Banksia integrifolia* subsp. *monticola*, *Bursaria spinosa* subsp. *spinosa*, *Leucopogon lanceolatus*, *Podolobium ilicifolium*, *Rubus parvifolius*.
- Ground: *Acaena novae-zelandiae*, *Desmodium varians*, *Dichondra repens*, *Galium propinquum*, *Geranium potentilloides*, *Glycine clandestina*, *Gonocarpus teucroides*, *Hibbertia scandens*, *Lomandra longifolia*, *Poa sieberiana* var. *sieberiana*, *Pteridium esculentum*, *Themeda australis*, *Wahlenbergia stricta*, *Viola betonicifolia*, *Viola hederacea*.

Justification of PCT: PCT present on upper slopes with a northern aspect in the western portion of the wind farm corridor. Species present include *Eucalyptus obliqua* and *Eucalyptus nobilis*.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
931	Messmate - Mountain Gum tall moist forest of the far southern New England Tableland Bioregion	<ul style="list-style-type: none"> • Characteristic canopy and understorey species assemblages are representative of the PCT within the subject land
934	Messmate open forest of the tableland edge of the NSW North Coast Bioregion and New England Tableland Bioregion	<ul style="list-style-type: none"> • Characteristic canopy and understorey species assemblages are representative of the PCT within the subject land
954	Mountain Ribbon Gum - Messmate open forest of escarpment ranges of the NSW North Coast Bioregion and New England Tableland Bioregion	<ul style="list-style-type: none"> • Characteristic canopy and understorey species assemblages are representative of the PCT within the subject land
1555	Mountain Gum - Messmate - Snow Gum grassy open forest of the New England Tablelands	<ul style="list-style-type: none"> • PCT noted as Open forests with a canopy dominated by <i>Eucalyptus dalrympleana</i> which is not representative of the PCT within the subject land
1574	Messmate grassy tall open forest on Barrington and Northern Tablelands escarpment	<ul style="list-style-type: none"> • Characteristic canopy species and ground-layer assemblages are representative of the PCT within the subject land

PCT 1194 - Snow Gum - Mountain Gum - Mountain Ribbon Gum open forest on ranges of the NSW North Coast Bioregion and eastern New England Tableland Bioregion



Vegetation formation: Wet Sclerophyll Forests (Grassy sub-formation)

Vegetation class: Northern Tableland Wet Sclerophyll Forests

Conservation status:

- EPBC Act: Not listed
- BC Act: Ribbon Gum—Mountain Gum—Snow Gum Grassy Forest/Woodland of the New England Tableland Bioregion (when present within the New England Tablelands IBRA Bioregion)

Extent within development footprint: 43.77 hectares

No. BAM plots: 14

Recorded species:

- Canopy: *Eucalyptus viminalis*, *Eucalyptus pauciflora*, *Eucalyptus nobilis*
- Shrub: *Acacia melanoxylon*, *Melicytus dentatus*, *Lomatia arborescens*, *Bursaria spinosa*
- Ground: *Anthoxanthum odoratum*, *Asperula conferta*, *Dichondra repens*, *Geranium homeanum*, *Hybanthus monopetalus*, *Microlaena stipoides*, *Poa sieberiana*, *Pteridium esculentum*, *Veronica plebeia*, *Stellaria pungens*, *Urtica incisa*, *Lomandra longifolia*

Justification of PCT: The PCT was found to commonly occur on slopes and plateaux at high elevation with the canopy a co-dominant mix of *Eucalyptus viminalis*, *Eucalyptus pauciflora*, and *Eucalyptus nobilis*. The understorey was found to be generally open and grassy with the ground layer supporting the highest diversity in native species recorded.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
525	Mountain Gum - Snow Gum grassy open forest at high altitudes in the Kaputar area of the Nandewar Bioregion	<ul style="list-style-type: none"> PCT noted as occurring on clay loam soil derived from basalt or trachyte substrate at high altitudes on plateau in a mountain landform pattern in Mount Kaputar National Park in the Nandewar Bioregion.
554	Ribbon Gum - Mountain Gum - Snow Gum grassy open forest or woodland of the New England Tableland Bioregion	<ul style="list-style-type: none"> PCT description notes Guyra Plateau almost 200km north of the subject land. Description notes the shrub layer is absent to very sparse which is not representative of the PCT within the subject land. Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land
1104	Ribbon Gum - Mountain Gum - Snow Gum Grassy Forest/Woodland of the New England Tableland Bioregion	<ul style="list-style-type: none"> PCT described as a grassy forest/woodland, and part of the Grassy Woodland vegetation formation. This is not representative of the PCT within the subject land which supports a shrubbier and fernier understorey. Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land
1198	Snow Gum - New England Peppermint grassy open forest of the New England Tableland Bioregion	<ul style="list-style-type: none"> PCT noted as occurring mainly on sandstone geologies not representative of the subject land. Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land

PCT 1604 - Narrow-leaved Ironbark - Grey Box - Spotted Gum shrub - grass woodland of the central and lower Hunter



Vegetation formation: Grassy Woodlands

Vegetation class: Coastal Valley Grassy Woodlands

Conservation status:

- EPBC Act: Not listed
- BC Act: Not listed

Extent within development footprint: 0.02 hectares

No BAM plots: 0

Characteristic species (NSW BioNet):

- Canopy: *Eucalyptus crebra*, *Eucalyptus moluccana*, *Corymbia maculata*.
- Shrub: *Bursaria spinosa*, *Olearia elliptica*.
- Ground: *Eremophila debilis*, *Cymbopogon refractus*, *Aristida ramosa*, *Aristida vagans*, *Microlaena stipoides*, *Austrodanthonia fulva*, *Cheilanthes sieberi*, *Lomandra multiflora*, *Brunoniella australis*.

Justification of PCT: PCT was found to occur on the road verge along the transport haul route as scattered occurrences of *Corymbia maculata* *Eucalyptus moluccana* and *Eucalyptus crebra*, over a generally disturbed understorey in low condition.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
623	Narrow-leaved Ironbark +/- Grey Box grassy woodland of the upper Hunter Valley, mainly Sydney Basin Bioregion	<ul style="list-style-type: none"> PCT noted as occurring on hillcrests and upper hillslopes, not representative of the PCTs occurrence within the subject land. Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land
1691	Narrow-leaved Ironbark - Grey Box grassy woodland of the central and upper Hunter	<ul style="list-style-type: none"> PCT was found to be grassier and less shrubby than areas of PCT 1604 within the subject land

PCT 1691 - Narrow-leaved Ironbark - Grey Box grassy woodland of the central and upper Hunter



Vegetation formation: Grassy Woodlands

Vegetation class: Coastal Valley Grassy Woodlands

Conservation status:

- EPBC Act: Central Hunter Valley eucalypt forest and woodland
- BC Act: Central Hunter Grey Box—Ironbark Woodland in the New South Wales North Coast and Sydney Basin Bioregions

Extent within development footprint: 0.04 hectares

No. Bam plots: 0

Characteristic species (NSW BioNet:

- Canopy: *Eucalyptus crebra*, *Eucalyptus moluccana*, *Brachychiton populneus*.
- Shrub: *Notelaea macrocarpa*.
- Ground: *Eremophila debilis*, *Aristida ramosa*, *Cymbopogon refractus*, *Chloris ventricosa*, *Calotis lappulacea*, *Dichondra repens*, *Eragrostis leptostachya*, *Microlaena stipoides*, *Austrostipa verticillata*.

Justification of PCT: PCT was found to occur on the road verge along the transport haul route, with impacts only to occur to a highly degraded edge of the PCT, too disturbed to meet the listing requirements as a BC Act or EPB Act TEC. The PCT occurs as scattered occurrences of *Eucalyptus crebra* and *Eucalyptus tereticornis*, over a grassy understorey in low to moderate condition to the north and away from the road verge.

Similar PCTs		Justification of best fit (BioNet PCT data)
PCT ID	PCT name	
623	Narrow-leaved Ironbark +/- Grey Box grassy woodland of the upper Hunter Valley, mainly Sydney Basin Bioregion	<ul style="list-style-type: none"> PCT noted as occurring on hillcrests and upper hillslopes, not representative of the PCTs occurrence within the subject land. Characteristic canopy and shrub species assemblages are representative of the PCT within the subject land
1604	Narrow-leaved Ironbark - Grey Box - Spotted Gum shrub - grass woodland of the central and lower Hunter	<ul style="list-style-type: none"> PCT was found to be shrubbier and less grassy than areas of PCT 1691 within the subject land

Appendix C Threatened species habitat suitability assessment

Table 103 Consideration of species requiring further assessment

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
Frogs											
<i>Litoria booroolongensis</i>	Booroolong Frog	Species		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Live along permanent streams with some fringing vegetation cover such as ferns, sedges, or grasses.;1 Adults occur on or near cobble banks and other rock structures within stream margins.;2 Shelter under rocks or amongst vegetation near the ground on the stream edge.;3 Sometimes bask in the sun on exposed rocks near flowing water during summer.;4 Breeding occurs in spring and early summer and tadpoles metamorphose in late summer to early autumn.;5 Eggs are laid in	Marginal habitat supported by a number of minor waterbodies within the subject land. Low quality potential habitat present where transmission line corridor crosses Wombramurra Creek	Moderate

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									submerged rock crevices and tadpoles grow in slow-flowing connected or isolated pools.;6		
<i>Litoria daviesae</i>	Davies' Tree Frog	Species		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes	Other; Streams or swamps or within 250 m of waterbodies	Davies' Tree Frog occurs in permanent, slow-flowing small streams above 400 m elevation, mostly in the headwaters of eastern-flowing streams (although it does occur in the headwaters of the western-flowing Peel River).;1 On the tablelands, riparian habitat may be montane heath or dry open forest with fringing tea tree, tussocks and ferns. Escarpment habitat is typically rainforest and wet sclerophyll with a rainforest understorey.;2 Breeding occurs in spring and early summer. Daytime calling is common during the breeding season. At night,	Marginal habitat supported by a number of minor waterbodies within the subject land. Habitats degraded on transmission line corridor	Moderate

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									males can be found calling from perched positions on trees and shrubs 0.5 - 1.5 m above streams.;3 The species has rarely been observed away from the riparian zone, implying a reliance on that zone for breeding and foraging. However, nothing is known of habitat use outside the breeding season.;4		
<i>Litoria subglandulosa</i>	Glandular Frog	Species		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes		Glandular Frogs may be found along streams in rainforest, moist and dry eucalypt forest or in subalpine swamps.;1 Breeding occurs in summer, and possibly in spring.;2	Marginal habitat supported by a number of minor waterbodies within the subject land. Habitats degraded on transmission line corridor. Species records associated with large areas on intact vegetation	Low

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
										to the east of the project site, with no records within 100kms of the project site.	
<i>Mixophyes balbus</i>	Stuttering Frog	Species		Yes	Variegated - 31-70% habitat retained	5 - 24 ha	Yes		Found in rainforest and wet, tall open forest in the foothills and escarpment on the eastern side of the Great Dividing Range.;1 Outside the breeding season adults live in deep leaf litter and thick understorey vegetation on the forest floor.;2 Feed on insects and smaller frogs.;3 Breed in streams during summer after heavy rain.;4 Eggs are laid on rock shelves or shallow riffles in small, flowing streams.;5 As the tadpoles grow they move to deep permanent pools and take approximately 12 months to metamorphose.;6	Marginal habitat supported by a number of minor waterbodies within the subject land. Habitats degraded on transmission line corridor. Species records associated with large areas on intact vegetation to the east of the project site, with no records within 100kms of the project site.	Low

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
<i>Philoria sphagnicolus</i>	Sphagnum Frog	Species	No	Yes	Intact (> 70% natural habitat retained)	5 - <25 ha	Yes	No	The habitat of the Sphagnum Frog is characterised by high moisture levels. They are typically found in high rainfall areas at high elevation in Sphagnum Moss beds or seepages on steep slopes. Habitat often occurs in rainforest (including Antarctic Beech forest) and wet sclerophyll forest. They can also occur at lower elevation (to about 250 m) in wet coastal foothills.;1 Sphagnum Frogs burrow in loose, moist soil or moss, under leaf litter often in soaks or seepages, or may use cracks and cavities behind and beside large or small waterfalls where the environment remains saturated with moisture.;2 They eat ants and other insects.;3 The	Typically found in high rainfall areas at high elevation in Sphagnum Moss beds or seepages on steep slopes. This habitat is not present within the development footprint. Records >50kms to the north-east of the subject land	Nil

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									Sphagnum Frog breeds in spring - summer and calls diurnally. Eggs are laid in moist locations such as rock crevices, under logs or in burrows in Shpagnum Moss. Larvae emerge from the nest after about one month.;4		
Birds											
<i>Anthochaera phrygia</i>	Regent Honeyeater	Species/ Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes	Other; As per mapped areas	The Regent Honeyeater is a flagship threatened woodland bird whose conservation will benefit a large suite of other threatened and declining woodland fauna. The species inhabits dry open forest and woodland, particularly Box-Ironbark woodland, and riparian forests of River Sheoak. Regent Honeyeaters inhabit woodlands that support a significantly high abundance and species	Potential forage habitat supported across the development footprint. Project site does not occur within mapped Important Areas for the species	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									richness of bird species. These woodlands have significantly large numbers of mature trees, high canopy cover and abundance of mistletoes.;1 Every few years non-breeding flocks are seen foraging in flowering coastal Swamp Mahogany and Spotted Gum forests, particularly on the central coast and occasionally on the upper north coast. Birds are occasionally seen on the south coast.;2 In the last 10 years Regent Honeyeaters have been recorded in urban areas around Albury where woodlands tree species such as Mugga Ironbark and Yellow Box were planted 20 years ago.;3 The Regent Honeyeater is a generalist		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>forager, although it feeds mainly on the nectar from a relatively small number of eucalypts that produce high volumes of nectar. Key eucalypt species include Mugga Ironbark, Yellow Box, White Box and Swamp Mahogany. Other tree species may be regionally important. For example, the Lower Hunter Spotted Gum forests have recently been demonstrated to support regular breeding events. Flowering of associated species such as Thin-leaved Stringybark</p> <p><i>Eucalyptus eugenioides</i> and other Stringybark species, and Broad-leaved Ironbark <i>E. fibrosa</i> can also contribute important nectar flows at times. Nectar and fruit from the mistletoes <i>Amyema</i></p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>miquelii, A. pendula</p> <p>and A. cambagei are also utilised. When nectar is scarce lerp and honeydew can comprise a large proportion of the diet. Insects make up about 15% of the total diet and are important components of the diet of nestlings.</p> <p>4 Colour-banding of Regent Honeyeater has shown that the species can undertake large-scale nomadic movements in the order of hundreds of kilometers. However, the exact nature of these movements is still poorly understood. It is likely that movements are dependent on spatial and temporal flowering and other resource patterns. To successfully manage the recovery of this species a</p>		

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									<p>full understanding of the habitats used in the non-breeding season is critical.;⁵ There are three known key breeding areas, two of them in NSW - Capertee Valley and Bundarra-Barraba regions. The species breeds between July and January in Box-Ironbark and other temperate woodlands and riparian gallery forest dominated by River Sheoak. Regent Honeyeaters usually nest in horizontal branches or forks in tall mature eucalypts and Sheoaks. Also nest in mistletoe haustoria.;⁶ An open cup-shaped nest is constructed of bark, grass, twigs and wool by the female. Two or three eggs are laid and incubated by the female for 14 days. Nestlings are</p>		

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									brooded and fed by both parents at an average rate of 23 times per hour and fledge after 16 days. Fledglings fed by both parents 29 times per hour.;7		
<i>Artamus cyanopterus cyanopterus</i>	Dusky Woodswallow	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Primarily inhabit dry, open eucalypt forests and woodlands, including mallee associations, with an open or sparse understorey of eucalypt saplings, acacias and other shrubs, and ground-cover of grasses or sedges and fallen woody debris. It has also been recorded in shrublands, heathlands and very occasionally in moist forest or rainforest. Also found in farmland, usually at the edges of forest or woodland.;1 Primarily eats invertebrates, mainly	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>insects, which are captured whilst hovering or sallying above the canopy or over water. Also, frequently hovers, sallies and pounces under the canopy, primarily over leaf litter and dead timber. Also occasionally take nectar, fruit and seed.</p> <p>;2 Depending on location and local climatic conditions (primarily temperature and rainfall), the dusky woodswallow can be resident year-round or migratory. In NSW, after breeding, birds migrate to the north of the state and to southeastern Queensland, while Tasmanian birds migrate to southeastern NSW after breeding. Migrants generally depart between March and May, heading south to breed again in</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>spring. There is some evidence of site fidelity for breeding. Although dusky wood swallows generally breed as solitary pairs or occasionally in small flocks, large flocks may form around abundant food sources in winter. Large flocks may also form before migration, which is often undertaken with other species. ;3 Nest is an open, cup-shape, made of twigs, grass, fibrous rootlets and occasionally casuarina needles, and may be lined with grass, rootlets or infrequently horsehair, occasionally unlined. Nest sites vary greatly, but generally occur in shrubs or low trees, living or dead, horizontal or upright forks in branches, spouts, hollow stumps or logs, behind loose bark or</p>		

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									in a hollow in the top of a wooden fence post. Nest sites may be exposed or well concealed by foliage.;4		
<i>Burhinus grallarius</i>	Bush Stone-curlew	Species		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	No	Fallen/standing dead timber including logs; Null	Inhabits open forests and woodlands with a sparse grassy ground layer and fallen timber.;1 Largely nocturnal, being especially active on moonlit nights.;2 Feed on insects and small vertebrates, such as frogs, lizards and snakes.;3 Nest on the ground in a scrape or small bare patch.;4 Two eggs are laid in spring and early summer.;5	Species occurs at altitudes much lower than the development footprint with the highest elevation record of the species within over 120kms of the project site at an altitude of 500 metres (approx.). The lowest point of the project site occurs along the transmission line at an altitude of 750 metres (approx.) and as such the development	Negligible

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
										<p>footprint does not support habitat for the species. Two records of the species occur at an elevation of approximately 1000 metres, one hears Armidale over 120kms from the project site, and the other in Washpool NP, over 270kms from the project site. When these records are compared to the remainder of the 1350 species' records in BioNet, these occurrences are considered to be a vagrant.</p>	

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<i>Callocephalon fimbriatum</i>	Gang-gang Cockatoo	Species/ Ecosystem		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	No	Hollow bearing trees; Eucalypt tree species with hollows greater than 9 cm diameter	In spring and summer, generally found in tall mountain forests and woodlands, particularly in heavily timbered and mature wet sclerophyll forests. ;1 In autumn and winter, the species often moves to lower altitudes in drier more open eucalypt forests and woodlands, particularly box-gum and box-ironbark assemblages, or in dry forest in coastal areas and often found in urban areas.;2 May also occur in sub-alpine Snow Gum (Eucalyptus pauciflora) woodland and occasionally in temperate rainforests.;3 Favours old growth forest and woodland attributes for nesting and roosting. Nests are located in hollows that are 10 cm in diameter or	Of the 16,000 records of the species in ebird (and >600 in BioNet), none occur north of Muswellbrook NSW, except occasional records along coast just south of Coffs Harbour. As such the development footprint does not support habitat for the species.	Negligible

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									larger and at least 9 m above the ground in eucalypts.;4		
<i>Calyptorhynchus lathamii</i>	Glossy Black-Cockatoo	Species/Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes	Forage: Presence of Allocasuarina and casuarina species Breeding: Hollow bearing trees; Living or dead tree with hollows greater than 15cm diameter and greater than 5m above ground.	Inhabits open forest and woodlands of the coast and the Great Dividing Range where stands of sheoak occur. Black Sheoak (Allocasuarina littoralis) and Forest Sheoak (A. torulosa) are important foods.;1 Inland populations feed on a wide range of sheoaks, including Drooping Sheoak, Allocasuarina diminuta, and A. gymnathera. Belah is also utilised and may be a critical food source for some populations.;2 In the Riverina, birds are associated with hills and rocky rises supporting Drooping Sheoak, but also recorded in open	Marginal potential forage habitat supported across the development footprint, very few Casuarina spp or Allocasuarina spp. have been recorded during floristic surveys of fauna habitat assessments. Breeding habitat potentially present in the form of hollow trees.	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									woodlands dominated by Belah (<i>Casuarina cristata</i>);3 Feeds almost exclusively on the seeds of several species of she-oak (<i>Casuarina</i> and <i>Allocasuarina</i> species), shredding the cones with the massive bill.;4 Dependent on large hollow-bearing eucalypts for nest sites. A single egg is laid between March and May.;5		
<i>Chthonicola sagittata</i>	Speckled Warbler	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		The Speckled Warbler lives in a wide range of <i>Eucalyptus</i> dominated communities that have a grassy understorey, often on rocky ridges or in gullies.;1 Typical habitat would include scattered native tussock grasses, a sparse shrub layer, some	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									eucalypt regrowth and an open canopy.;2 Large, relatively undisturbed remnants are required for the species to persist in an area.;3 The diet consists of seeds and insects, with most foraging taking place on the ground around tussocks and under bushes and trees.;4 Pairs are sedentary and occupy a breeding territory of about ten hectares, with a slightly larger home-range when not breeding.;5 The rounded, domed, roughly built nest of dry grass and strips of bark is located in a slight hollow in the ground or the base of a low dense plant, often among fallen branches and other litter. A side entrance allows the bird to walk directly inside.;6 A clutch of 3-4 eggs is laid, between		

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									August and January, and both parents feed the nestlings. The eggs are a glossy red-brown, giving rise to the unusual folk names 'Blood Tit' and 'Chocolatebird'; ⁷ Some cooperative breeding occurs. The species may act as host to the Black-eared Cuckoo; ⁸ Speckled Warblers often join mixed species feeding flocks in winter, with other species such as Yellow-rumped, Buff-rumped, Brown and Striated Thornbills; ⁹		
<i>Climacteris picumnus victoriae</i>	Brown Treecreeper (eastern subspecies)	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Found in eucalypt woodlands (including Box-Gum Woodland) and dry open forest of the inland slopes and plains inland of the Great Dividing Range; mainly inhabits woodlands dominated by stringybarks or other rough-barked	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>eucalypts, usually with an open grassy understorey, sometimes with one or more shrub species; also found in mallee and River Red Gum (<i>Eucalyptus camaldulensis</i>)</p> <p>Forest bordering wetlands with an open understorey of acacias, saltbush, lignum, cumbungi and grasses; usually not found in woodlands with a dense shrub layer; fallen timber is an important habitat component for foraging; also recorded, though less commonly, in similar woodland habitats on the coastal ranges and plains.;¹ Sedentary, considered to be resident in many locations throughout its range; present in all seasons or year-round at many sites; territorial year-round,</p>		

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									<p>though some birds may disperse locally after breeding.;2 Gregarious and usually observed in pairs or small groups of 8 to 12 birds; terrestrial and arboreal in about equal proportions; active, noisy and conspicuous while foraging on trunks and branches of trees and amongst fallen timber; spend much more time foraging on the ground and fallen logs than other treecreepers.</p> <p>;3 When foraging in trees and on the ground, they peck and probe for insects, mostly ants, amongst the litter, tussocks and fallen timber, and along trunks and lateral branches; up to 80% of the diet is comprised of ants; other invertebrates (including spiders, insects larvae,</p>		

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									<p>moths, beetles, flies, hemipteran bugs, cockroaches, termites and lacewings) make up the remaining percentage; nectar from Mugga Ironbark (<i>Eucalyptus sideroxylon</i>) and paperbarks, and sap from an unidentified eucalypt are also eaten, along with lizards and food scraps; young birds are fed ants, insect larvae, moths, crane flies, spiders and butterfly and moth larvae;⁴ Hollows in standing dead or live trees and tree stumps are essential for nesting.</p> <p>⁵ The species breeds in pairs or co-operatively in territories which range in size from 1.1 to 10.7 ha (mean = 4.4 ha). Each group is composed of a breeding pair with retained</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									male offspring and, rarely, retained female offspring. Often in pairs or cooperatively breeding groups of two to five birds. ;6		
<i>Daphoenositta chrysoptera</i>	Varied Sittella	Ecosystem		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes		Inhabits eucalypt forests and woodlands, especially those containing rough-barked species and mature smooth-barked gums with dead branches, mallee and <i>Acacia</i> woodland.;1 Feeds on arthropods gleaned from crevices in rough or decorticated bark, dead branches, standing dead trees and small branches and twigs in the tree canopy.;2 Builds a cup-shaped nest of plant fibres and cobwebs in an upright tree fork high in the living tree canopy, and often re-uses the same fork or tree	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									in successive years.;3 Generation length is estimated to be 5 years.;4		
<i>Glossopsitta pusilla</i>	Little Lorikeet	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Forages primarily in the canopy of open <i>Eucalyptus</i> forest and woodland, yet also finds food in <i>Angophora</i> , <i>Melaleuca</i> and other tree species. Riparian habitats are particularly used, due to higher soil fertility and hence greater productivity.;1 Isolated flowering trees in open country, e.g. paddocks, roadside remnants and urban trees also help sustain viable populations of the species.;2 Feeds mostly on nectar and pollen, occasionally on native fruits such as mistletoe, and only rarely in	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>orchards;3 Gregarious, travelling and feeding in small flocks (<10), though often with other lorikeets. Flocks numbering hundreds are still occasionally observed and may have been the norm in past centuries.;4 Roosts in treetops, often distant from feeding areas.;5 Nests in proximity to feeding areas if possible, most typically selecting hollows in the limb or trunk of smooth-barked Eucalypts. Entrance is small (3 cm) and usually high above the ground (2–15 m). These nest sites are often used repeatedly for decades, suggesting that preferred sites are limited. Riparian trees often chosen, including species like</p> <p><i>Allocasuarina</i>;</p> <p>6 Nesting season extends</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									from May to September. In years when flowering is prolific, Little Lorikeet pairs can breed twice, producing 3-4 young per attempt. However, the survival rate of fledglings is unknown.;7		
<i>Grantiella picta</i>	Painted Honeyeater	Ecosystem		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes		Inhabits Boree/ Weeping Myall (Acacia pendula), Brigalow (A. harpophylla) and Box-Gum Woodlands and Box-Ironbark Forests.;1 A specialist feeder on the fruits of mistletoes growing on woodland eucalypts and acacias. Prefers mistletoes of the genus Amyema;2 Insects and nectar from mistletoe or eucalypts are occasionally eaten.;3 Nest from spring to autumn in a small, delicate nest hanging within the outer canopy of	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									drooping eucalypts, she-oak, paperbark or mistletoe branches;4		
<i>Haliaeetus leucogaster</i>	White-bellied Sea-Eagle	Species/ Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	No	Waterbodies; Within 1km of a rivers, lakes, large dams or creeks, wetlands and coastlines; Living or dead mature trees within suitable vegetation within 1km of a rivers, lakes, large dams or creeks, wetlands and coastlines	Habitats are characterised by the presence of large areas of open water including larger rivers, swamps, lakes, and the sea.;1 Occurs at sites near the sea or sea-shore, such as around bays and inlets, beaches, reefs, lagoons, estuaries and mangroves; and at, or in the vicinity of freshwater swamps, lakes, reservoirs, billabongs and saltmarsh. ;2 Terrestrial habitats include coastal dunes, tidal flats, grassland, heathland, woodland, and forest (including rainforest). ;3 Breeding habitat consists of mature tall open forest, open forest, tall woodland, and swamp sclerophyll forest close to	Project site does not occur within 1km of a rivers, lakes, large dams or creeks, wetlands and coastlines. Where Peel River occurs within 1km of the development footprint it is a minor watercourse.	Negligible

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>foraging habitat.</p> <p>Nest trees are typically large emergent eucalypts and often have emergent dead branches or large dead trees nearby which are used as 'guard roosts. Nests are large structures built from sticks and lined with leaves or grass.</p> <p>;4 Feed mainly on fish and freshwater turtles, but also waterbirds, reptiles, mammals and carrion.;5 Hunts its prey from a perch or whilst in flight (by circling slowly, or by sailing along 10–20 m above the shore). Prey is usually carried to a feeding platform or (if small) consumed in flight, but some items are eaten on the ground.;6 May be solitary, or live in pairs or small family groups consisting of a pair of</p>		

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									adults and dependent young. ;7 Typically lays two eggs between June and September with young birds remaining in the nest for 65-70 days.;8		
<i>Hamirostra melanosternon</i>	Black-breasted Buzzard	Species/ Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes	Waterbodies; Land within 40 m of riparian woodland on inland watercourses /waterholes containing dead or dying eucalypts	Lives in a range of inland habitats, especially along timbered watercourses which is the preferred breeding habitat.;1 Also hunts over grasslands and sparsely timbered woodlands.;2 Not a powerful hunter, despite its size, mostly taking reptiles, small mammals, birds, including nestlings, and carrion.;3 Also specialises in feeding on large eggs, including those of emus, which it cracks on a rock.;4 Breeds from August to October near water in a tall tree. The	Riparian habitats are degraded within the development footprint.	Negligible

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									stick nest is large and flat and lined with green leaves. Normally two eggs are laid.;5		
<i>Hieraaetus morphnoides</i>	Little Eagle	Species/ Ecosystem		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes	Other; Nest trees - live (occasionally dead) large old trees within vegetation.	Occupies open eucalypt forest, woodland or open woodland. Sheoak or <i>Acacia</i> woodlands and riparian woodlands of interior NSW are also used.;1 Nests in tall living trees within a remnant patch, where pairs build a large stick nest in winter.;2 Lays two or three eggs during spring, and young fledge in early summer.;3 Preys on birds, reptiles and mammals, occasionally adding large insects and carrion.;4	Potential forage and breeding habitat supported across the development footprint	Moderate
<i>Lathamus discolor</i>	Swift Parrot	Species/ Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes	Other; As per mapped areas	Migrates to the Australian south-east mainland between February and October.;1 On the mainland they occur in	Potential forage habitat supported across the development footprint. Project	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>areas where eucalypts are flowering profusely or where there are abundant lerp (from sap-sucking bugs) infestations.;² Favoured feed trees include winter flowering species such as Swamp Mahogany Eucalyptus robusta, Spotted Gum Corymbia maculata, Red Bloodwood C. gummifera, Forest Red Gum E. tereticornis, Mugga Ironbark E. sideroxylon, and White Box E. albens;³ Commonly used lerp infested trees include Inland Grey Box E. microcarpa, Grey Box E. moluccana, Blackbutt E.</p>	site does not occur within mapped Important Areas for the species	

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									<p>pilularis, and Yellow Box E. melliodora;4 Return to some foraging sites on a cyclic basis depending on food availability.;5 Following winter they return to Tasmania where they breed from September to January, nesting in old trees with hollows and feeding in forests dominated by Tasmanian Blue Gum Eucalyptus globulus;6 </p>		
<i>Lophoictinia isura</i>	Square-tailed Kite	Species/ Ecosystem		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes	Other; Nest trees	Found in a variety of timbered habitats including dry woodlands and open forests. Shows a particular preference for timbered watercourses.;1 In arid north-western NSW, has been observed in stony country with a ground cover of chenopods and	Potential forage and breeding habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									grasses, open acacia scrub and patches of low open eucalypt woodland.;2 Is a specialist hunter of passerines, especially honeyeaters, and most particularly nestlings, and insects in the tree canopy, picking most prey items from the outer foliage.;3 Appears to occupy large hunting ranges of more than 100km2.;4 Breeding is from July to February, with nest sites generally located along or near watercourses, in a fork or on large horizontal limbs.;5		
<i>Melanodryas cucullata</i>	Hooded Robin (south-eastern form)	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Prefers lightly wooded country, usually open eucalypt woodland, acacia scrub and mallee, often in or near clearings or open areas.;1 Requires structurally diverse habitats	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									featuring mature eucalypts, saplings, some small shrubs and a ground layer of moderately tall native grasses.;2 Often perches on low dead stumps and fallen timber or on low-hanging branches, using a perch-and-pounce method of hunting insect prey.;3 Territories range from around 10 ha during the breeding season, to 30 ha in the non-breeding season.;4 May breed any time between July and November, often rearing several broods.;5 The nest is a small, neat cup of bark and grasses bound with webs, in a tree fork or crevice, from less than 1 m to 5 m above the ground.;6 The nest is defended by both sexes with displays of injury-feigning, tumbling across		

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									the ground.;7 A clutch of two to three is laid and incubated for fourteen days by the female. Two females often cooperate in brooding.;8		
<i>Ninox connivens</i>	Barking Owl	Species/ Ecosystem		Yes	Fragmented - 11-30 % habitat retained	25 - 100 ha	Yes	Hollow bearing trees; Living or dead trees with hollows greater than 20 cm diameter and greater than 4m above the ground.	Inhabits woodland and open forest, including fragmented remnants and partly cleared farmland. It is flexible in its habitat use, and hunting can extend in to closed forest and more open areas. Sometimes able to successfully breed along timbered watercourses in heavily cleared habitats (e.g. western NSW) due to the higher density of prey found on these fertile riparian soils.;1 Roost in shaded portions of tree canopies, including tall midstorey trees with dense foliage such as	Potential forage and breeding habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>Acacia and Casuarina species. During nesting season, the male perches in a nearby tree overlooking the hollow entrance.;2 Preferentially hunts small arboreal mammals such as Squirrel Gliders and Common Ringtail Possums, but when loss of tree hollows decreases these prey populations the owl becomes more reliant on birds, invertebrates and terrestrial mammals such as rodents and rabbits. Can catch bats and moths on the wing, but typically hunts by sallying from a tall perch.;3 Requires very large permanent territories in most habitats due to sparse prey densities. Monogamous pairs hunt over as much as 6000</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>hectares, with 2000 hectares being more typical in NSW habitats.;4 Two or three eggs are laid in hollows of large, old trees. Living eucalypts are preferred though dead trees are also used. Nest sites are used repeatedly over years by a pair, but they may switch sites if disturbed by predators (e.g. goannas).;5 Nesting occurs during mid-winter and spring, being variable between pairs and among years. As a rule of thumb, laying occurs during August and fledging in November. The female incubates for 5 weeks, roosts outside the hollow when chicks are 4 weeks old, then fledging occurs 2-3 weeks later. Young are dependent on their parents for several months.;6 Territorial pairs</p>		

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									respond strongly to recordings of Barking Owl calls from up to 6 km away, though humans rarely hear this response farther than 1.5 km. Because disturbance reduces the pair's foraging time and can pull the female off her eggs even on cold nights, recordings should not be broadcast unnecessarily nor during the nesting season.;7		
<i>Ninox strenua</i>	Powerful Owl	Species/ Ecosystem		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes	Hollow bearing trees; Living or dead trees with hollow greater than 20cm diameter	The Powerful Owl inhabits a range of vegetation types, from woodland and open sclerophyll forest to tall open wet forest and rainforest.;1 The Powerful Owl requires large tracts of forest or woodland habitat but can occur in fragmented landscapes as well. The species breeds and hunts in open or	Potential forage and breeding habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>closed sclerophyll forest or woodlands and occasionally hunts in open habitats. It roosts by day in dense vegetation comprising species such as Turpentine <i>Syncarpia glomulifera</i>, Black She-oak <i>Allocasuarina littoralis</i>, Blackwood <i>Acacia melanoxylon</i>, Rough-barked Apple <i>Angophora floribunda</i>, Cherry Ballart <i>Exocarpus cupressiformis</i> and a number of eucalypt species. ;2 The main prey items are medium-sized arboreal marsupials, particularly the Greater Glider, Common Ringtail Possum and Sugar Glider. There may be marked regional differences in the</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>prey taken by Powerful Owls. For example, in southern NSW, Ringtail Possum make up the bulk of prey in the lowland or coastal habitat. At higher elevations, such as the tableland forests, the Greater Glider may constitute almost all of the prey for a pair of Powerful Owls. Flying foxes are important prey in some areas; birds comprise about 10-50% of the diet depending on the availability of preferred mammals. As most prey species require hollows and a shrub layer, these are important habitat components for the owl.</p> <p>;3 Pairs of Powerful Owls demonstrate high fidelity to a large territory, the size of which varies with habitat quality and thus prey</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>densities. In good habitats a mere 400 can support a pair; where hollow trees and prey have been depleted the owls need up to 4000 ha.;4 Powerful Owls nest in large tree hollows (at least 0.5 m deep), in large eucalypts (diameter at breast height of 80-240 cm) that are at least 150 years old. While the female and young are in the nest hollow the male Powerful Owl roosts nearby (10-200 m) guarding them, often choosing a dense "grove" of trees that provide concealment from other birds that harass him.;5 Powerful Owls are monogamous and mate for life. Nesting occurs from late autumn to mid-winter but is slightly earlier in north-eastern NSW (late</p>		

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									summer - mid autumn). Clutches consist of two dull white eggs and incubation lasts approximately 38 days.;6		
<i>Pachycephala olivacea</i>	Olive Whistler	Ecosystem		Yes	Variegated - 31-70% habitat retained	25 - 100 ha	Yes		Mostly inhabit wet forests above about 500m. During the winter months they may move to lower altitudes.;1 Forage in trees and shrubs and on the ground, feeding on berries and insects.;2 Make nests of twigs and grass in low forks of shrubs.;3 Lay two or three eggs between September and January;4	Potential forage habitat supported across the development footprint	Moderate
<i>Petroica boodang</i>	Scarlet Robin	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		The Scarlet Robin lives in dry eucalypt forests and woodlands. The understorey is usually open and grassy with few scattered shrubs.;1 This species lives in both mature and regrowth vegetation. It occasionally	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									occurs in mallee or wet forest communities, or in wetlands and tea-tree swamps.;2 Scarlet Robin habitat usually contains abundant logs and fallen timber: these are important components of its habitat.;3 The Scarlet Robin breeds on ridges, hills and foothills of the western slopes, the Great Dividing Range and eastern coastal regions; this species is occasionally found up to 1000 metres in altitude.;4 The Scarlet Robin is primarily a resident in forests and woodlands, but some adults and young birds disperse to more open habitats after breeding.;5 In autumn and winter many Scarlet Robins live in open grassy woodlands, and grasslands		

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									<p>or grazed paddocks with scattered trees.;⁶ The Scarlet Robin is a quiet and unobtrusive species which is often quite tame and easily approached.;⁷ Birds forage from low perches, fence-posts or on the ground, from where they pounce on small insects and other invertebrates which are taken from the ground, or off tree trunks and logs; they sometimes forage in the shrub or canopy layer.;⁸ Scarlet Robin pairs defend a breeding territory and mainly breed between the months of July and January; they may raise two or three broods in each season.;⁹ This species' nest is an open cup made of plant fibers and cobwebs and is built in the fork of tree usually more</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									than 2 metres above the ground; nests are often found in a dead branch in a live tree, or in a dead tree or shrub.; ¹⁰ Eggs are pale greenish-, bluish- or brownish-white, spotted with brown; clutch size ranges from one to four.; ¹¹ Birds usually occur singly or in pairs, occasionally in small family parties; pairs stay together year-round.; ¹² In autumn and winter, the Scarlet Robin joins mixed flocks of other small insectivorous birds which forage through dry forests and woodlands.; ¹³		
<i>Petroica phoenicea</i>	Flame Robin	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Breeds in upland tall moist eucalypt forests and woodlands, often on ridges and slopes.; ¹ Prefers clearings or areas with open understoreys.; ² The	Potential forage habitat supported across the development footprint	Moderate

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									ground layer of the breeding habitat is dominated by native grasses and the shrub layer may be either sparse or dense.;3 Occasionally occurs in temperate rainforest, and also in herbfields, heathlands, shrublands and sedgelands at high altitudes.;4 In winter, birds migrate to drier more open habitats in the lowlands (i.e. valleys below the ranges, and to the western slopes and plains).;5 Often occurs in recently burnt areas; however, habitat becomes unsuitable as vegetation closes up following regeneration.;6 In winter lives in dry forests, open woodlands and in pastures and native grasslands, with or without scattered trees.;7 In winter,		

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									occasionally seen in heathland or other shrublands in coastal areas.;8 Birds forage from low perches, from which they sally or pounce onto small invertebrates which they take from the ground or off tree trunks, logs and other coarse woody debris.;9 Flying insects are often taken in the air and sometimes gleans for invertebrates from foliage and bark.;10 In their autumn and winter habitats, birds often sally from fence-posts or thistles and other prominent perches in open habitats.;11 Occur singly, in pairs, or in flocks of up to 40 birds or more; in the non-breeding season they will join up with other insectivorous birds in mixed feeding		

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									flocks.;12 Breeds in spring to late summer.;13 Nests are often near the ground and are built in sheltered sites, such as shallow cavities in trees, stumps or banks.;14 Builds an open cup nest made of plant materials and spider webs.;15 Eggs are oval in shape and are pale bluish- or greenish-white and marked with brownish blotches; clutch size is three or four eggs.;16		
<i>Stagonopleura guttata</i>	Diamond Firetail	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Found in grassy eucalypt woodlands, including Box-Gum Woodlands and Snow Gum Eucalyptus pauciflora Woodlands.;1 Also occurs in open forest, mallee, Natural Temperate Grassland, and in secondary grassland derived from other	Potential forage habitat supported across the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									communities.;2 Often found in riparian areas (rivers and creeks), and sometimes in lightly wooded farmland.;3 Feeds exclusively on the ground, on ripe and partly-ripe grass and herb seeds and green leaves, and on insects (especially in the breeding season).;4 Usually encountered in flocks of between 5 to 40 birds, occasionally more.;5 Groups separate into small colonies to breed, between August and January.;6 Nests are globular structures built either in the shrubby understorey, or higher up, especially under hawk's or raven's nests.;7 Birds roost in dense shrubs or in smaller nests built especially for		

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									roosting.;8 Appears to be sedentary, though some populations move locally, especially those in the south.;9 Has been recorded in some towns and near farm houses.;10		
<i>Tyto novaehollandiae</i>	Masked Owl	Species/ Ecosystem		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes	Hollow bearing trees; Living or dead trees with hollows greater than 20cm diameter.	Lives in dry eucalypt forests and woodlands from sea level to 1100 m.;1 A forest owl, but often hunts along the edges of forests, including roadsides.;2 The typical diet consists of tree-dwelling and ground mammals, especially rats.;3 Pairs have a large home-range of 500 to 1000 hectares.;4 Roosts and breeds in moist eucalypt forested gullies, using large tree hollows or sometimes caves for nesting.;5	Potential forage and breeding habitat supported across the development footprint	Moderate
<i>Tyto tenebricosa</i>	Sooty Owl	Species/ Ecosystem		Yes	Intact - over 70% natural habitat retained	> 100 ha	Yes	Caves; Caves or clifflines/ledg	Occurs in rainforest, including dry rainforest, subtropical and warm	Potential forage and breeding habitat	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
								es Hollow bearing trees; Living or dead trees with hollows greater than 20cm diameter.	temperate rainforest, as well as moist eucalypt forests.;1 Roosts by day in the hollow of a tall forest tree or in heavy vegetation; hunts by night for small ground mammals or tree-dwelling mammals such as the Common Ringtail Possum (Pseudocheirus peregrinus) or Sugar Glider (Petaurus breviceps);.2 Nests in very large tree-hollows.;3	supported across the development footprint	
Mammals											
<i>Aepyprymnus rufescens</i>	Rufous Bettong	Species		Yes	Variegated - 31-70% habitat retained	< 5 ha	Yes		Rufous Bettongs inhabit a variety of forests from tall, moist eucalypt forest to open woodland, with a tussock grass understorey. A dense cover of tall native grasses is the preferred shelter.;1 They sleep during the day in cone-	Marginal potential habitat occurs within the subject land and habitats within the transmission line corridor are degraded	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									shaped nests constructed of grass in a shallow depression at the base of a tussock or fallen log.; ² At night they feed on grasses, herbs, seeds, flowers, roots, tubers, fungi and occasionally insects.; ³		
<i>Cercartetus nanus</i>	Eastern Pygmy-possum	Species		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes		Found in a broad range of habitats from rainforest through sclerophyll (including Box-Ironbark) forest and woodland to heath, but in most areas woodlands and heath appear to be preferred, except in north-eastern NSW where they are most frequently encountered in rainforest.; ¹ Feeds largely on nectar and pollen collected from banksias, eucalypts and bottlebrushes; an important pollinator of heathland plants such as	Potential habitat is present within the development footprint.	Moderate

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									<p>banksias; soft fruits are eaten when flowers are unavailable.;² Also feeds on insects throughout the year; this feed source may be more important in habitats where flowers are less abundant such as wet forests.;³ Shelters in tree hollows, rotten stumps, holes in the ground, abandoned bird-nests, Ringtail Possum (<i>Pseudocheirus peregrinus</i>) dreys or thickets of vegetation, (e.g. grass-tree skirts); nest-building appears to be restricted to breeding females; tree hollows are favoured but spherical nests have been found under the bark of eucalypts and in shredded bark in tree forks.;⁴ Appear to be mainly solitary, each individual using several</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>nests, with males having non-exclusive home-ranges of about 0.68 hectares and females about 0.35 hectares.;5 Young can be born whenever food sources are available, however most births occur between late spring and early autumn.;6 Agile climbers, but can be caught on the ground in traps, pitfalls or postholes; generally nocturnal.;7 Frequently spends time in torpor especially in winter, with body curled, ears folded and internal temperature close to the surroundings.;8 </p>		
<i>Chalinolobus dwyeri</i>	Large-eared Pied Bat	Species		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes		Roosts in caves (near their entrances), crevices in cliffs, old mine workings and in the disused, bottle-shaped mud nests of the Fairy	Habitat occurs within and adjacent to the development footprint	Recorded by survey

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>Martin (<i>Petrochelidon ariel</i>), frequenting low to mid-elevation dry open forest and woodland close to these features. Females have been recorded raising young in maternity roosts (c. 20-40 females) from November through to January in roof domes in sandstone caves and overhangs. They remain loyal to the same cave over many years;¹ Found in well-timbered areas containing gullies;² The relatively short, broad wing combined with the low weight per unit area of wing indicates maneuverable flight. This species probably forages for small, flying insects below the forest canopy;³ Likely to hibernate through the</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									coolest months.;4 It is uncertain whether mating occurs early in winter or in spring.;5		
<i>Dasyurus maculatus</i>	Spotted-tailed Quoll	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Recorded across a range of habitat types, including rainforest, open forest, woodland, coastal heath and inland riparian forest, from the sub-alpine zone to the coastline.;1 Individual animals use hollow-bearing trees, fallen logs, small caves, rock outcrops and rocky-cliff faces as den sites.;2 Mostly nocturnal, although will hunt during the day; spends most of the time on the ground, although also an excellent climber and will hunt possums and gliders in tree hollows and prey on roosting birds.;3 Use communal 'latrine sites',	Habitat occurs within and adjacent to the development footprint	Recorded by survey

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>often on flat rocks among boulder fields, rocky cliff-faces or along rocky stream beds or banks. Such sites may be visited by multiple individuals and can be recognised by the accumulation of the sometimes characteristic 'twisty-shaped' faeces deposited by animals.;⁴ A generalist predator with a preference for medium-sized (500g-5kg) mammals. Consumes a variety of prey, including gliders, possums, small wallabies, rats, birds, bandicoots, rabbits, reptiles and insects. Also eats carrion and takes domestic fowl.;⁵ Females occupy home ranges of 200-500 hectares, while males occupy very large home ranges from 500 to over 4000 hectares. Are known</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									to traverse their home ranges along densely vegetated creeklines.;6 Average litter size is five; both sexes mature at about one year of age. Life expectancy in the wild is about 3-4 years.;7		
<i>Falsistrellus tasmaniensis</i>	Eastern False Pipistrelle	Ecosystem		Yes	Variegated - 31-70% habitat retained	5 - 24 ha	Yes		Prefers moist habitats, with trees taller than 20 m.;1 Generally roosts in eucalypt hollows, but has also been found under loose bark on trees or in buildings.;2 Hunts beetles, moths, weevils and other flying insects above or just below the tree canopy.;3 Hibernates in winter.;4 Females are pregnant in late spring to early summer.;5	Habitat occurs within and adjacent to the development footprint	Recorded by survey
<i>Macropus parma</i>	Parma Wallaby	Species		Yes	Variegated - 31-70% habitat retained	5 - 24 ha	Yes		Preferred habitat is moist eucalypt forest with thick, shrubby understorey, often	Potential habitat occurs in higher condition areas	Moderate

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									with nearby grassy areas, rainforest margins and occasionally drier eucalypt forest.;1 Typically feed at night on grasses and herbs in more open eucalypt forest and the edges of nearby grassy areas.;2 During the day they shelter in dense cover.;3	connected to Ben Halls Gap Nature Reserve. Potential habitats within the transmission line corridor are degraded	
<i>Micronomus norfolkensis</i>	Eastern Coastal Free-tailed Bat	Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Occur in dry sclerophyll forest, woodland, swamp forests and mangrove forests east of the Great Dividing Range.;1 Roost mainly in tree hollows but will also roost under bark or in man-made structures.;2 Usually solitary but also recorded roosting communally, probably insectivorous.;3	Habitat occurs within and adjacent to the development footprint	Recorded by survey
<i>Miniopterus australis</i>	Little Bent-winged Bat	Species/ Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes	Caves; Cave, tunnel, mine, culvert or	Moist eucalypt forest, rainforest, vine thicket, wet and dry sclerophyll forest,	Habitat occurs within and adjacent to the	Recorded by survey

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
								other structure known or suspected to be used for breeding including species records in BioNet with microhabitat code 'IC – in cave'; observation type code 'E nest-roost'; with numbers of individuals >500; or from the scientific literature.	Melaleuca swamps, dense coastal forests and banksia scrub. Generally found in well-timbered areas.; ¹ Little Bentwing-bats roost in caves, tunnels, tree hollows, abandoned mines, stormwater drains, culverts, bridges and sometimes buildings during the day, and at night forage for small insects beneath the canopy of densely vegetated habitats.; ² They often share roosting sites with the Common Bentwing-bat and, in winter, the two species may form mixed clusters.; ³ In NSW the largest maternity colony is in close association with a large maternity colony of Large Bentwing-bats (<i>Miniopterus orianae</i>) and appears to depend on the large colony	development footprint	

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									to provide the high temperatures needed to rear its young.;4 Maternity colonies form in spring and birthing occurs in early summer. Males and juveniles disperse in summer.;5 Only five nursery sites /maternity colonies are known in Australia.;6		
<i>Miniopterus orianae oceanensis</i>	Large Bent-winged Bat	Species/ Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes	Caves; Cave, tunnel, mine, culvert or other structure known or suspected to be used for breeding including species records with microhabitat code "IC - in cave;"	Caves are the primary roosting habitat, but also use derelict mines, storm-water tunnels, buildings and other man-made structures.;1 Form discrete populations centered on a maternity cave that is used annually in spring and summer for the birth and rearing of young.;2 Maternity caves have very specific temperature and humidity regimes.;3 At other times	Habitat occurs within and adjacent to the development footprint	Recorded by survey

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
								observation type code "E nest-roost;" with numbers of individuals >500	of the year, populations disperse within about 300 km range of maternity caves.;4 Cold caves are used for hibernation in southern Australia.;5 Breeding or roosting colonies can number from 100 to 150,000 individuals.;6 Hunt in forested areas, catching moths and other flying insects above the tree tops.;7		
<i>Myotis macropus</i>	Southern Myotis	Species		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Generally roost in groups of 10 - 15 close to water in caves, mine shafts, hollow-bearing trees, storm water channels, buildings, under bridges and in dense foliage.;1 Forage over streams and pools catching insects and small fish by raking their feet across the water surface.;2 In NSW	Habitat occurs within and adjacent to the development footprint	Recorded by survey

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									females have one young each year usually in November or December.;3		
<i>Nyctophilus corbeni</i>	Corben's Long-eared Bat	Ecosystem		Yes	Fragmented - 11-30 % habitat retained	5 - 24 ha	Yes		Inhabits a variety of vegetation types, including mallee, bullocke Allocasuarina leuhmanni and box eucalypt dominated communities, but it is distinctly more common in box/ironbark/cypress-pine vegetation that occurs in a north-south belt along the western slopes and plains of NSW and southern Queensland.;1 Roosts in tree hollows, crevices, and under loose bark.;2 Slow flying agile bat, utilising the understorey to hunt non-flying prey - especially caterpillars and beetles - and will even hunt on the ground.;3 Mating takes	Potential habitat is present within the development footprint	Moderate

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									place in autumn with one or two young born in late spring to early summer.;4		
<i>Petauroides volans</i>	Greater Glider	Ecosystem		Yes	Variegated - 31-70% habitat retained	5 - 24 ha	Yes			Habitat occurs within and adjacent to the development footprint	Recorded by survey
<i>Petaurus australis</i>	Yellow-bellied Glider	Ecosystem		Yes	Variegated - 31-70% habitat retained	25 - 100 ha	Yes		Occur in tall mature eucalypt forest generally in areas with high rainfall and nutrient rich soils.;1 Forest type preferences vary with latitude and elevation; mixed coastal forests to dry escarpment forests in the north; moist coastal gullies and creek flats to tall montane forests in the south.;2 Feed primarily on plant and insect exudates, including nectar, sap, honeydew and manna with pollen and insects providing protein.;3 Extract sap by	Habitat occurs within and adjacent to the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									incising (or biting into) the trunks and branches of favoured food trees, often leaving a distinctive V-shaped scar.;4 Live in small family groups of two - six individuals and are nocturnal.;5 Den, often in family groups, in hollows of large trees.;6 Very mobile and occupy large home ranges between 20 to 85 ha to encompass dispersed and seasonally variable food resources.;7		
<i>Petaurus norfolcensis</i>	Squirrel Glider	Species		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Inhabits mature or old growth Box, Box-Ironbark woodlands and River Red Gum forest west of the Great Dividing Range and Blackbutt-Bloodwood forest with heath understorey in coastal areas.;1 Prefers mixed species stands with a shrub or Acacia midstorey.;2 Live	Potential habitat is present within the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									in family groups of a single adult male one or more adult females and offspring.;3 Require abundant tree hollows for refuge and nest sites.;4 Diet varies seasonally and consists of <i>Acacia</i> gum, eucalypt sap, nectar, honeydew and manna, with invertebrates and pollen providing protein.;5		
<i>Petrogale penicillata</i>	Brush-tailed Rock-wallaby	Species		Yes	Fragmented - 11-30 % habitat retained	5 - 24 ha	Yes		Occupy rocky escarpments, outcrops and cliffs with a preference for complex structures with fissures, caves and ledges, often facing north.;1 Shelter or bask during the day in rock crevices, caves and overhangs and are most active at night when foraging.;2 Browse on vegetation in and adjacent to rocky areas eating	Potential habitat is present within the development footprint	Moderate

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									grasses and forbs as well as the foliage and fruits of shrubs and trees.; ³ Highly territorial and have strong site fidelity with an average home range size of about 15 ha. Males tend to have larger home ranges than females.; ⁴ The home range consists of a refuge area and a foraging range linked by habitually used commuting routes.; ⁵ Females settle in or near their mother's range, while males mainly disperse between female groups within colonies, and less commonly between colonies.; ⁶ Dominant males associate and breed with multiple females.; ⁷ Breeding occurs throughout the year with a peak in births between February and May, especially in the southern		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									parts of the range and at higher altitudes.;8		
<i>Phascogale tapoatafa</i>	Brush-tailed Phascogale	Species		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		Prefer dry sclerophyll open forest with sparse groundcover of herbs, grasses, shrubs or leaf litter.;1 Also inhabit heath, swamps, rainforest and wet sclerophyll forest.;2 Agile climber foraging preferentially in rough barked trees of 25 cm DBH or greater.;3 Feeds mostly on arthropods but will also eat other invertebrates, nectar and sometimes small vertebrates.;4 Females have exclusive territories of approximately 20 - 40 ha, while males have overlapping territories often greater than 100 ha.;5 Nest and shelter in tree hollows with entrances 2.5 - 4 cm wide and use	Potential habitat is present within the development footprint. However, BioNet notes the species occurrences in the following IBRA subregions relevant to the project site. Walcha Plateau IBRA - Known to occur, but a geographic restriction exists stating "East of the Tia River". This river's headwaters occur >50kms north-east of the assessment area. Nearest record of	Negligible

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									many different hollows over a short time span.; ⁶ Mating occurs May - July; males die soon after the mating season whereas females can live for up to three years but generally only produce one litter.; ⁷	the species is 56kms east. Tomala IBRA - species known, with no geographic restrictions listed. However, only records of the species comprise an inaccurate record (10kms) noted as Mount Royal SF (or NP) from 1991, one more low accuracy (10kms) in similar location (but in Barrington Tops IBRA), one further single record in the IBRA from 1974, and >66kms from the assessment area.	

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
										Peel IBRA - Species predicted to occur (i.e. not known), no geographic restrictions listed. Species never recorded in IBRA.	
<i>Phascolarctos cinereus</i>	Koala	Species/ Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes	Other; Areas identified via survey as important habitat (see comments)	Inhabit eucalypt woodlands and forests.; ¹ Feed on the foliage of more than 70 eucalypt species and 30 non-eucalypt species, but in any one area will select preferred browse species.; ² Inactive for most of the day, feeding and moving mostly at night.; ³ Spend most of their time in trees, but will descend and traverse open ground to move between trees.; ⁴ Home range size varies with quality of habitat, ranging from less than two ha to several	Habitat occurs within and adjacent to the development footprint	Recorded by survey

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									hundred hectares in size.;5 Generally solitary, but have complex social hierarchies based on a dominant male with a territory overlapping several females and subordinate males on the periphery.;6 Females breed at two years of age and produce one young per year.;7		
<i>Phoniscus papuensis</i>	Golden-tipped Bat	Ecosystem		Yes	Variegated - 31-70% habitat retained	< 5 ha	Yes		Found in rainforest and adjacent wet and dry sclerophyll forest up to 1000m. Also recorded in tall open forest, <i>Casuarina</i> -dominated riparian forest and coastal <i>Melaleuca</i> forests.;1 Bats will fly up to two kilometres from roosts to forage in rainforest and sclerophyll forest on mid and upper-slopes.;2 Roost	Potential habitat is present within the development footprint	Moderate

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									<p>mainly in rainforest gullies on small first- and second-order streams in usually abandoned hanging Yellow-throated Scrubwren and Brown Gerygone nests modified with an access hole on the underside. Bats may also roost under thick moss on tree trunks, in tree hollows, dense foliage and epiphytes.;³ Bats will use multiple roost and change roosts regularly.;⁴ Bats roost individually or in small colonies which can contain up to approximately 20 bats of both males and females or just a single sex.;⁵ Maternity roosts may occur away from water sources with one maternity roost found 450m upslope of the nearest water course in a broken bough.;⁶ Specialist feeder</p>		

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									on small web-building spiders.;7 There is one breeding cycle per year.;8		
<i>Potorous tridactylus</i>	Long-nosed Potoroo	Species	No	Yes	Variegated (between 31 and 70% habitat retained)	5 - <25 ha	Yes	Dense shrub layer or alternatively high canopy cover exceeding 70% (i.e. to capture populations inhabiting wet sclerophyll and rainforest))	Inhabits coastal heaths and dry and wet sclerophyll forests. Dense understorey with occasional open areas is an essential part of habitat, and may consist of grass-trees, sedges, ferns or heath, or of low shrubs of tea-trees or melaleucas. A sandy loam soil is also a common feature.;1 The fruit-bodies of hypogeous (underground-fruited) fungi are a large component of the diet of the Long-nosed Potoroo. They also eat roots, tubers, insects and their larvae and other soft-bodied animals in the soil.;2 Often digs small holes in the ground in a similar way to	Inhabits coastal heaths and dry and wet sclerophyll forests. Dense understorey with occasional open areas is an essential part of habitat, and may consist of grass-trees, sedges, ferns or heath, or of low shrubs of tea-trees or melaleucas. A sandy loam soil is also a common feature. this habitat is not present with the development footprint,	Low

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									bandicoots.;3 Mainly nocturnal, hiding by day in dense vegetation - however, during the winter months animals may forage during daylight hours.;4 Individuals are mainly solitary, non-territorial and have home range sizes ranging between 2-5 ha.;5 Breeding peaks typically occur in late winter to early summer and a single young is born per litter. Adults are capable of two reproductive bouts per annum.;6	furthermore the species was not recorded during terrestrial mammal camera trap surveys. Records <35-70kms to the south-east, east, north-east of the subject land	
<i>Pteropus poliocephalus</i>	Grey-headed Flying-fox	Species/ Ecosystem		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes	Other; Breeding camps	Occur in subtropical and temperate rainforests, tall sclerophyll forests and woodlands, heaths and swamps as well as urban gardens and cultivated fruit crops.;1 Roosting camps are generally located within	Potential forage habitat supported across the development footprint	Recorded by survey

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									20 km of a regular food source and are commonly found in gullies, close to water, in vegetation with a dense canopy.;2 Individual camps may have tens of thousands of animals and are used for mating, and for giving birth and rearing young.;3 Annual mating commences in January and conception occurs in April or May; a single young is born in October or November.;4 Site fidelity to camps is high; some camps have been used for over a century.;5 Can travel up to 50 km from the camp to forage; commuting distances are more often <20 km.;6 Feed on the nectar and pollen of native trees, in particular Eucalyptus , Melaleuca and Banksia , and		

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									fruits of rainforest trees and vines.;7 Also forage in cultivated gardens and fruit crops.;8		
<i>Saccolaimus flaviventris</i>	Yellow-bellied Sheathtail-bat	Ecosystem		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	Yes		Roosts singly or in groups of up to six, in tree hollows and buildings; in treeless areas they are known to utilise mammal burrows.;1 When foraging for insects, flies high and fast over the forest canopy, but lower in more open country.;2 Forages in most habitats across its very wide range, with and without trees; appears to defend an aerial territory.;3 Breeding has been recorded from December to mid-March, when a single young is born.;4 Seasonal movements are unknown; there is speculation about a migration to southern	Habitat occurs within and adjacent to the development footprint	Recorded by survey

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									Australia in late summer and autumn.;5		
<i>Scoteanax rueppellii</i>	Greater Broad-nosed Bat	Ecosystem		Yes	Variegated - 31-70% habitat retained	5 - 24 ha	Yes		Utilises a variety of habitats from woodland through to moist and dry eucalypt forest and rainforest, though it is most commonly found in tall wet forest.;1 Although this species usually roosts in tree hollows, it has also been found in buildings.;2 Forages after sunset, flying slowly and directly along creek and river corridors at an altitude of 3 - 6 m.;3 Open woodland habitat and dry open forest suits the direct flight of this species as it searches for beetles and other large, slow-flying insects; this species has been known to eat other bat species.;4 Little is known of its reproductive	Habitat occurs within and adjacent to the development footprint	Recorded by survey

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									cycle, however a single young is born in January; prior to birth, females congregate at maternity sites located in suitable trees, where they appear to exclude males during the birth and raising of the single young.;5		
<i>Thylogale stigmatica</i>	Red-legged Pademelon	Ecosystem		Yes	Fragmented - 11-30 % habitat retained	5 - 24 ha	Yes		Inhabits forest with a dense understorey and ground cover, including rainforest, moist eucalypt forest and vine scrub.;1 Wet gullies with dense, shrubby ground cover provide shelter from predators.;2 In NSW, rarely found outside forested habitat.;3 They disperse from dense shelter areas to feed from late afternoon to early morning, favouring native grasses and herbs on the edge of the forest.;4 Also known to	Potential forage habitat supported across the development footprint	Moderate

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									feed on fruits, young seedling leaves and stems, fungi and ferns.;5		
<i>Vespadelus trouhoni</i>	Eastern Cave Bat	Species		Yes	Fragmented - 11-30 % habitat retained	5 - 24 ha	Yes	Caves; Within two kilometres of rocky areas containing caves, overhangs, escarpments, outcrops, crevices or boulder piles, or within two kilometres of old mines, tunnels, old buildings or sheds."	Very little is known about the biology of this uncommon species.;1 A cave-roosting species that is usually found in dry open forest and woodland, near cliffs or rocky overhangs; has been recorded roosting in disused mine workings, occasionally in colonies of up to 500 individuals.;2 Occasionally found along cliff-lines in wet eucalypt forest and rainforest.;3 Little is understood of its feeding or breeding requirements or behaviour.;4	Habitat occurs within and adjacent to the development footprint	Recorded by survey
Reptiles											
<i>Hoplocephalus bitorquatus</i>	Pale-headed Snake	Species		Yes	Fragmented - 11-30 % habitat retained	< 5 ha	No		The Pale-headed Snake is a highly cryptic species that can spend weeks at a time	Species known only to occur at altitudes much	Negligible

BAM candidate species identification			Step 1: Identify threatened species for assessment					Step 2	Step 3: Identify candidate species credit species for further assessment		
Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									hidden in tree hollows.;1 Found mainly in dry eucalypt forests and woodlands, cypress forest and occasionally in rainforest or moist eucalypt forest.;2 In drier environments, it appears to favour habitats close to riparian areas.;3 Shelter during the day between loose bark and tree-trunks, or in hollow trunks and limbs of dead trees.;4 The main prey is tree frogs although lizards and small mammals are also taken.;5 The Pale-headed Snake is relatively unusual amongst elapid snakes in that it is well adapted to climbing trees.;6	lower than the development footprint, within highest elevation BioNet records including 550m elevation (approx.) north of Bindarri NP (>200kms from the project site), 390m elevation (approx.) west of Kwiambal NP (>150kms from the project site) and 375m elevation (approx.) west of Gunnedah (>100kms from the project site). The lowest point of the project site occurs along the transmission line at an altitude of	

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										750 metres (approx.) and as such the development footprint does not support habitat for the species.	
<i>Hoplocephalus stephensii</i>	Stephens' Banded Snake	Species	No	Yes	Variegated (between 31 and 70% habitat retained)	25 - <100 ha	Yes	Hollow bearing trees;Or within 500 m of this habitat Other;Within 500 m of aboreal vine tangles Fallen/standing dead timber including logs;Or within 500 m of this habitat	Rainforest and eucalypt forests and rocky areas up to 950 m in altitude.;1 Stephens' Banded Snake is nocturnal, and shelters between loose bark and tree trunks, amongst vines, or in hollow trunks limbs, rock crevices or under slabs during the day.;2 At night it hunts frogs, lizards, birds and small mammals.;3	Rainforest and eucalypt forests and rocky areas up to 950 m in altitude. The species uses very old primary forest with many large old hollow bearing trees. Habitat needs to be well connected and geographically large. Potential habitat combining old primary forest	

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										<950m elevation does not occur within the development footprint. Records >35kms to the east of the subject land	
<i>Uvidicolus sphyrurus</i>	Border Thick-tailed Gecko	Species		Yes	Relictual - 10% or less habitat retained	< 5 ha	Yes		As implied by another of its common names (Granite Thick-tailed Gecko), this species often occurs on steep rocky or scree slopes, especially granite. Recent records from basalt and metasediment slopes and flats indicate its habitat selection is broader than formerly thought and may have extended into areas that were cleared for agriculture.; ¹ Favours forest and woodland areas with boulders, rock slabs, fallen timber and deep leaf litter. Occupied sites often	Species distribution is north of the project site and has never been recorded (or predicted to occur in) Tomala or Walcha Plateau IBRA subregions. Peel IBRA has records 20-25kms north of the site across cleared land, which are at the southern extent of the species'	Low

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									have a dense tree canopy that helps create a sparse understorey.;2 These Geckos are active at night and shelter by day under rock slabs, in or under logs, and under the bark of standing trees.	occurrence. Peel IBRA abuts parts of the subject land and includes the western 60% of the transmission line corridor.	
Plants											
<i>Acacia atrox</i>	Myall Creek Wattle	Species	N/A	Yes	N/A	N/A	Yes	N/A	Species grows in soils ranging from deep black clay over basalt to shallow red stony loams on the upper slope and crest of a low hill. Currently known from two populations near Delungra and Gurley. There individuals grow in a partly cleared paddock of box woodland with a native grassy understorey.	Known populations more than 200km north/northwest of the assessment area. No records within proximity to the site. Potential habitat in PCT599 is marginal and unlikely to support the species.	Unlikely
<i>Callistemon</i>	Callistemon	Species	N/A	Yes	N/A	N/A	Yes	N/A	Habitats range from	Habitats range	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
<i>pungens</i>	pungens								riparian areas dominated by Casuarina cunninghamiana subsp. cunninghamiana to woodland and rocky shrubland.;1 Often in rocky watercourses, usually with sandy granite (occasionally basalt) creek beds;2 Flowers over spring and summer, mostly in November.;3	from riparian areas dominated by Casuarina cunninghamiana subsp. cunninghamiana to woodland and rocky shrubland. Often in rocky watercourses, usually with sandy granite (occasionally basalt) creek beds. Marginal habitat may occur along the transmission line corridor however the subject land occur outside the known area of occurrence of the species. Records <80-100kms to the north and north-	

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
										east, with the species never having been recorded south of Tamworth / Port Macquarie	
<i>Eucalyptus nicholii</i>	Narrow-leaved Black Peppermint	Species	N/A	No	N/A	N/A	Yes	N/A	Typically grows in dry grassy woodland, on shallow soils of slopes and ridges- prefers infertile soils derived from granite or metasedimentary rock. The species occurs from Nundle to north of Tenterfield being most common in the central portions of its range.	Potential habitat within grassy woodland and dry sclerophyll forests within the site	Possible
<i>Chiloglottis anaticeps</i>	Bird Orchid	Species	N/A	Yes	N/A	N/A	Yes	N/A	Bird Orchid often grows near streams or on the edges of low, flat rock outcrops.; ¹ It grows in eucalypt forest in areas with very little ground cover, in gravelly loam soils.; ²	Often grows near streams or on the edges of low, flat rock outcrops, in grows in eucalypt forest in areas with very little ground cover, in	No

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										<p>gravely loam soils. Very broad habitat descriptions do not fit well with the habitats or soils present within the development footprint. Furthermore the subject land occurs well outside the species known area of occurrence..</p> <p>Records of two known areas of occurrence occur >100kms and >150 kms to the north-east of the subject land</p>	
<i>Chiloglottis platyptera</i>	Barrington Tops Ant	Species	N/A	Yes	N/A	N/A	Yes	N/A	Grows in moist areas in tall open Eucalypt forest with a	Potential habitat within grassy	Likely

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	Orchid								grassy understorey and also around rainforest edges; generally on rich brown loam soils. Known to occur within the area including at Ben Halls Gap Nature Reserve.	woodland and open forests within the site.	
<i>Cryptostylis hunteriana</i>	Leafless Tongue Orchid	Species	N/A	Yes	N/A	N/A	Yes	N/A	Does not appear to have well defined habitat preferences and is known from a range of communities, including swamp-heath and woodland.; ¹ The larger populations typically occur in woodland dominated by Scribbly Gum (<i>Eucalyptus sclerophylla</i>), Silvertop Ash (<i>E. sieberi</i>), Red Bloodwood (<i>Corymbia gummifera</i>) and Black Sheoak (<i>Allocasuarina littoralis</i>); appears to prefer open areas in the understorey of this community and is often found in association	EPBC Act SPRAT profile sates habitat associated for the species in the NSW Northern Tables lands regions include New England Blackbutt (<i>Eucalyptus andrewsii</i>) Grassy Forest and New England Blackbutt Shrubby Forest and Large-fruited Blackbutt (<i>Eucalyptus</i>	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									with the Large Tongue Orchid (<i>C. subulata</i>) and the Tartan Tongue Orchid (<i>C. erecta</i>); ² Little is known about the ecology of the species; being leafless it is expected to have limited photosynthetic capability and probably depends upon a fungal associate to meet its nutritional requirements from either living or dead organic material; ³ In addition to reproducing from seed, it is also capable of vegetative reproduction and thus forms colonies which can become more or less permanent at a site; ⁴ On the Central Coast of NSW, populations have been recorded in woodland dominated by Scribbly Gum (<i>Eucalyptus haemastoma</i>), Brown	pyrocarpa) / Strawberry Gum (<i>Eucalyptus olida</i>) Woodland. These habitats do not occur within the subject land. Records >35-40kms to the north-east of the subject land	

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									Stringybark (Eucalyptus capitellata), Red Bloodwood (Corymbia gummifera) and also associated with Large Tongue Orchid (C. subulata) and the Tartan Tongue Orchid (C. erecta);,5		
<i>Dichanthium setosum</i>	Bluegrass	Species	N/A	Yes	N/A	N/A	Yes	N/A	Found in heavy basaltic black soil and red-brown loams with clay subsoil. Often in moderately disturbed areas including cleared woodland, grassy roadside remnants and high disturbed pasture; associated with species including Eucalyptus albens, E.melanophloia, E.melliadora, E.viminalis, Myoporum debile, Aristida ramosa, Themeda triandra, Poa sieberiana, Bothriochloa ambigua, Medicago minima.	Potential habitat within dry sclerophyll forests, derived native grassland and forested wetlands within the site.	Possible

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<i>Digitaria porrecta</i>	Finger Panic Grass	Species	N/A	Yes	N/A	N/A	Yes	N/A	Occurs in native grassland, woodland or open forest with a grassy understorey on richer soils. Typically associated with E.albens, Acacia pendula Austrostipa aristiglumis, Enteropogon acicularis, Cyperus bifax, Hibiscus trionum and Neptunia gracilis.	Habitat within box woodland marginal for the species. No other suitable habitat within the site.	Unlikely
<i>Diuris pedunculata</i>	Small Snake Orchid	Species	N/A	Yes	N/A	N/A	Yes	N/A	The Small Snake Orchid grows on grassy slopes or flats.;1 Often on peaty soils in moist areas.;2 Also on shale and trap soils, on fine granite, and among boulders.;3 It flowers during late September-October.;4 Pollination is mostly by sexual deception, with the Small Snake-orchid attracting mostly males of a native bee Halictus lanuginosus. However, the flowers produce nectar and emit a	Grows on grassy slopes or flats, often on peaty soils in moist areas, also on shale and trap soils, on fine granite, and among boulders. This habitat is not present within the development footprint. Nearest record occurs	No

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									strong scent that attracts a range of other pollinators;5]	approximately 7km to the north of the subject land. However the records is noted as a Royal Botanic Gardens Herbarium Specimen Register (with an "endDate" noted as 29/10/2000), and despite its relatively high level of accuracy (50m) it is located in the centre of a pine plantation that was recently cleared and is now regrowing, thus puts its accuracy into question. The next nearest records occur >60kms to the	

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										north, >115kms to the north-east, and >100kms to the south-east.	
<i>Eucalyptus magnificata</i>	Northern Blue Box	Species	N/A	Yes	N/A	N/A	Yes	N/A	Grassy open forest or woodland on shallow, sandy or loamy soils.; ¹ Occurs on moderately hilly sites and at the edge of gorges, usually at altitudes from 900 - 1050 m.; ² Known populations are small, numbering 5-400 plants per location.; ³	Occurs in grassy open forest or woodland on shallow, sandy or loamy soils, on moderately hilly sites and at the edge of gorges, usually at altitudes from 900 - 1050 m. Potential habitat combining grassy open forest or woodland on shallow, sandy or loamy soils at 900 - 1050 m elevation does not occur within the development footprint.	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
										Records >90-100kms to the north-east of the subject land	
<i>Eucalyptus oresbia</i>	Small-fruited Mountain Gum	Species	N/A	Yes	N/A	N/A	Yes	N/A	Found at altitudes between 800 and 1100 m in very steep valleys and deeply incised creeklines with primarily south to southwest exposure (i.e. warm yet moist). ;1 Three subpopulations exist with some outlying individuals.;2 Replaced by Eucalyptus pauciflora, E. elliptica and E. melliodora in less rocky habitat.;3 Seedlings more commonly observed in disturbed areas and rare elsewhere. ;4 Age structure suggests populations are largely regrowth after disturbance from mining, forestry and road construction during	Steep valleys and deeply incised creeklines do not occur within the development footprint	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									the past century (data collected by J.T. Hunter and L. M. Copeland) ;5		
<i>Euphrasia ciliolata</i>	Polblue Eyebright	Species	N/A	Yes	N/A	N/A	Yes	N/A	Occurs on the edge of montane and sub-alpine swamps and on open grassy slopes bordering swamps, Snow Grass meadows, Snow Gum woodland, open boggy meadows amidst Black Sallee woodland, and in seasonally inundated upland grassland.;1 Flowers December to May.;2 Polblue Eyebright is an 'annual', ie each plant lives for one year and dies after flowering and fruit-set has finished. Some plants may merely die back to ground level, leaving the underground parts to resprout the following season. This annual	Species Flowers December to May and occurs on the edge of montane and sub-alpine swamps and on open grassy slopes bordering swamps, Snow Grass meadows, Snow Gum woodland, open boggy meadows amidst Black Sallee woodland, and in seasonally inundated upland grassland. Potential habitat for the species within the	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									'disappearance' means that surveys for this plant must be confined to its flowering season.; ³ Polblue Eyebright is partly parasitic upon other plants, such as Snow Grass, and may only be able to grow in close proximity to such host plants.; ⁴	development footprint, comprising PCT 586, is degraded by weed invasion. Furthermore the species was not recorded during targeted meander and plot surveys undertaken within degraded habitats in March 2021. Records >15 kms to the north-east of the subject land	
<i>Homoranthus prolixus</i>	Granite Homoranthus	Species	N/A	Yes	N/A	N/A	Yes	On or near granite outcrops and slabs or within 100m	Grows in heath patches, in skeletal soil among crevices of granite outcrops within the Ironbark Nature Reserve (east of Barraba) and neighbouring	No suitable habitat within the site	Unlikely

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									properties. The species has not been recorded in a survey of other granitic outcrop areas in the region.		
<i>Monotaxis macrophylla</i>	Large-leafed Monotaxis	Species	N/A	Yes	N/A	N/A	Yes	N/A	Found in coastal heath, arid shrubland, forests and montane heath from sea level to 1300m altitude, subject to regular fire.	No suitable habitat within the site	Unlikely
<i>Picris evae</i>	Hawkweed	Species	N/A	Yes	N/A	N/A	Yes	N/A	Occurs within open Eucalypt forest including a canopy of E.melliodora, E.crebra, E.populnea, E.albens, Angophora subvelutina, Allocasuarina torulosa, Cunninghamiana with a Dichanthium grassy understorey. Recorded north of Inverell and at Oxley Park Tamworth.	Open Eucalypt woodland within site does not support Dichanthium spp. dominated ground layer and is marginal for the species.	Unlikely
<i>Polygala linariifolia</i>	Native Milkwort	Species	N/A	Yes	N/A	N/A	Yes	N/A	Sandy soils in dry eucalypt forest and woodland with a sparse understorey. The species has been recorded from the Inverell and	Potential habitat within PCT 1194	Possible.

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									Torrington districts growing in dark sandy loam on granite in shrubby forest of Eucalyptus caleyi, Eucalyptus dealbata and Callitris, and in yellow podsolic soil on granite in layered open forest		
<i>Prasophyllum sp. Wybong</i>	Prasophyllum sp. Wybong	Species	N/A	Yes	N/A	N/A	Yes	N/A	A perennial orchid, appearing as a single leaf over winter and spring;1 Flowers in spring and dies back to a dormant tuber over summer and autumn;2 Known to occur in open eucalypt woodland and grassland ;3	BioNet notes that the species is known to occur in open eucalypt woodland and grassland, however species records to not occur in habitats that are remotely similar to those present within the development footprint. Impacts associated with the transport haul routes are	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
										to highly degraded road edges, that do not support habitat for the species. Records >100kms to the north-west and >75kms to the south-west of the subject land	
<i>Pterostylis elegans</i>	Elegant Greenhood	Species	N/A	Yes	N/A	N/A	Yes	N/A	Known to occur on red-brown loams at elevations between 950m and 1200m. Found among grass and shrubs in tall open Eucalypt forest.	BioNet notes that the species restricted distribution from the Barrington Tops to the Walcha district, which is outside the occurrence of the subject land. Records >20-25kms to the east of the subject land	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
<i>Pterostylis riparia</i>	Pterostylis riparia	Species	Species	N/A	Yes	N/A	N/A	Yes	Grows on the edge of small streams under shrubs.;1 Forms small clonal colonies.;2 May be inundated during times of high flow.;3	BioNet notes the species grows on the edge of small streams under shrubs, and is restricted to the Barrington Tops. Riparian habitats within the development footprint are degraded and generally relate to ephemeral first order watercourses at the highest point of the catchment. These areas do not support habitat for the species. Records >30-50kms to the east and south-east of the subject land	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
<i>Commersonia procumbens</i>	Commersonia procumbens	Species	N/A	N/A	N/A	N/A	Yes	Piliga sandstone	Grows in sandy sites, often along roadsides. The species is often found as a pioneer species of disturbed habitats. It has been recorded colonising disturbed areas such as roadsides, the edges of quarries and gravel stockpiles and a recently cleared easement under power lines.	No PCTs known to be associated with the species occur within the site	Unlikely
<i>Senna acclinis</i>	Rainforest Cassia	Species	Yes	N/A	N/A	N/A	Yes	N/A	Grows on the margins of subtropical, littoral and dry rainforests.; ¹ Often found as a gap phase shrub.; ² Flowering occurs in spring and summer and the fruit is ripe in summer and autumn.; ³ Primarily pollinated by a variety of bees.; ⁴	Species grows on the margins of subtropical, littoral and dry rainforests. The subject land does not support habitat for the species. Records >40-60kms to the east and south-east of the subject land	No

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
<i>Tasmannia glaucifolia</i>	Fragrant Pepperbush	Species	Yes	N/A	N/A	N/A	Yes	N/A	Usually grows in or near Antarctic Beech <i>Nothofagus moorei</i> rainforest along streams in mountain areas at altitudes of between 1200 and 1500 m altitude. Also occurs in tall scrub, on seepage lines in tall eucalypt forest and in grassy woodland.	Eucalypt forest within PCT 934, 931 and 927 offers marginal habitat for the species.	Possible
<i>Tasmannia purpurascens</i>	Broad-leaved Pepperbush	Species	Yes	N/A	N/A	N/A	Yes	N/A	Grows in tall, moist eucalypt forest, sub-alpine woodland and cool temperate rainforest. Has been observed growing on cleared land, logged forest and graded fire trails.	Suitable habitat within open woodland and forest within the site (PCT 934, 931, 927 and 1194)	Likely
<i>Thesium australe</i>	Austral Toadflax	Species	Yes	N/A	N/A	N/A	Yes	N/A	It occurs in shrubland, grassland or woodland, often on damp sites. Vegetation types include open grassy heath dominated by <i>Leptospermum myrtifolium</i> , <i>Hakea microcarpa</i> , <i>Callistemon</i>	Suitable habitat within open woodland, Eucalypt forest and derived native grasslands.	Possible

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									sieberi, Grevillea lanigera, Epacris microphylla and Poa spp., Kangaroo Grass grassland surrounded by Eucalyptus woodland; and grassland dominated by Cymbopogon refractus.		
<i>Tylophora linearis</i>	Tylophora linearis	Species	Yes	N/A	N/A	N/A	Yes	N/A	Grows in dry scrub and open forest. Records from low altitude (300-400m) sedimentary flats and dry woodlands of E.fribosa, E.sideroxylon, E.albens, Callitris endlicheri, Callitris glaucophylla and Allocasuarina luehmannii.	Associated PCTs within the subject land occur at higher altitudes than recorded for the species.	Unlikely
<i>Asterolasia</i> sp. 'Dungowan Creek'	Dungowan Starbush	Species	Yes	N/A	N/A	N/A	Yes	N/A	In the vicinity of Dungowan Dam the Dungowan Starbush grows in rocky alluvial soil along a creekbank dominated by Casuarina cunninghamiana with or without Eucalyptus viminalis. Recent populations have been found growing near (100-	Marginal habitat within PCT 934.	Possible.

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Species name	Common name	Credit class	Species geographic constraints	Species associated with site PCT?	Native vegetation cover required	Required patch size	Requires further assessment	Habitat constraints	Suitable habitat	Habitat assessment	Likelihood of occurrence
									150m) major drainage lines on lower and mid slopes in open forest in moderately deep brown loamy soils. Overstorey trees at these locations were dominated by Eucalyptus obliqua and E. nobilis with or without E. radiata ssp. sejuncta.		
<i>Homopholis belsonii</i>	Belson's Panic	Species	Yes	N/A	N/A	N/A	Yes	N/A	Grows in dry woodland (e.g. Belah) often on poor soils. Found mostly on heavy texture cracking soils derived from basalt or alluvials between 00-520m altitude.	Site lacks suitable habitat.	Unlikely
<i>Euphrasia arguta</i>	Euphrasia arguta	Species	Yes	N/A	N/A	N/A	Yes	N/A	Plants from the Nundle area have been recorded in Eucalypt forest with a mixed grass and shrub understorey. Also know to occur in highly disturbance areas including road edges	Suitable habitat within the assessment area	Likely

* Conservation advice taken from BioNet and Commonwealth SPRAT databases

Appendix D Collision Risk Model Report



Hills of Gold Wind Farm Bird Collision Risk Assessment

Prepared for Arup Group Pty Ltd
30 September 2020

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1. Introduction

This document presents a collision risk assessment for birds at the proposed Hills of Gold Wind Farm, near Nundle New South Wales (NSW).

Biosis Pty Ltd has conducted a range of ecological assessments for the project. These surveys have included seasonal bird utilisation surveys at 21 sites across the wind farm study area. The data collected during those surveys forms the basis of the collision risk model presented in this report.

A background to risk modelling is provided in Section 1.1, and a description of the Biosis collision risk model can be found in Smales et al. (2013), provided in Appendix 1.

The risk modelling for Hills of Gold Wind Farm followed a three-stage approach, as detailed in this report. In the first stage, a brief evaluation was made of three different models of turbine that are under consideration for the project. The object of this stage was to ascertain which turbine might represent a 'worst-case' collision risk. This turbine was then used for subsequent stages of modelling with the intention that if either of the other two turbines are chosen for operational use by the project, the potential collision risk they pose will be lower than that presented herein.

In the second stage, the collision risk model was run to obtain results expressed as the probable annual number of flights at risk of collision (see Section 1.2.3) for all species that were recorded flying at rotor-swept height for the 'worst-case' turbine.

In the third stage, a likely size of the site-population was estimated for species of raptors and those values were incorporated into the model to obtain results expressed as the probable annual number of collisions (see Section 1.2.3) for those species.

1.1 Background to quantitative risk modelling

Collisions of birds and bats with wind turbines have been documented to occur at various frequencies around the world. Quantitative modelling to estimate the number of collision mortalities of threatened and non-threatened taxa is widely used as part of environmental impact assessments for proposed wind energy facilities (Masden & Cook 2016).

The impact of any collisions on the viability of threatened and non-threatened fauna populations is more important than determination of simple numbers of mortalities, and population models can be used in combination with results of collision risk models to evaluate such impacts, but population modelling would be a separate exercise to the collision risk modelling presented here.

Modelling of collision risk is reliant on empirical data for flights by species at the wind farm site. There is no practicable method to obtain species-specific flight data for bats that are likely to utilise the site and bat-call data does not provide information about the number of flights by bats. The modelling presented here is therefore confined to diurnal birds.

Mathematical modelling of risk is intended to provide an articulated, transparent and replicable evaluation of what may occur in the real world. The rationale behind projections is explicitly stated in the mathematics of a model, which means that the logical consistency of the predictions can be easily evaluated. The explicit nature of inputs and rigour entailed in modelling means that the process is replicable and consistent and it is open to analysis, criticism or modification when new information becomes available. Modelling is designed as a mechanism to evaluate uncertainties – if there was no uncertainty there would be no need to use a model. As a consequence of

uncertainty in various parameters, some assumptions are required and while it is necessary to include some assumptions and arbitrary choices when deciding on the structure and parameters of a model, these choices are explicit. To the extent feasible, assumptions are informed by the best available information.

Models are also valuable for their heuristic capacities as they focus attention on important processes and parameters entailed in risk (Brook et al. 2002). Their very nature facilitates incorporation of information as it is learnt (Burgman 2005) and refinements should thus be expected of any model.

Most factors related to the layout, dimensions and geometry of turbines are known. The risk modelling detailed here entails the use of informed assumptions related particularly to the flights of birds. The bird utilisation data collected from the site provides an empirical basis for extrapolations required for use in the model. We consider the assumptions and values used are reasonable and they are informed by available information about the ecology of relevant species. As a consequence, we consider the results of modelling detailed here provide a basis for evaluation of probable effects of the proposed Hills of Gold Wind Farm on relevant species of birds.

The only alternative to a quantitative modelling approach is one of qualitative subjective judgement. All the benefits of using mathematical modelling outlined above are difficult, if not impossible to achieve with a purely qualitative assessment.

1.2 Turbine collision risk model

The risk of birds colliding with turbines at the proposed Hills of Gold Wind Farm has been assessed using the Biosis Pty. Ltd. Deterministic Collision Risk Model. The model was first developed in 2002 and has been refined over time to incorporate new data and knowledge, and has been applied at a wide range of proposed wind farm sites in Australia. A full description of the model (Smales *et al.* 2013) is provided in Appendix 1.

1.2.1 Overview of the model

The collision risk model takes account of bird flights that occur within the height zone that will be occupied by turbines. Data for the number of flights and their heights was documented by a regime of fixed-time point counts at locations representative of future turbine locations across the site. The model uses the empirical sample of flight data for each species and extrapolates that to determine a potential number of such flights that might occur over an entire 12-month period. This factor takes into account what is known about seasonal presence of particular species that may be migratory or may be present only for part of the year for other reasons.

In the model, the turbine is decomposed into its static and dynamic components. The entire turbine (including the tower, nacelle and the rotor *when stationary*) represents the static component. The dynamic component is the volume swept by the leading edge of the rotor blades in the time it takes the species of interest to pass across the depth of the swept disk.

Since the turbine tower below rotor swept height is always a static component and poses minimal collision risk, the model takes this into account by dividing flights into those below turbine rotor height, and those within the height zone swept by turbine rotors and allocating different risk rates to these height classes.

The risk assessment accounts for a combination of variables that are specific to the proposed wind farm and to data for birds from the vicinity of the farm. They include the following:

- The numbers flights of each species below rotor height, and for which just the lower portion of turbine towers present a collision risk.
- The numbers of flights at heights within the zone swept by turbine rotors, and for which the upper portion of towers, nacelles and rotors present a collision risk.
- The numbers of bird movements-at-risk, as recorded during timed point counts, extrapolated to determine an estimated number of movements-at-risk the species makes in an entire year. Account is taken of the portion of the year that birds are within proximity of the site and that they may thus be at risk.
- The mean area (m² per turbine), of tower, nacelle and stationary rotor blades of a wind generator that present a risk to birds. Thus, the mean area presented by a turbine is between the maximum (where the direction of the bird is perpendicular to the plane of the rotor sweep) and the minimum (where the direction of the bird is parallel to the plane of the rotor sweep). The mean presented area is determined from turbine specifications supplied to Biosis for specific make and model of a turbine. It represents the average area presented to an incoming flight from any direction.
- The additional area (m² per turbine) presented by the movement of rotors during the potential flight of a bird through a turbine. This information is determined via a calculation involving species-specific, independent parameters of flight speed and body length and supplied turbine specifications.
- The model assumes that all turbines in the site represent equal risk.
- A calculation of the average number of turbines a bird is likely to encounter in a given flight through the site. This is based on the scattered configuration of turbines in the landscape and the total number of turbines proposed for the project.

1.2.2 Avoidance rate

Results are provided for various avoidance rates. Avoidance rate is the capacity for a bird to avoid a collision, whether that occurs due to a cognitive response on the part of a bird or not. Thus at the extremes of the rates applied, a 0.90 avoidance rate equates to one flight in 10 in which a bird takes no action to avoid a turbine and a 0.99 avoidance rate equates to one flight in 100 in which it does not avoid a turbine. Based on experience with a wide range of bird species, it is certain that virtually all species have high capacity to avoid collision with the static components of turbines. Avoidance rate for these components is thus consistently considered to be 0.999 in the modelling. Various avoidance rates are modelled for the dynamic turbine components because it is not certain how adept various species may be at evading collision with the moving rotor. For this reason, results are provided for 0.90, 0.95, 0.98 and 0.99 avoidance rates for the dynamic components (moving rotor) of turbines.

It should be noted that internationally there is very little empirical evidence for the actual avoidance rate for any bird species and for this reason it is prudent to provide a range of estimates that are considered to be reasonable. The evidence that is available suggests that avoidance capacity is species-specific and that the great majority of birds have very high avoidance capability that is higher than 0.98. Nonetheless, the avoidance rate of some large raptors in Australia appears to be between 0.93 and 0.95 (Smales et al. 2013; Smales 2017).

1.2.3 Result metrics – number of flights at risk vs number of collisions

Generally, the model's results are expressed as the number of flights at risk of collision per annum for each species. This is a relative measure that permits us to compare risk rates associated with various turbines or turbine configurations. It does not necessarily equate to the number of collisions that might occur because we do not know how many individual birds of each species use the site and may thus be at risk. The difference between flights at risk of collision and number of actual collisions can be simply explained by way of an example. If there are just two individuals of a given species occupying the wind farm, they may make multiple flights that could result in collisions, however the maximum number of fatalities that could occur is two. As can be seen from this example, the number of actual collisions can be no greater than the number of flights at risk, and if the site-population is small but the birds fly actively within the site, the number of collisions will always be considerably lower than the number of flights at risk.

In cases where a good estimate of the site-population for particular species can be made, the model permits that to be incorporated to provide results expressed as an annual estimate of collisions.

Existing knowledge of the population dynamics for most of the species at the Hills of Gold site, is not sufficient to allow an estimate to be made for their site-populations. However, for two resident raptors, an estimate of their possible site-populations has been made and the model has been run to provide a projection of results for them as an annual estimate of collisions.

The model cannot forecast the frequency of collisions around the predicted annual average and it is important to recognize that the number of any actual collisions that might occur can be expected to vary from year to year in a distribution around the average.

All results are provided to two significant figures, however they represent annual 'average' results and, of course actual bird fatalities will always be measured in numbers of individuals and the average results of modelling must represent a distribution that can be expected to vary from year-to-year around the mean.

2. Preliminary evaluation of turbine options

An array of 70 wind turbines is proposed for the Hills of Gold Wind Farm. At present the project is considering options for three different models of turbine (Vestas 5.6, GE 5.5, SGRE 6). The three turbines differ in various aspects that may affect the collision risk they pose to birds in flight. The differences include rotor-swept area, rotor-sweep height above the ground and rotor speed. As a consequence, the risk to various species of birds will differ between them in response to the documented data for flight heights.

The Biosis turbine collision risk model was initially applied to a single species (Wedge-tailed Eagle) for each of the three turbines with a view to providing a preliminary consideration of how they might differ and to determine which turbine might represent a 'worst-case' collision risk. Wedge-tailed Eagle was chosen for this purpose because it had the greatest number of flights recorded at the site that were within rotor-swept height for all three turbines. An avoidance rate of 0.95 was used (see below and Appendix 1). The size of the population of Wedge-tailed Eagles at risk for the project was determined as set out under *Raptor populations at-risk*, below.

Results of the preliminary assessment of the three turbines are shown in Table 1.

Table 1 Comparison of three turbine models for a configuration of 70 of each turbine. Results show projected annual collision mortalities of Wedge-tailed Eagles at 0.95 avoidance rate.

Turbine model	Projected annual collision mortalities of Wedge-tailed Eagles at 0.95 avoidance rate
Vestas 5.6	3.71
GE 5.5	2.90
SGRE 6	3.23

On the basis of the preliminary evaluation it was determined that the Vestas 5.6 turbine represents the likely 'worst-case' collision risk for birds at the proposed Hills of Gold Wind Farm. While it has a slightly smaller rotor-swept area than the SGRE 6 turbine, it has a greater rotor speed and that can factor significantly in collision risk for birds.

3. Model inputs and assumptions

The Biosis collision risk model requires a range of numeric inputs, to quantify the number of turbines, key dimensions of turbines, and to estimate bird utilisation characteristics, including the number of flights within and outside of rotor swept height for species to be included in the model.

3.1 Wind farm and turbine parameters

The collision risk model requires input values for 36 turbine parameters that include number and layout of turbines and multiple aspects of turbine dimensions and geometry.

Following the results of the preliminary evaluation, the collision risk model was run for the Vestas 5.6 turbine.

Key parameters used in this modelling are:

- Number of turbines: 70
- Turbine type: based on Vestas 5.6
- Turbine tower height: 139 m
- Rotor diameter: 162 m
- Rotational speed: 12.1 rpm

The rotor-swept area is 17331.5 m². The 162 metres diameter blades have a length of 81 metres, resulting in rotor swept height between 58 and 220 metres above the ground.

The landscape configuration of the proposed Hills of Gold Wind Farm is essentially a linear row of turbines. This is different from most wind energy facilities in Australia in which turbines are scattered across a site. In the more usual scattered or 'clustered' array, a bird has a high probability of encountering multiple turbines in a given flight. The configuration of Hills of Gold turbines is such that a bird is likely to encounter multiple turbines only in the rare event that it flies directly along the row of turbines. The collision risk model has a built-in function to account for this difference whereby the turbine array can have any setting from 100% of turbines fully clustered to 0% in which turbines are entirely linearly configured. Given the slight sinuosity of the ridge-top array, this factor was set to 5%.

3.2 Bird species data

Following the results of the preliminary evaluation, the collision risk model was run for all species for which there were any flights recorded at rotor-swept height for the Vestas 5.6 turbine. That included all flights documented from between 58 and 220 metres above the ground.

A total of 51 species of birds were recorded during investigations of the Hills of Gold site. While all of them may have capacity to fly at rotor swept height, 18 species were recorded doing so, and thus have data available for use in the model. They are listed in Table 3, below.

White-breasted Woodswallow is the only one of the modelled species that is known to be present seasonally as its population migrates to inland and northern Australia during the winter. It was modelled as being present at the site for nine months per annum.

With the exception of two species of raptors (Section 3.2.1), it is not feasible to estimate the site-population sizes of the species modelled. As a consequence, model results for those 16 species are expressed as the

number of flights at risk of collision per annum. The resident raptors, Nankeen Kestrel and Wedge-tailed Eagle, tend to occupy territories that remain stable over periods of several years and, because they are apex predators they occur at relatively low densities. In light of published studies for those species, Biosis undertook a process to estimate the sizes of their potential populations for the site. For those two species it was thus possible to run the model to provide results expressed as annual estimate of collisions.

3.2.1 Estimating site-populations for raptors

Information was collated from published sources to ascertain the likely number of Wedge-tailed Eagles and Nankeen Kestrels that might occupy the site and thus be at some risk of turbine collision. Brown Goshawk was recorded once only and is not considered further. There are a number of relevant studies of Wedge-tailed Eagle, but less for Nankeen Kestrel. Information for the latter was drawn from data collated in Marchant and Higgins (1993). The population dynamics of neither species has been studied at the Hills of Gold Wind Farm site.

The first item of information relates to average home-range size or documented spacing between home-ranges and to the number of birds that might occupy a given home-range. The second requires an understanding of the likely number of flying birds that might occupy a given territory. For the two species in question, published data indicates that they function as territorial pairs that usually attempt to raise one brood per annum. For the purposes of considering collision risk, the number of birds occupying a territory thus includes the adult pair and the average number of their offspring that fledge.

Wedge-tailed Eagle

Cherriman (2007) provided an overview of studies, including his own, that have investigated the size of Wedge-tailed Eagle territories in temperate regions. Territory sizes in studies near Perth (Cherriman 2007); at two other sites in the south-east of Western Australia (Ridpath and Brooker 1987); near Canberra in south-eastern Australia (Leopold and Wolfe 1970); and, in South Australia (Rowe et al. 2017) were all between 31 km² and 42 km². Foster and Wallis (2010) studied the species west of Melbourne and recorded nearest-neighbour distances averaged 4.7 kilometres. In a study in western NSW, Sharp et al. (2001) found the mean distance to nearest neighbour between Wedge-tailed Eagle nests was in the order of 1 pair per 3–9 km². They noted this was considerably higher than that noted in other semi-arid zone studies (~1 pair per 40–48 km²).

Using a conservative mean Wedge-tailed Eagle territory size of 30 km², the average diameter of a territory would be slightly greater than 12 kilometres. As a consequence, we have based the modelling exercise for Wedge-tailed Eagles on the assumption that the 26 kilometre linear array of the proposed wind farm may intersect with three territories, occupied by six adult birds.

Cherriman (2013) reported that breeding productivity (number of chicks fledged) was 0.73 young per pair, across 15 occupied territory-years. Debus et al. (2007) recorded very similar results with 10 young produced in 12 pair-years, equating to 0.8 young fledged per pair per year. On the basis of those studies, we have conservatively assumed that, on average, three pairs will be accompanied by a total of three flying juveniles, bringing the average site-population of Wedge-tailed Eagles to a total of 9. Hence we have modelled for this number of birds as being at potential risk of collision.

During field investigations of the site, Biosis staff documented one instance each in which three, four and five Wedge-tailed Eagles were observed simultaneously.

Nankeen Kestrel

Near Armidale, NSW, one pair of Nankeen Kestrels occupied at least 200 hectares (2 km²) (Genelly 1978) and active nests were recorded approximately 1 kilometre apart (Baker-Gabb 1985). Near Mildura, Victoria, 12 pairs were documented from an area with a 10 kilometre radius (i.e. approx. 314 km²), and 25 nests averaged

1-3.6 kilometre apart equating to 1 pair per 5.4 km² (Baker-Gabb 1984). At Millewa, Victoria, Campbell (1986) reported an average of 1 active pair per 5.3 km² [all references in Marchant and Higgins (1993)].

Using a conservative mean Nankeen Kestrel territory size of 5.3 km², the average diameter of a territory would be approximately 2.6 kilometre . As a consequence, we have based the modelling exercise for Nankeen Kestrels on the assumption that the 26 kilometre linear array of the proposed wind farm may intersect with 10 territories, occupied by 20 adult birds.

Baker-Gabb (1984) reported a mean number of 1.3 fledglings per territorial pair. On the basis of that study, we have assumed that, on average, 10 pairs will be accompanied by a total of 13 flying juveniles, bringing the average site-population of Nankeen Kestrels to a total of 33 and we have modelled for this number of birds as being at potential risk of collision.

4. Model results

4.1 Raptors

Collision risk model results for Nankeen Kestrel and Wedge-tailed Eagle are shown in Table 2. As discussed earlier, informed assumptions have been made for the possible site-population sizes of these two species, and results for them are provided here expressed as projected numbers of annual average collisions. Results are provided for four potential avoidance rates.

Experience with these two species at wind energy facilities in south-eastern Australia demonstrates that both Nankeen Kestrels and Wedge-tailed Eagles collide with wind turbines (Moloney et al. 2019). For Wedge-tailed Eagles there is some published empirical data (Smales et al. 2013) and more recent unpublished data for actual mortalities available to validate the outputs of the Biosis collision risk model. That evidence suggests that the model's projections accurately equate to avoidance capacity of between 0.90 and 0.95. As with any forward-projection modelling, the accuracy of the results presented here for the proposed Hills of Gold Wind Farm, will depend upon the precision of all assumptions used for the modelling process.

Table 2 Collision risk model results for 70 x Vestas 5.6 turbines for two raptors at Hills of Gold Wind Farm site

Common name	Scientific name	Dynamic rotor avoidance rate			
		0.90	0.95	0.98	0.99
Nankeen Kestrel	<i>Falco cenchroides</i>	0.36	0.20	0.10	0.07
Wedge-tailed Eagle	<i>Aquila audax</i>	5.86	3.71	1.77	0.98

4.2 Other species

Collision risk model results for all 18 species of birds that were documented flying within rotor-swept height of the Vestas 5.6 turbines as proposed for Hills of Gold Wind Farm, are shown in Table 3. As discussed earlier in the report, information about the possible site-population sizes of 16 of these species is not available, and results for them are provided here expressed as projected numbers of annual flights that may be at risk of turbine collisions. Results are provided for four potential avoidance rates.

Fewer than 20 flights were recorded during the total of all point count field observations for the species shaded grey in Table 3. In cases such as these where the sample size of flights is low, it is possible that the model results may be less reliable than they are for species that were recorded more frequently. If the low number of observations for those species indicates that they occur relatively infrequently, or make few flights, that may still indicate that their risk is relatively low. We include them here for completeness, but under the caveat that the model's estimates for them may be less certain than the results for species with a greater number of records.

Table 3 Results for 70 x Vestas 5.6 turbines for 18 species of birds recorded within RSH at Hills of Gold Wind Farm site

Common name	Scientific name	Dynamic rotor avoidance rate			
		0.90	0.95	0.98	0.99
Brown Goshawk	<i>Accipiter fasciatus</i>	0.24	0.12	0.05	0.03
Galah	<i>Cacatua roseicapilla</i>	0.34	0.17	0.07	0.04
Nankeen Kestrel	<i>Falco cenchroides</i>	0.36	0.20	0.10	0.07
White-browed Treecreeper	<i>Climacteris affinis</i>	1.08	0.54	0.22	0.11
Australian Magpie	<i>Gymnorhina tibicen</i>	1.07	0.61	0.33	0.24
Yellow-tailed Black-Cockatoo	<i>Calyptorhynchus funereus</i>	3.27	1.64	0.67	0.34
Laughing Kookaburra	<i>Dacelo novaeguineae</i>	0.73	0.38	0.16	0.09
Red Wattlebird	<i>Anthochaera carunculata</i>	0.59	0.38	0.25	0.21
Spotted Pardalote	<i>Pardalotus punctatus</i>	1.03	0.52	0.21	0.11
Common Starling	<i>Sturnus vulgaris</i>	0.30	0.16	0.08	0.05
Rainbow Lorikeet	<i>Trichoglossus haematodus</i>	0.52	0.26	0.11	0.06
Crimson Rosella	<i>Platycercus elegans</i>	0.62	0.35	0.20	0.14
Little Wattlebird	<i>Anthochaera chrysoptera</i>	1.70	0.87	0.36	0.20
Pied Currawong	<i>Strepera graculina</i>	1.99	1.02	0.43	0.24
White-breasted Woodswallow	<i>Artamus leucorhynchus</i>	1.67	0.84	0.35	0.18
Australian Raven	<i>Corvus coronoides</i>	4.02	2.04	0.86	0.46
Sulphur-crested Cockatoo	<i>Cacatua galerita</i>	3.95	2.01	0.84	0.45
Wedge-tailed Eagle	<i>Aquila audax</i>	9.46	4.78	1.97	1.03

5. Conclusion

A total of 51 species of birds were recorded during investigations of the Hills of Gold site. While all of them may have capacity to fly at rotor-swept height, 18 species were recorded doing so and the Biosis turbine collision risk model was run for them. The modelling was undertaken for 70 x Vestas 5.6 turbines after preliminary assessment suggested that this turbine is likely to represent a greater collision risk than two other types of turbines under consideration for operational use by the project.

None of the species involved are listed within any category of threat status under New South Wales or Commonwealth legislation.

Informed assumptions were able to be developed and employed for the potential site-population sizes of Nankeen Kestrels and Wedge-tailed Eagles and this permitted the model to provide projections expressed as average numbers of potential collisions per annum for those two species. Depending upon avoidance capacity and all other assumptions used for Nankeen Kestrels the model returned a likely range of between 0.36 and 0.07 collisions for that species per annum. Under the same caveats for Wedge-tailed Eagles, the likely range was between 5.86 and 0.98 collisions per annum. Empirical evidence from some wind farms in south-eastern Australia suggest that avoidance capacity for this species at those sites has been between 0.90 and 0.95.

For 16 other species, of birds collision risk modelling provided results expressed as average numbers of their flights that might be at risk of turbine collisions. For nine of those species the model indicates that they might make between one and four flights per annum that would be at some risk of collision assuming their collision avoidance capacity was no greater than 0.90. This is considered to be a very low avoidance rate and most birds appear to avoid turbine collisions at a significantly higher rate than that. On the basis of the bird utilization data collected for these birds at the site, and other assumptions entailed in the modelling, it is reasonable to conclude that the number of actual collisions that might occur per annum for all of these species would be lower than the number of their flights-at-risk.

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Appendix 1



A Description of the Biosis Model to Assess Risk of Bird Collisions With Wind Turbines

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ABSTRACT We describe the model of Biosis Propriety Limited for quantifying potential risk to birds of collisions with wind turbines. The description follows the sequence of the model's processes from input parameters, through modules of the model itself. Aspects of the model that differentiate it from similar models are the primary focus of the description. These include its capacity to evaluate risk for multi-directional flights by its calculation of a mean presented area of a turbine; its use of bird flight data to determine annual flux of movements; a mathematical solution to a typical number of turbines that might be encountered in a given bird flight; capacity to assess wind-farm configurations ranging from turbines scattered in the landscape to linear rows of turbines; and the option of assigning different avoidance rates to structural elements of turbines that pose more or less risk. We also integrate estimates of the population of birds at risk with data for numbers of their flights to predict a number of individual birds that are at risk of collision. Our model has been widely applied in assessments of potential wind-energy developments in Australia. We provide a case history of the model's application to 2 eagle species and its performance relative to empirical experience of collisions by those species. © 2013 The Wildlife Society.

KEY WORDS bird, collision, model, risk, turbine, wind energy.

A number of mathematical models have been developed for the purposes of either describing the interaction of a bird with a wind turbine or to predict the risks of bird collisions with turbines (Tucker 1996a, b; Podolsky 2003, 2005; Bolker et al. 2006; Band et al. 2007). Tucker (1996a, b) and Band et al. (2007) detailed their models in the peer-reviewed literature. The collision risk model developed by Biosis Propriety Limited has been widely used to assess wind-energy developments in Australia since 2002, but it has not previously been described in detail. Given high levels of interest in effects of wind turbines on fauna, we believe it is important for the model to be accessible.

Our model provides a predicted number of collisions between turbines and a local or migrating population of birds. It has the potential to be modified to accommodate Monte-Carlo simulation, although at its core it uses a deterministic approach. It is modular by design, and allows various customizations, depending upon the unique configuration of the wind facility and characteristics of the taxa modeled.

The initial calculation involves species-specific parameters for speed and size of birds and specifications of the turbine, including its dimensions and rotational speed of its blades. Using these parameters, we derive the mean area of turbine

presented to a bird in flight. This allows the model to accommodate flight approaches from any potential direction. Alternatively, unidirectional flights can be modeled by using the relevant turbine surface area presented to birds approaching from a given direction.

Data for bird flights are collected at the wind-farm site according to a specific and consistent field methodology. These data are used to determine the flux (density) of bird flights. When combined with turbine specifications, this yields the probability of collision during a single flight-turbine interaction. The density flux approach has not been used for this application previously.

The number of movements at risk of collision with one turbine is then scaled according to a typical number of turbines that a bird might encounter in a given flight. This is further refined by a metric for the capacity of the particular species to avoid collisions. Where a population census or estimate is available for the number of birds that may be at risk, a further deduction is used to attribute the number of flights-at-risk to individuals, and hence provide a final model output as the number of individuals at risk of collisions. The ability to transform from flights-at-risk to individuals-at-risk has been uniquely developed and applied as a routine component of our model.

DESCRIPTION OF THE MODEL

The model requires data for input parameters and, using these, functions in a sequence of modules (Fig. 1).

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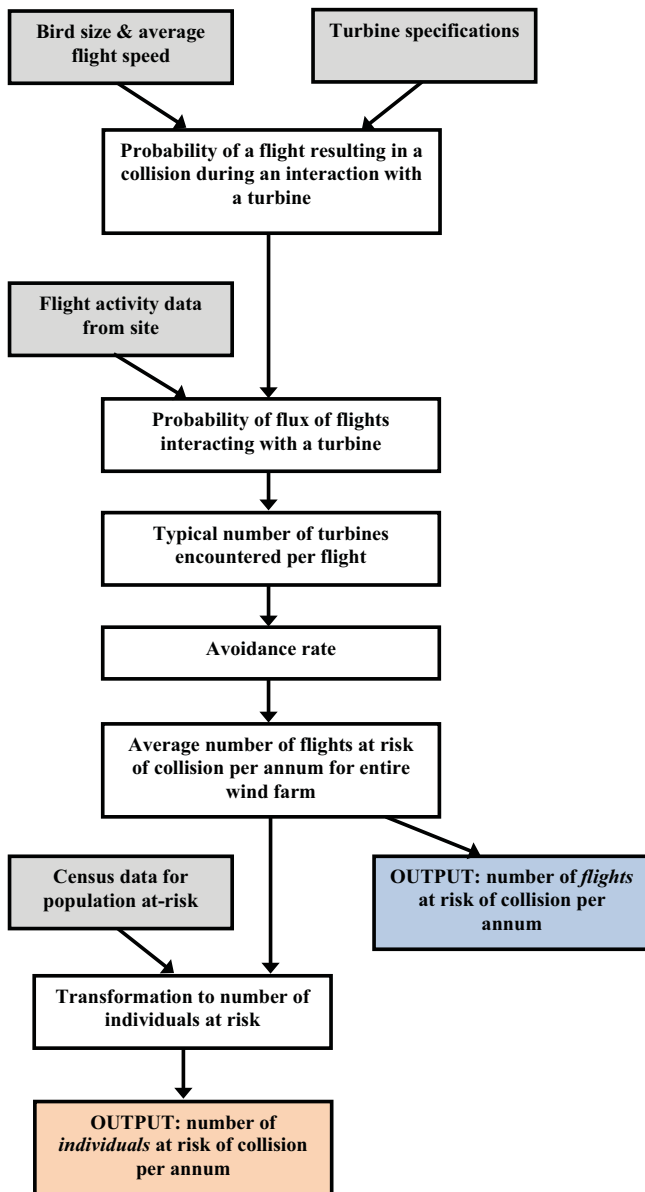


Figure 1. Overview of the collision risk model that quantifies risk to birds of colliding with wind turbines, showing input parameters (gray boxes), modules, and sequence.

Model Inputs

Turbine parameters.—The primary risk faced by a flying bird, whether it may strike or be struck by a turbine, is that the machine presents a potential obstacle in its path. Ultimately this equates to the surface area of the turbine presented to the bird from whatever its angle of approach. Other models, such as probably Band et al. (2007), use individualistic representations of birds. Our model uses a projection of the presented area onto all possible flight angles. For this reason, multiple dimensions of turbine components and rotor speed for the particular type of turbine are used as input values to the risk model. Turbine specifications are as provided by the machine's manufacturer.

The modeled wind turbine consists of 2 fundamental components representing potentially different risks. We refer

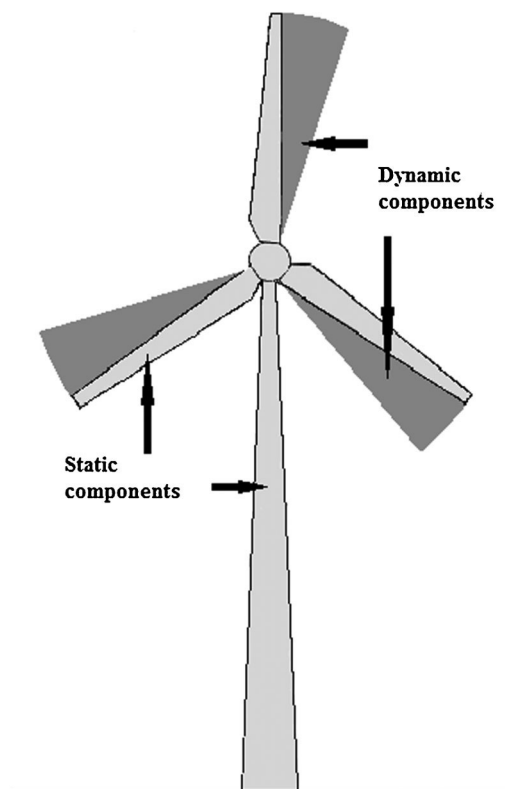


Figure 2. Schematic indication of the static and dynamic components of a wind turbine that may be encountered by a flying bird. The dynamic component is the area swept by rotor blades during the time that a bird of a particular species would take to pass through the rotor-swept zone.

to these as the static and dynamic components (Fig. 2). The static areas of a turbine include all surfaces of the entire machine comprising a tower, which in current turbines is a simple taper with known base and top diameters; a rectangular nacelle housing the generator; a hemi-spherical hub; and rotor blades that taper in 2 planes. The dynamic component is the area swept by the leading edges of rotor blades during the time that a bird would take to pass through the rotor-swept zone.

Size and flight speed of birds.—For each taxon, the model requires values for the total length of the bird in flight, from bill tip to tip of the tail or outstretched legs, and the average speed of the species' flights. We obtained bird lengths either from museum specimens or from standard ornithological texts.

Accurate determinations of bird flight speeds can be complex and difficult to obtain (Videler 2005, Pennycuik 2008) and published data are not available for most species. However, published radar studies (e.g., Bruderer 1995, Bruderer and Boldt 2001) provide ranges of flight speeds for a variety of species, including congenics with similar morphologies and ecological traits to a number of species we have assessed. Use of radar to collect bird flight data at the wind-farm site may provide flight speeds for species of interest. We consider that average ground speed (as opposed to air speed) is appropriate for modeling of multidirectional movements of birds.

Bird flight data.—The model requires data from the wind-farm site for the number of flights made by species of interest within a measured time and volume of airspace. Movement data may be obtained from fixed-time point counts using a methodology adapted from Reynolds et al. (1980), incorporating an effective detection range (Buckland et al. 1993). It may be collected by human observers or by using horizontal and vertical radar combined with call recording or visual species identification (e.g., Gauthreaux and Belser 2003, Desholm et al. 2006). Data represent the number of flights that birds make within a cylinder of airspace that is centered horizontally on the observer and the height of which is the maximum reached by rotor blades of the turbines. The data collection regime is designed with the aim of providing a representative sample of flight activity across the local range of diel, seasonal, and other environmental variables.

Model Modules

Probability of a single flight interacting with a turbine.—In some situations, such as during highly directional migratory passage, the presented area of turbines is determined from the angle of the birds' flight relative to the compass orientation of turbines. However, for the great majority of species (including temporary or permanent residents at an on-shore wind farm) this does not apply, and flights can be expected to approach turbines from any direction. For this situation, all dimensions of the turbine contribute to the area with which a flying bird might collide and the model uses a simple integration to determine a mean presented area. This represents a substantial advance over other collision risk models that depend on the assumption of a specific angle of approach as a bird encounters a turbine (e.g., Tucker 1996a, b; Bolker et al. 2006; Band et al. 2007).

We calculate the area presented by the static components of a turbine using a conservative assumption that none of them overlap or obscure any others. The area of each component is calculated individually, and these are then summed to determine a total static area for the turbine. Static areas are calculated from the simple length \times width dimensions of all components visible by line of sight. These are then projected onto an arbitrary approach direction (effectively scaling by the cosine of the approach angle). For example, viewed directly from one side, only the side panel of the nacelle is visible. However, approached from 45° to the turbine, both the front and side panels are visible, and are thus scaled by $\cos(45) \approx 1/\sqrt{2}$ to match that particular angle of view.

We calculate the dynamic area, swept during the movement of blades, from the dimensions of the stationary blades and the distance they travel at their average speed during the time taken by a bird to fly through the rotor-swept area. We assume that all flights involve forward movement, so the swept-area is derived from the length and speed of the particular species of bird, in combination with the thickness of the sweeping blade.

Each rotor blade is tapered in 2 planes. Thus the thickness of the blades, used to determine the time taken for a bird to cross through the swept area, is actually a function of the

point in the rotor radius at which an individual bird's flight intersects the swept area. This presents a complication that we overcome by defining an effective blade, which is a simple rectangular cross-section that sweeps out precisely the same volume of space as the physical blade. In doing so, we calculate a constant thickness of blade that accounts for the fact that the thinner tips actually sweep far more space than the thicker base of the blade. This ensures also that our flux calculation is not compromised by introduction of a spatial variation at odds with other aspects of the model.

A further input parameter is the percentage of time per annum when rotors are not turning due to inappropriate wind speeds and routine turbine maintenance. Prior to commissioning of a wind farm, wind speed data are usually gathered and the expected percentage of downtime due to inappropriate wind speeds is determined. During downtime periods the rotor simply stops turning; and so risks associated with dynamic components only are reduced by this percentage of time, while all static components of the turbine remain as potential obstacles to flying birds.

Combining all presented areas of the turbine.—Modeling for multidirectional bird movements requires no dependence on approach angles nor on complexities of interactions between flight direction and wind direction. We thus reduce the turbine to its mean presented area. This is solved by the equation

$$\frac{1}{\pi} \int_0^{\pi} A(\theta) d\theta$$

where A is the presented area of the turbine as a function of approach angle θ . We solve this numerically using a trapezoidal integrator (Press et al. 1992).

Probability of multiple flights interacting with a turbine.—Because counts of bird flights have been made across the wind-farm site and there is no obligatory relationship between point-count locations and particular sites proposed for turbines, we combine the data collected from all point counts. This provides a measure of flight activity, which is assumed to be constant across the site. Thus the field data reduce to a single ratio value for the subject species, which is the sum of all flights documented during all counts divided by the total time of observations. This equates to a maximum likelihood estimation of the mean of an assumed Poisson distribution.

To calculate a number of flights at risk of collision, we first reduce documented bird movements (M) to a measure of flux (F) using the equation

$$F = \frac{M}{T_{\text{obs}} A_{\text{obs}}}$$

where T_{obs} is the combined total time of all point counts and A_{obs} is the area of the vertical plane dissecting the observation cylinder. This flux is a measure of bird movements per time per square meter of vertical airspace. The third dimension, volume of airspace, is redundant (or tacit) due to the

assumption that, unless involved in a collision, flight paths do not end arbitrarily in space.

We next multiply activity measure by the number of minutes in which the species is active during the 24-hour diel period, T , and the total presented area of the turbine, A . For year-round resident species, the “active minutes” are calculated for the entire year, while for seasonal or migratory species, they are calculated for the portion of the year that the species is present at the site. This then gives a measure of risk to the bird movements, $M_{\text{risk}} = \text{FTA}$.

Because the flight data are a measure of movements by the species in question and do not discriminate the number of individuals making the movements, the measure (M_{risk}) quantifies the total movements-at-risk for the species and does not reflect risk to individual birds.

To determine a risk rate from total of recorded movements-at-risk, it is necessary to extrapolate to a total number of expected bird movements per annum, M_{yearly} . We calculate this from the flight data, extrapolating the movements to a yearly total through the equation

$$M_{\text{yearly}} = M \frac{T_{\text{yearly}}}{T_{\text{obs}}}$$

We then deduce a probability of flights at risk of collision as $M_{\text{risk}}/M_{\text{yearly}}$. Note that T_{year} is the total time in a year, and not the diel activity period of the species, which has already been factored into the calculation of movements at risk.

The resultant value is now a probability of flights being at risk of collision with a single turbine. To this point, no account is taken of the bird’s own ability to avert a collision. This is modified later through use of an avoidance factor.

Estimating number of turbines encountered per flight.—Every turbine is presumed to represent some risk for birds, so the total number of turbines proposed for the wind farm is an input to the model. Turbine layout of modern wind farms is primarily determined by the wind resource and turbines are micro-sited accordingly. Consequently, the machines are usually scattered on the landscape. Older wind farms had turbines arrayed in rows, and occasional modern facilities may be linear where they follow a single topographic feature.

To account for the number of turbines with which a single flight might interact, it would be necessary either to know precisely the route of every flight or to make informed assumptions about flight paths. The manner in which turbines are arrayed in the landscape is important to ascertain a typical number of turbines that a bird might encounter in a given flight. This number differs according to whether turbines are in a scattered array or a single row, and these require different calculations.

For a row of turbines, the likely number of encounters can be visualized by considering a row of N turbines in plan view and a flight path at angle Φ to the row. A flight directly along the line of turbines (Φ') will interact with all N turbines. As the angle of flight relative to the row increases toward 90° , flight paths have potential to interact with fewer turbines until an angle (Φ'') is reached at which the path has potential to interact with a maximum of one turbine.

For a single row of turbines, we define the piecewise smooth function, which gives the number of turbines for a given angle of crossing with,

$$n_{\text{interaction}} = \begin{cases} N, & \text{if } \theta \leq \phi' \\ \cot(\theta), & \text{if } \phi' < \theta \leq \phi'' \\ 1, & \text{if } \phi'' < \theta \leq \frac{\pi}{2} \end{cases}$$

This gives us an expected number of interactions as

$$\langle n_{\text{interaction}} \rangle = \frac{2}{\pi} \left[N \arctan\left(\frac{1}{N}\right) + \frac{\pi}{4} - \ln\left(\sqrt{2} \sin\left(\arctan\left(\frac{1}{N}\right)\right)\right) \right]$$

For scattered turbine arrays it is not realistic to assume that a bird will encounter all turbines in the wind farm in a given flight. We assume each flight has potential to cross between any 2 points on the outer edges of the farm. Given the size of most on-shore wind farms, this is a reasonable assumption for typical species of concern, such as raptors. When multiple flight paths are drawn randomly across the plan view of a wind farm, some paths may be circuitous and have potential to encounter many turbines, while others will pass through a small portion of the site and have potential to encounter relatively few turbines.

To deduce an average number of turbines likely to be encountered by any flight we use a topological, non-affine mapping technique. This spatial transformation can be illustrated as follows: if we were to throw a lasso around the perimeter of the site and shorten it to its minimum, we would find that all the turbines had collected in a circle. A straight flight path through this “lassoed” site is mathematically equivalent to a random walk across the unconstrained layout. The average of all flight paths crossing the center of this remapped farm will intersect with \sqrt{N} turbines (where N is the total no. of turbines in the wind farm). This value is used in the model for the number of turbines that might be encountered per flight within a scattered turbine array.

For arrays that are neither entirely scattered nor linear, the model employs a simple weighted average of the values for fully scattered and entirely linear arrays.

Application of turbine avoidance capacity.—Birds have substantial ability to avoid obstacles; therefore, it is necessary to incorporate this capacity into the model. In common with other workers (Percival et al. 1999), we use “avoidance” in specific reference to behavior on the part of a bird that averts a potential collision with a turbine. The “avoidance rate” equates to the proportion of flights that might otherwise have involved interaction with a turbine but where the bird alters course and the flight does not result in a collision. For the purposes of the model it is of no consequence whether or not this is a result of a cognitive response by the bird to the presence of the turbine.

Turbine avoidance remains little-studied for any species, and empirical information about actual avoidance can be obtained for a given site only by studying the responses of birds in the presence of operational turbines (Chamberlain et al. 2006). One recent investigation has compared flight behaviors of 2 species of eagles in the presence of turbines at

2 operating wind farms with their behaviors at a site without turbines (Hull and Muir 2013).

Avoidance rate is incorporated into the model by scaling the movements at risk by $(1 - v)$, where v is a measure of the bird's ability to avoid objects. In this scenario, $v = 0$ corresponds to a blind, non-responsive projectile, and $v = 1$ represents a perfectly responsive bird able to avoid any object.

A novel feature of our model is its capacity to apply different avoidance values to the static and dynamic portions of a turbine. As noted by Martin (2011), birds are known to collide with both stationary and moving parts of turbines. This aspect of our model allows for differences in capacity of birds to detect and avoid the large, static components of modern turbines relative to their capacity to detect and avoid the small and fast-moving leading edges of rotor blades.

Size of population at risk.—When information about the size of the population at-risk is available, this can be factored directly into our model to provide results in the form of an expected number of individuals at risk of collision per annum. This is an important consideration because an input measured in terms of bird movements cannot provide an output in terms of individual birds. This aspect appears to have been largely overlooked by other workers, although Chamberlain et al. (2006) alluded to the use of a number of flights only, without incorporation of the number of individuals, as a potential issue in evaluation of collision estimates provided by the Band model (Band et al. 2007).

To deduce a predicted number of individual birds that are at risk of collision, a valid estimate is required of the number of individuals that may interact with turbines at the wind farm in the course of a year. If it is not feasible to obtain this for a species, then the output of the collision risk model will necessarily be the number of flights-at-risk per annum. Although this metric is not predictive of the number of individuals that might collide, it permits risk to be compared for various designs of a wind farm or between one facility and another. In rare cases, such as where there is a single migration passage through the site per annum, the number of movements may equate with the number of individual birds that are at risk. The great majority of risk modeling we have undertaken has been for raptors that are year-round residents. Due to their territoriality and relatively low densities, our studies at wind-farm sites have been able to ascertain the number of individuals using a site per annum, including both resident adults and juveniles, with a high level of confidence. For some other species, such as cranes (Gruidae), we have undertaken home-range studies to determine numbers present during the breeding season, and we have obtained local census data to estimate numbers of individuals that might encounter turbines during non-breeding seasons.

Given a population estimate, the number of flights at risk is attributed equally to the relevant number of individuals through the simple relation $M_{\text{individuals}} = \text{Yearly Movements} / \text{Population}$. We can then attribute individual mortality through

$$\text{mortality} = \text{Population} \left(1 - \frac{\text{Movements At Risk}}{\text{Yearly Movements}} \right)^{M_{\text{individuals}}}$$

MODEL VALIDATION

The model we describe here has been used to assess potential turbine collision risk for numerous species of birds for 23 commercial-scale wind farms proposed in Australia and one in Fiji. Eleven of these facilities have subsequently been built and are now operational. The model's projections have been used by regulatory authorities in determination of approval or modification to wind-farm designs for a range of species of concern. These include taxa as diverse as the orange-bellied parrot (*Neophema chrysogaster*), wedge-tailed eagle (*Aquila audax*), brolga (*Grus rubicunda*), and the large and readily observable Pacific fruit-bat (*Pteropus tonganus*) in Fiji.

The model's performance can be validated only when it can be compared with post-construction mortality data that are sufficient to permit calculation of an actual annual mortality rate and a 95% confidence interval for that rate. Conditions of regulatory approval for most wind farms that have been built to-date in Australia have varied considerably between state jurisdictions and over time. Generally they have not required rigorous investigation or public reporting of avian collisions that occur during operation. We have thus had limited opportunity to validate our model against empirical information for actual collisions. However, where these are available, we can compare the model's predicted average estimates with the measured confidence interval for actual mortalities to assess its predictive capacity. We present one such case study below.

Comparing the Model's Predictions With Empirical Data—A Case History

Substantial investigations have been undertaken at Bluff Point and Studland Bay wind farms in northwestern Tasmania entailing a number of studies of wedge-tailed eagle and white-bellied sea-eagle (*Haliaeetus leucogaster*). These have included utilization surveys designed to measure eagle activity before and after development of the wind farm; collision monitoring; eagle breeding success; eagle behaviors and movements relative to turbines and observers; and investigations and trials aimed at reduction of collisions (Hull et al. 2013). Commissioning of turbines began at Bluff Point Wind Farm in 2002 and at Studland Bay Wind Farm in 2007. Bluff Point Wind Farm consisted of 37 Vestas V66 turbines in a scattered array on an area of 1,524 ha. Studland Bay Wind Farm was situated 3 km south of Bluff Point and comprised 25 Vesta V90 turbines in a scattered array over an area of 1,410 ha. Both wind farms were close to the coast of northwestern Tasmania and resident white-bellied sea-eagles and Tasmanian subspecies of wedge-tailed eagle (*A. a. fleayi*) occurred at both sites.

Monitoring Eagle Flights

Movement data for both species were collected during point counts at Bluff Point Wind Farm site in 3 years prior to construction of turbines and in 4 years after they commenced operating. At Studland Bay, they were collected in 6 years prior to turbine construction and in 3 years after turbines commenced operation. As prescribed by regulatory authorities, point counts were undertaken in the austral autumn and spring. Ten replicate point counts were made in each season

at 18 locations per wind farm. There were 545 point counts undertaken at Bluff Point between 1999 and 2007 and 854 point counts at Studland Bay between 1999 and 2009.

Collision Risk Model Results

We used the model to estimate risk based on movement data collected prior to construction for populations of 6 wedge-tailed eagles and 4 white-bellied sea-eagles at-risk per annum at each of the 2 wind farms.

State regulatory authorities have required that the collision risk model be re-run with the accumulated sum of eagle movement data obtained during the entire period of both pre-construction and operation of the 2 wind farms spanning the period from 1999 to 2009 (Table 1). We modeled static avoidance rate at 99% in all cases.

Documented Eagle Collisions

Carcass monitoring surveys were conducted at the Bluff Point and Studland Bay wind farms since they commenced operating. Fences to exclude mammalian scavengers were maintained at 27% of turbines across the 2 sites. All turbines, both fenced and unfenced, were searched routinely within a 100-m radius of the tower base. Search frequency was initially informed by trials to determine rates of loss to scavengers and of observers' capacity to detect carcasses. Since 2007, searches were carried out twice weekly during periods that may have represented higher risk to the species (i.e., eagle display period Jun–Aug, inclusive; and eagle fledging period mid-Dec–Feb, inclusive) and fortnightly outside these periods (Hull et al. 2013). Assessment of the extent of undetected eagle collisions (Hydro Tasmania 2012; Hull et al. 2013) concluded that it is unlikely that significant numbers of eagle carcasses were missed because they are conspicuous; the search zone around turbines was adequate to detect eagle carcasses where they will fall after colliding with turbines (Hull and Muir 2010); personnel on site had capacity to detect carcasses that may have been moved from the formal search zones; eagle carcasses in vegetation were found not to decompose readily and, even when scavenged, remains were identifiable; avian scavengers did not remove all evidence of carcasses and, although mammalian scavengers could remove carcasses, this was controlled at the subset of fenced turbines; survey intensity was informed by predetermined scavenger removal rates; and, although a small number of eagles survived collision with a turbine, in all documented cases such birds were unable to fly and are likely to have been detected because

both scavenger exclusion and farm fences prevented them from leaving the site.

Comparison of Collision Risk Model Estimates With Actual Mortality Rates

Given constraints of statistically low collision numbers, the model's estimates of annual collisions, based on the combined total of movement data from pre-construction and operation of the 2 wind farms from 1999 until 2009 (Table 1), compare well with actual mortality of the 2 eagle species at both wind farms (Table 2). The model's estimate of the number of wedge-tailed eagle collisions per annum at Bluff Point at a 95% avoidance rate was 1.5, which is the same as the mean number of documented mortalities per annum. Estimates provided for this case by model iterations for 90% and 95% avoidance rates fell within the 95% confidence interval of measured mortality rates. The model's estimates for number of collisions at a 95% avoidance rate for white-bellied sea-eagles at Bluff Point (0.5) and for wedge-tailed eagles at Studland Bay (1.1; Table 1) also closely approximated the mean numbers of documented mortalities per annum for the 2 species (0.4 and 1.0, respectively; Table 2). For those cases, the model's estimates for the range of avoidance rates between 90% and 99% fell within the 95% confidence interval of measured mortality rates. No white-bellied sea-eagle collisions have yet been reported from Studland Bay so, to date, the model's estimates are higher than actual experience for that species there.

MANAGEMENT IMPLICATIONS

We consider that there are 2 different, although not mutually exclusive, applications for modeling of bird collision risks at prospective wind farms. These are to provide projections of long-term effects of a particular wind-energy facility on key bird species; and to determine relative risks for key species that are associated with different wind-farm sites, different portions of large wind farms, and different types of turbines and/or turbine configurations.

In many respects, we consider the latter use of collision risk modeling is the most important contribution it offers. This application provides a tool for planning of wind farms to avoid, reduce, or mitigate potential risks to birds. The model we describe here has now been used in such an iterative manner for a number of prospective sites to evaluate relative risks to key species posed by different types, sizes, numbers, and layouts of turbines.

The integration in our model of data for numbers of bird flights with numbers of birds in the population at-risk is key to the accurate prediction of potential numbers of collisions. This aspect appears not to have been adequately considered previously but has real implications to the appropriate determination of actual risks posed by a wind farm. Our model's use of bird flight data to determine annual flux of movements; a mathematical solution to the typical number of turbines that might be encountered in a bird flight; capacity to assess wind-farm configurations ranging from turbines scattered in the landscape to linear rows of turbines; and the option of assigning different avoidance rates to components

Table 1. Modeled mean annual turbine collision estimates for 2 eagle species based on movement data collected over the span of pre-construction and operation of 2 wind farms in northwestern Tasmania, Australia, from 1999 to 2009. Estimates are shown for 4 potential dynamic avoidance rates. Static avoidance rate was modeled at 99% in all cases

Dynamic avoidance rate (%)	White-bellied sea-eagle		Wedge-tailed eagle	
	Bluff Point	Studland Bay	Bluff Point	Studland Bay
90	0.9	0.8	2.7	1.9
95	0.5	0.4	1.5	1.1
98	0.2	0.2	0.7	0.5
99	0.1	0.1	0.4	0.3

Table 2. Average annual mortality rate and variance for 2 eagle species based on carcasses detected at 2 wind farms in northwestern Tasmania, Australia

Wind farm	White-bellied sea-eagle		Wedge-tailed eagle	
	Mean annual mortality	Annual variance (95% CI)	Mean annual mortality	Annual variance (95% CI)
Bluff Point 2002–2012	0.4	0.1–1.0	1.5	0.8–2.6
Studland Bay 2007–2012	0.0	0.0–0.7	1.0	0.3–2.2

of turbines that pose more or less risk, all represent refinements designed to improve the predictive capacity of turbine collision risk modeling.

In the cases outlined here, where long-term mortality data sets have permitted validation of the model's collision estimates at given avoidance rates, the two have closely approximated each other. We will seek further opportunities to compare the results of our model with empirical mortality information from operating wind farms, with a view to wider application of the model.

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The model described here is the property of Biosis Propriety Limited, an environmental consultancy business incorporated in Australia. It is used commercially by Biosis Propriety Limited.

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Appendix E Serious and irreversible impact assessments

SAIL assessment for White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland

White Box-Yellow Box-Blakely's Red Gum Grassy Woodland and Derived Native Grassland (Box Gum Woodland) is a CEEC listed under the BC Act and EPBC Act. The vegetation community has an open woodland or open forest structure and often contains a high diversity of groundcover species. Its distribution is strongly associated with more fertile soils on lower elevations across the known range in Queensland, New South Wales and Victoria. The geographic range of Box Gum Woodland is quite broad, ranging from the Queensland border in the north, to the Victorian border in the south. In NSW, occurs in the following bioregions NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highlands, NSW South Western Slopes, South East Corner and Riverina.

Over much of its range, the CEEC has been subject to extensive clearing and modification for agriculture and grazing, so it often occurs as derived native grasslands with no overstorey. It now mostly occurs as fragmented, isolated and modified fragments. Key historical and current threats to this CEEC include clearing for agriculture and urban development.

There are four PCTs within the development footprint that align with the floristic description of this TEC, including:

- PCT 433 - White Box grassy woodland to open woodland on basalt flats and rises in the Liverpool Plains sub-region, Brigalow Belt South Bioregion.
- PCT434 - White Box grass shrub hill woodland on clay to loam soils on volcanic and sedimentary hills in the southern Brigalow Belt South Bioregion
- PCT 492 - Silvertop Stringybark - Yellow Box - Apple Box - Rough-barked Apple shrub grass open forest mainly on southern slopes of the Liverpool Range, Brigalow Belt South Bioregion.
- PCT 599- Blakely's Red Gum - Yellow Box grassy tall woodland on flats and hills in the Brigalow Belt South Bioregion and Nandewar Bioregion.

The CEEC is listed in the BioNet Threatened Biodiversity Data Collection (TBDC) as an entity at risk of SAIL based on the following principles (DPIE 2019):

- Principle 1: an ecological community that is currently observed, estimated, inferred or reasonably suspected to be in a rapid rate of decline.
- Principle 2: an ecological community that is observed, inferred or reasonably suspected to be severely degraded or disturbed.

Given the absence of definitive impact thresholds stated for the community, the potential for a SAIL will be determined by the consent authority, guided by the additional assessment provided below.

Impacts to Box Gum Woodland in the context of this SAIL assessment are mapped on Figure 29.

1. Impacts to the CEEC and the action and measures taken to avoid the direct and indirect impact on the CEEC at risk of an SAIL.

It has been conservatively assumed that all condition states of the above listed PCTs support the required floristic diversity to represent the CEEC. Within the development footprint, there is a total of 8.15 hectares of Box Gum Woodland, which has been considered to meet the listing requirements of both the EPBC Act TEC and the BC Act.

The CEEC was found to occur along the transmission line corridor, mainly to the west of the wind farm, with a small area in the central portion of the development site downslope (and north) of the wind farm itself. Furthermore the CEEC was found to occur at the far northern end of the access track servicing the central

portion of the transmission line, and within two areas requiring upgrades for the transport route including just east of Nundle, and at Devil's Elbow (Figure 9).

The patches of Box Gum Woodland within the development footprint are predominantly located in the Nandewar Bioregion (6.75 hectares), with a small area (1.40 hectares) located in the NSW North Coast Bioregion.

Actions and measures to avoid direct impacts

Throughout the development of the project layout, design decisions have been implemented to avoid impacts to Box Gum Woodland. This has included early biodiversity surveys, prior to development of the preferred corridor.

A preliminary biodiversity assessment (Arup 2019) was undertaken prior to the development of the wind farm layout and the selection of the preferred transmission line corridor. This preliminary assessment highlighted areas of key ecological concern and allowed for avoidance of these areas during the wind farm concept design. During the wind farm layout design, workshops were held between project ecologists, civil engineers and wind modellers to further optimise layout options and ensure impacts to the areas of mapped Box Gum Woodland.

During the design development phase, a wider landscape was reviewed for potential transmission line corridor and seven potential transmission line routes were identified to understand visual impact and willingness to reach land agreements. Desktop and field validated vegetation and habitat maps were reviewed and transmission line options assessed for likely impacts to significant biodiversity features, with a focus on minimising impacts to TECs. A desktop assessment was undertaken to identify the potential impacts to native vegetation communities for each of the seven options using the State Vegetation Type Mapping for the alignments.

Following the review of each of these seven options two preferred routes were selected and an optimisation consisting of a 200 metre corridor was undertaken to adjust the routes to minimise further impact around mapped PCTs and TECs. The transmission line corridor was then further refined to a single option a 60 metre wide impact areas was determined, resulting in impacts to 31.4 hectares of exotic grassland, and 53.5 hectares of native vegetation, of which 3.1 hectares comprises Box Gum Woodland CEEC. This level of impact includes those areas determined as impacted by AECOM (2021) where the vegetation occurs within valleys spanned by the overhead lines (refer section 7.3 for more detail).

Impacts to Box Gum Woodland also occur along the transport route where upgrades are required to allow delivery of turbine components to site. Previous impacts higher condition patches of the CEEC at Devil's Elbow have now been removed from the Project, with the majority of impacts to the CEEC to facilitate site access and component transport to isolated and edged effected patches. One higher condition patch will however be impacted adjacent to Crawney Road.

Overall design refinements undertaken since the exhibited BDAR have resulted in a reduction of assessed impact to Box Gum Woodland CEEC from 13.3 hectares to 8.15 hectares.

Furthermore the current transmission line development footprint considers a conservative 'worst case' clearing footprint for the transmission line easement, assessing complete clearing within the areas of the 60 metre easement not spanned by the overhead wires (AECOM 2021).

Planning for the construction of the transmission line corridor has also been completed as part of the impact assessment. This has included consideration for the placement of access tracks into the main transmission line corridor. Existing road infrastructure and farm tracks have been prioritised to provide the required access during construction and operation. By utilising pre-existing infrastructure, the Project has avoided and minimised impacts to sensitive ecological areas, including areas of Box Gum Woodland.

Actions and measures to avoid indirect impacts

As outlined above, the current transmission line development footprint considers a conservative 'worst case' clearing footprint for the transmission line easement, assessing complete clearing within the areas of the 60 metre easement not spanned by the overhead wires (AECOM 2021). Actual impacts, as determined to occur by the AECOM (2021) (Appendix I) will likely be smaller than, and contained entirely within, the 60 metre wide development footprint. This detailed investigation has shown that impacts to native vegetation can be avoided along the transmission line due to spanning and implementation of a construction methodology that allows vegetation in gullies to be completely avoided, and that these areas will not be subject to ongoing indirect impacts from the overhead wires. Impacts to over 1 kilometre of Box Gum Woodland CEEC (PCT 492) has been confirmed as avoided as a result.

Opportunities to complete revegetation works using species characteristic of Box Gum Woodland have been committed to in Section 8.9 of this BDAR, and the current concept design assumes that a large portion of the transmission line corridor can be subject to revegetation works using native species. Options to plant trees, shrubs and groundcovers will be explored where appropriate safety and operational constraints allow taller vegetation to be established. Where this is not possible, such as locations underneath the transmission line where suitable clearance is required, opportunities to plant groundcover species that occur in Box Gum Woodland will be included in revegetation species mixes. Revegetation of road batters associated with Devil's Elbow upgrade works will ensure only Box Gum Woodland characteristic species are used. With all revegetation seed / plantings sourced to ensure appropriate local provenance.

Additional indirect impacts associated with construction will be managed through preparation and implementation of a Biodiversity Management Plan that will highlight ecologically sensitive areas, including areas Box Gum Woodland CEEC, and ensure these areas are maintained as no-go zones, ensure current condition of retained areas are maintained and improved, and no indirect impacts associated with water, materials storage, access etc occur and ongoing weed control will occur.

2a. Evidence of reduction in geographic distribution, as the current total geographic extent of the TEC in NSW and the estimated reduction in geographic extent of the TEC since 1970 (not including impacts of the proposal). (SAIL Principle 1)

Species and ecological communities that have undergone large reductions or are likely to undergo large reductions in the future are considered to be at greater risk of extinction than those that have undergone or are likely to undergo smaller reductions (NSW TSSC 2018).

To be considered under this principle, the ecological community should have been observed, estimated, inferred, or reasonably suspected to have undergone, or be projected to undergo, a very large reduction in distribution (DPIE 2019).

The Final Determination for the listing of Box Gum Woodland as a CEEC under the BC Act states the community undergone "a very large reduction in geographic distribution" evidenced by the community having been extensively cleared throughout its range, and remnants typically are small, isolated, highly fragmented, that occur in predominantly cleared landscapes and exhibit a highly modified understorey (TSSC 2006). Based on a compilation of available maps depicting the current extent of the community at the national scale, TSSC (2006) estimated that less than 5% of the original distribution remained (NSW TSSC 2020a).

The NSW Threatened Species Scientific Committee (TSSC) Conservation Assessment of Box Gum Woodland (NSW TSSC 2020b) presents indicative estimates of the historical decline in geographic distribution compiled by the Commonwealth TSSC (2006) for state jurisdictions from sub-jurisdictional vegetation maps. The total reduction in NSW is stated as being approximately 93% of the pre-1750 distribution of the community (with 250,729 hectares remaining from a historical area of 3,717,366 hectares). It is noted that there are

uncertainties around the pre-1750 distribution of the community, however the plausible ranges for variants of the community estimated to have been most extensively distributed in NSW suggest that these have almost certainly been reduced to less than 10% of their pre-1750 distribution (NSW TSSC 2020b).

Data is not available on the community's reduction in geographic extent since 1970 (i.e. over the last 50 years), however NSW TSC (2020b) states there is evidence that clearing of Box Gum Woodland CEEC is ongoing and has increased in recent years. During the period 2009 – 2016 it is noted that:

- An average of 395 ha of Grassy Woodland (sensu Keith 2004, of which Box Gm Woodland CEEC is a major component) was lost annually across NSW to agriculture-related activities (cropping, conversion to pasture and thinning)
- A further 155 ha per annum of Grassy Woodland is lost due to infrastructure developments (NSW DPIE 2019).
- Losses due to agriculture rose during the period 2016-2017 to 654 ha (166% of the average over the preceding seven years) and to 1,344 ha (340%) for the period 2017-2018.
- Losses attributable to infrastructure rose to 216 ha (138% of the 2009-2016 average) and 589 ha (378% of the 2009-2016 average), respectively (NSW DPIE 2019).

It should be noted that these figures include other forms of grassy woodland communities, and impacts to Box Gum Woodland CEEC only form a sub-component of these impacted figures, however the data illustrates ongoing pressures and an expected ongoing reduction in geographic extent of the TEC since 1970.

2b. Extent of reduction in ecological function for the TEC using evidence that describes the degree of environmental degradation or disruption to biotic processes, as indicated by i. change in community structure, ii. change in species composition, iii. disruption of ecological processes, iv. invasion and establishment of exotic species, v. degradation of habitat, and vi. fragmentation of habitat. (SAIL Principle 2)

Reduction in ecological function relates to the IUCN principle of “very small population size” which for ecological communities means communities have very high levels of either environmental degradation or disruption of biotic processes, and interactions have an increased risk of failure to sustain their characteristic native species assemblages (Bland et al. 2016).

Ecological communities that are considered to have a very large degree of environmental degradation or disruption of biotic processes or interactions are those with:

- ≥90% extent and severity where the disruption or impacts are measured since 1970.
- ≥80% extent and severity where the disruption or impacts are over a 50-year period, either in the past, future, or any part of the past, present and future (as per (Bland et al. 2016). (DPIE 2019).

Box Gum Woodland CEEC is listed as being Data Deficient for an assessment of environmental degradation of ecological community by NSW TSSC (2020a), however it is listed in the same document as being subject to very large disruption of biotic processes or interactions.

NSW TSSC (2020a) states that Box Gum Woodland CEEC is subject to a number of threatening processes that have negatively impacted upon biotic processes and interactions throughout its range and are likely to cause continuing decline in the future. An almost complete conversion of the community to agricultural production has occurred which invariably includes the removal and/or thinned of the tree canopy resulting in op-down pressures on the ecosystem, with follow-on grazing of domestic stock being the most widespread activity. The impacts of grazing vary depending on the historical grazing regime (timing, intensity, continuity), methods employed to improve pasture (fertilizer application, augmentation with exotic or native species) and the extent of associated impacts on soil structure and biota (soil erosion, compaction).

Grazing has also been shown to lead to a reduction in understorey species diversity and richness due to the loss of native species that are both highly palatable and intolerant of grazing by domestic stock, with many previously widespread species now confined to the least disturbed remnants NSW TSSC (2020a). Shifts in the dominance of pasture species have also been observed as grazing intensity increases and is attributed to differential palatability and resilience to grazing among species, and the reduction of native plant cover by grazing presents opportunities for the invasion of the community by exotic plant species NSW TSSC (2020a).

NSW TSSC (2020b) also states that Box Gum Woodland CEEC is subject to a number of other threatening processes associated with fragmentation, increased soil salinity, inappropriate fire regimes, and reduced recruitment of tree species.

2c. Evidence of restricted geographic distribution, based on the TEC's geographic range in NSW based on i. extent of occurrence, ii. area of occurrence and iii. number of threat-define locations. (SAIL Principle 3)

The geographic distribution of ecological communities is defined by the area of occupancy, sensu (Bland et al. 2016). Ecological communities with a very limited geographic distribution have an area of occupancy of less than or equal to two 10 x 10 km grid cells (200 km²) or an extent of occurrence of ≤1,000 km², sensu (Bland et al. 2016), and one of the following:

- An observed or inferred continuing decline in:
 - A measure of spatial extent appropriate to the ecological community.
 - A measure of environmental quality appropriate to characteristic biota of the ecological community.
 - A measure of disruption to biotic interactions appropriate to the characteristic biota of the ecological community.
- Observed or inferred threatening processes that are likely to cause continuing declines in geographic distribution, environmental quality or biotic interactions within the next 20 years.
- An ecological community that exists at one location (DPIE 2019).

NSW TSSC (2020a) states that the geographic distribution of Box Gum Woodland CEEC is not restricted.

The best estimate of the extent of occurrence (EOO) is 702,800 km², based on a minimum convex polygon enclosing likely occurrences of the community, the method of assessment recommended by (Bland et al. 2016). The best estimate of the area of occupancy (AOO) is 151,100 km² based on 10 x 10 km grid cells (with a minimum of 1% occupied by the community), the scale recommended for assessing AOO by (Bland et al. 2016). The best estimates of EEO and AOO derive from a compilation of maps from multiple sources. Not all of the areas occupied by the community are covered by maps of appropriate scale and accuracy. Therefore, the values for EEO and AOO quoted above may underestimate the true values.

2d. Evidence that the TEC is unlikely to respond to management. (SAIL Principle 4)

This principle encompasses two components, firstly whether there are any particular traits of the community which limits its' response to management, and secondly whether there are any key threatening processes affecting the community which cannot be effectively managed (DPIE 2019).

The Commonwealth TSSC (2006) states that Box Gum Woodland CEEC has suffered a severe decline in extent and condition, and remaining areas are generally small and highly fragmented, and that the key threats to the survival of the ecological community include clearing, grazing and weed invasion. Other threats include salinity, nutrient enrichment, altered fire regimes and the effects of fragmentation.

The priority recovery and threat abatement actions required for the listed ecological community include:

- protection of remnants of the listed ecological community through the development of conservation agreements and covenants;
- protection of remnants from weeds, particularly Coolatai Grass, by preventing soil disturbance in and around remnants, and the speedy eradication of any new invasion;
- avoid the use of fertilisers in or near remnants
- avoid soil disturbance in or near remnants, such as ripping planting lines and road grading;
- in very small derived grassland sites, avoid planting trees as they may reduce the floral diversity through competition for light, nutrients and water;
- planting and other rehabilitation-focused disturbance should focus on the edges of patches, expanding them, rather than within the patches;
- expansion and connection of existing remnants;
- exclusion of continuous grazing from remnants is important, coupled with weed management and control;
- use strategic grazing (incorporating rest at appropriate times) in areas still containing a diverse native understorey;
- burning or slashing if native tussock grasses have built up to a high level, to open inter- tussock spaces for tree seedlings, forbs and shrubs to establish; and,

Further information on recovery actions and conservation measures are detailed in the *National Recovery Plan for White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland* (DECC 2009).

Furthermore the TBDC lists 10 management actions to aid the threat abatement and recovery of the CEEC.

Based on the existence of a national recovery plan providing a guidance as to the management of the CEEC, the community is not considered to be unlikely to respond to management.

3. Where the TBDC indicates data is 'unknown' or 'data deficient' for a TEC, the assessor must record this in the BDAR or BCAR.

Not applicable.

4a. The impact on the geographic extent of the TEC, by estimating the total area of the TEC to be impacted by the proposal.

The majority of the impacted area of Box Gum Woodland TEC within the development footprint have been calculated based on a worst case clearing footprint requirement for the transmission line corridor. This corridor has been calculated for a maximum clearing footprint that is 60 metre wide. This width of corridor is unlikely be required, and is considered to be the maximum width is sufficient to capture any indirect impacts to adjacent retained areas of Box Gum Woodland CEEC.

A summary of the area of direct and indirect impact assessed as total loss, with PCTs and the measured vegetation integrity score is provided in the below table.

PCT and condition class	Area (ha)	Vegetation integrity score*
433 - Moderate	0.01	99.9
434 - Low	0.01	99.9
492 - High	0.01	93.0
492 - Moderate	1.42	93.0

PCT and condition class	Area (ha)	Vegetation integrity score*
492 – Low	0.63	60.3
492 – DNG	1.10	59.9
599 – High	0.81	99.9
599 – Moderate	0.50	99.9
599 – Low	3.66	99.9

* The vegetation integrity scores provided in the above table are artificially inflated due to the use of benchmark values to compensate for a reduced number of vegetation integrity plots in the following vegetation zones 433 – *Moderate*, 434 – *Low*, 492 – *DNG*, 599 – *High*, 599 – *Moderate*, 599 – *Low* (refer Section 4.1 for more detail).

Approximately 67 % of the impacts to Box Gum Woodland (5.4 hectares) as a result of the project will occur on areas of DNG or that have been assessed as occurring in Low condition.

The construction and operation of the transmission line or section of open road built to the required Australian Standards is unlikely to result in changes to any abiotic factors that are critical to the long term survival of Box Gum Woodland in areas adjacent to the development footprint. The construction of the transmission line will not require substantial earthworks that could impact on surface water or groundwater flow patterns. The current design has also allowed for several access tracks into the transmission line corridor to limit the need for substantial earthworks along the length of the alignment to enable construction.

The impacts to Box Gum Woodland associated with the project are confirmed to generally low condition vegetation within a fragmented landscape associated with the transmission line and its access racks, and to one patch of higher condition vegetation off Crawney Road. Neither will result in clearing of vegetation unique to the locality, or to species locally common when compared to elsewhere in the development footprint or broader locality. As such the project is not considered likely to cause a substantial change in the species composition of an occurrence of Box Gum Woodland such that it would continue to decline.

The project is unlikely to result in substantial additional impacts to characteristic and functionally important species beyond these areas that has been assessed as being lost during construction. Opportunities to include characteristic trees, shrubs and groundcover species from the Box Gum Woodland TEC will be included as part of revegetation works following construction.

The proposed development will not directly lead to altered fire regimes in areas of adjacent retained Box Gum Woodland. Recommendations to implement a suitable fire regime to manage biodiversity and fuel load may be required in areas of retained Box Gum Woodland adjacent to the transmission line corridor.

As outlined above impacts to Box Gum Woodland associated with the project are confirmed to be generally low condition vegetation within a fragmented landscape associated with the transmission line and its access tracks, and to one higher condition patch of vegetation adjacent to Crawney Road.

Box Gum Woodland present along the transmission line and in the surrounding landscape largely occurs as fragmented and isolated patches in an over-cleared landscape. The construction and operation of the transmission line will not result in negative ongoing impacts to the CEEC in the landscape, nor will it increase existing pressures associated with fragmentation, isolation and edge effects.

Based on the above additional indirect impacts to Box Gum Woodland are not expected to be substantial or significant as a result of the project and therefore only the direct impacts are considered in the calculation of the project's impacts on the geographic extent of the TEC.

As outlined above, the NSW (2020b) states the current geographic extent of Box Gum Woodland as being estimated as 250,729 hectares. The project will remove a total of 8.15 hectares of Box Gum Woodland, or approximately 0.003% of the extant area of the community.

4b. The extent that the proposed impacts are likely to contribute to further environmental degradation or the disruption of biotic processes of the TEC by i. estimating the size of any remaining, but now isolated, areas of the TEC; including areas of the TEC within 500m of the development footprint or equivalent area for other types of proposals.

Patches of potential Box Gum Woodland CEEC present in the wider landscape surrounding the project area occurs in an already highly fragmented state. GIS was used to determine the range and average size of mapped (Biosis 2021, DPIE 2019, DPIE 2015) occurrences of PCTs representative of Box Gum Woodland CEEC within a 500 metre buffer of the development footprint. The results of which are provided below both for those patches intersected by the development footprint (i.e. subject to vegetation removal) and those patches not intersected by the development footprint (i.e. not directly impact by the project). It should be noted that only the broad components of the development footprint proximal to impacted Box Gum Woodland have been included in these calculations, i.e. large areas of the transport route, where no Box Gum Woodland is impacted, have been excluded.

Mapped areas of known and potential Box Gum Woodland within 500m not directly impacted:

- Size range: <0.001 ha to 305 ha
- Average size: 6.69 ha
- Median: 1.23 ha
- Total no. mapped patches (dissolved Box Gum Woodland PCT polygons): 197

Mapped areas of known and potential Box Gum Woodland within 500m directly impacted:

- Size range: 2.81 ha to 164 ha
- Average size: 29.99 ha
- Median: 6.52 ha
- Total no. mapped patches: 8

It can be seen that there are a large number of mapped known and potential Box Gum Woodland patches within 500 metres of the development footprint, ranging from <0.001 hectares to >300 hectares in area. Patches directly impacted by the project are generally larger with a larger minimum size, average size and median size, than those patches not impacted by the project. This mainly relates to larger patches of PCT 492 mapped within the intact vegetation to the north and east of the subject land. The project is not expected to result in impacts that will substantially increase fragmentation, and/or increase perimeter to area ratios of retained patches of known and potential Box Gum Woodland vegetation based on both the generally large patch sizes of impacted patches and high level of variation in patch sizes within 500 metres of the development footprint.

ii. describing the impacts on connectivity and fragmentation of the remaining areas of TEC measured by:

- ***distance between isolated areas of the TEC, presented as the average***
- ***distance if the remnant is retained AND the average distance if the remnant is removed as proposed, and***
- ***estimated maximum dispersal distance for native flora species characteristic of the TEC, and***

- ***other information relevant to describing the impact on connectivity and fragmentation, such as the area to perimeter ratio for remaining areas of the TEC as a result of the development***

GIS was used to undertake a nearest neighbour analysis of mapped (DPIE 2019, DPIE 2015) occurrences of potential Box Gum Woodland CEEC prior to and post vegetation removal to determine the distance between impacted areas of the CEEC before and after the proposed development. The average and median distance between mapped occurrences of potential Box Gum Woodland CEEC within a 500 m buffer of the impact area, include:

- An average before development separation distance of 75.6 m
- An average after development separation distance of 85.3 m
- A median before development separation distance of 20.11 m
- A median after development separation distance of 15.06 m.

As can be seen from the above calculations the proposed vegetation removal will result in an increase of the average separation distance between patches of known and potential Box Gum Woodland CEEC by approximately 10 meters, but will result in a decrease in the median separation distance by 5 metres. Furthermore it should be noted that the increase in mean separation distance stated above is being driven by the removal of one small patch (<0.1 hectares) of known Box Gum Woodland CEEC vegetation resulting in the increased isolation of another larger patch (approx. 3.29 hectares) of potential Box Gum Woodland CEEC vegetation, such that the patch of the CEEC nearest the larger retained patch post development is over 2.9 kilometres away. If that outlier is excluded from the 'post-clearing' dataset, the development also results in a decrease of mean separation distance from 75.6 metres prior to development to 71.7 metres after development.

These decreases in mean and median separation distance are the result of the very large study area and 500 metre buffer considered in this analysis that supports larger and well separated patches of potential Box Gum Woodland CEEC vegetation the majority of which are not directly impacted by the project, the effect of the relatively narrow vegetation clearing within the development footprint creating new patches separated by between 5 and up to 60 metres, and the generally isolated nature of the majority of the impacted Box Gum Woodland CEEC polygons within the development footprint.

Native flora species characteristic of the CEEC include a range trees, shrubs, grasses, forbs and other groundcover species, the majority of which are dispersed via wind or animal vectors, with some species primary method of dispersal likely to be via non-flying insects such as ants. The expected changes to average separation distance between patches of retained known and potential Box gum Woodland CEEC within 500 m of the impact area are not expected to result in a significant or substantial impediment to the dispersal of native species between these patches, in an already highly fragmented landscape.

It is noted in EPBC Act conservation advice documents that allowances can be made for "breaks" of up to 30 metres between areas of MNES habitat, and that such breaks, which may be the result of watercourses, tracks, paths, roads, etc., do not significantly alter the overall functionality of the ecological community, or habitat (CoA 2020). As such, breaks in connectivity caused by the development footprint, of between 5 and 30 metres are not considered to be substantial in nature.

Along the transmission line and its access tracks the CEEC occurs in an already highly fragmented landscape, with the impacted patches of the CEEC, and the patches mapped in the surrounding landscape largely occurring as isolated patches of vegetation in an over-cleared landscape. Some larger patches of potential CEEC also occur on steeper slopes in more intact vegetation where the CEEC is associated with PCT 488 Silvertop Stringybark - Yellow Box +/- Nortons Box grassy woodland on basalt hills, however these areas are mainly south of the development footprint and on the southern side of the ridgeline. Impacts associated with

the construction and operation of the project's transmission line (and associated access tracks) will not result in fragmentation of any substantial patches of the CEEC, nor will it increase fragmentation in the landscape.

Near Nundle, a small isolated patch of Box Gum Woodland will be impacted by the required transport route upgrade works to allow for wind turbine components to be transported to the site. This impact will occur to a roadside vegetation in low condition, and within a landscape of scattered paddock trees, and retained vegetation near creeks. This impact will not fragment or increase fragmentation of the CEEC in this location.

It is therefore considered that the project will not fragment or increase fragmentation of Box Gum Woodland CEEC.

iii. describing the condition of the TEC according to the vegetation integrity score for the relevant vegetation zone(s) (Section 4.3). The assessor must also include the relevant composition, structure and function condition scores for each vegetation zone.

Box Gum Woodland occurs as four separate PCTs across nine vegetation zones. The relevant composition, structure and function condition scores are provided below.

PCT and condition class	Comp. score	Struc. score	Func. score	VI score	Comments
433 - Moderate	99.7	100	100	99.9	Condition of PCT 433 has been assumed to be benchmark due to a shortfall in collection of BAM plot data. The actual condition of the vegetation is 'moderate' being subject to historical clearing and ongoing agricultural processes.
433 - Low	99.7	100	100	99.9	Condition of PCT 433 has been assumed to be benchmark due to a shortfall in collection of BAM plot data. The actual condition of the vegetation is 'low' being subject to a high level of historical clearing and ongoing agricultural processes.
492 - High	100	80.9	99.5	93.0	Condition of PCT 492 has been assumed where a shortfall in in collection of BAM plot data has occurred. This has increased the attribute and VI scores for each condition state, with the exception of PCT 492 Low. The actual condition of the vegetation ranges from 'high' to 'DNG' depending on the level of historical clearing, ongoing agricultural processes and edge affectedness.
492 - Moderate	98.4	99.7	81.9	93.0	
492 - Low	68.4	60	53.5	60.3	
492 - DNG	75	57.2	50	59.9	
599 - High	99.7	100	100	99.9	Condition of PCT 599 has been assumed to be benchmark due to a shortfall in collection of BAM plot data. The actual condition of the vegetation ranges from 'high' to 'low' depending on the level of historical clearing, ongoing agricultural processes and edge affectedness.
599 - Moderate	99.7	100	100	99.9	
599 - Low	99.7	100	100	99.9	

No specific measures are proposed to contribute to the recovery of Box Gum Woodland in the IBRA subregion, with the exception of the goal of seeking local offsets in the form of Biodiversity Stewardship Sites established in the local area.

It cannot be guaranteed that any Biodiversity Stewardship Sites that are established will directly benefit Box Gum Woodland however rehabilitation of native vegetation in the locality is likely to provide indirect benefits to Box Gum Woodland in the locality.

The project's proposed offset strategy of targeting local properties for the establishment of Biodiversity Stewardship Sites provides potential opportunities for strategic enhancement of local habitat connectivity. Such enhancements could occur along the southern side of the ridgeline between Ben Halls Gap Nature Reserve and Crawney Pass National Park, and over Crawney Mountain to Wallabadah Nature Reserve, linking the three conservation areas. This enhancement of local connectivity can be achieved through the in-perpetuity conservation agreements being pursued over a number of properties along the ridge line, which will improve the biodiversity values on the land and increase habitat connectivity. Connectivity enhancements realised in this strategic location will not only offset direct impacts resulting from the project, but also allow for potential indirect impacts associated with disruption of habitat connectivity to be mitigated against and offset through the establishment of a managed corridor linking local conservation reserves and high-quality habitats.

Areas of the transmission line alignment and road batters surrounding the future bypass of Devil's Elbow will also be subject to revegetation works, with species characteristic of Box Gum Woodland to be included in plant species mixes where appropriate. This will have an indirect benefit to Box Gum Woodland, mainly along the transmission line, through increasing propagative material and decreasing area available for colonisation by existing species in some more degraded areas.

Large Bent-winged Bat SAI assessment

The Large Bent-winged Bat is a small, cave-dwelling insectivorous bat of the *Miniopteridae* family. It is recognised as a subspecies of the Common Bent-winged Bat (also known as the Large Bent-wing Bat) *Miniopterus orianae*. In south-eastern Australia, there are two subspecies of the Common Bent-winged Bat, which are morphologically similar but genetically distinct forming separate maternity colonies (Cardinal & Christidis 2000). The Southern Bent-winged Bat *M. o. bassanii* occurs in south-west Victoria and south-east South Australia; while the Large Bent-winged Bat *M.o.oceanensis* is distributed along the east coast of Australia from Cape York to southern Victoria (Lumsden & Jemison 2015).

The Large Bent-winged Bat is listed as Vulnerable under the NSW BC Act. It is considered a species credit species under the BAM specifically in relation to the species' breeding habitat. This breeding habitat is defined in the Atlas of NSW Wildlife (DPIE 2022) as any:

'Cave, tunnel, mine, culvert or other structure known or suspected to be used for breeding including species records with microhabitat code "IC - in cave;" observation type code "E nest-roost;" with numbers of individuals >500 or from the scientific literature'.

Impacts to these breeding habitat features, and the area within a 100 metre radius buffer around an accurate GPS point location centred on the cave / feature entrance, are what are considered to be potentially serious and irreversible for the species (DPIE 2022).

Amendments to the project design, have since resulted in impacts to these areas that previously occurred within the project footprint being successfully avoided. However, following correspondence with BCS (April 2022), the species is still considered at risk of a SAI as a result of the potential operational impacts of the project.

Internationally and in Australia, microbats are known to collide with the blades of wind turbines. The primary potential cause of impacts to bats, including the Large Bent-winged Bat, as a result of the project is considered to be mortalities due to collisions with turbines.

'Collision' is used here in reference to incidents in which a bat physically strikes, or is struck by, the moving blades of a turbine and to the potential for barotrauma. Barotrauma in bats was described by Baerwald et al.(2008) as the fatal effect on an animal's respiratory tract due to its encountering a rapid change in air pressure close to a moving turbine blade. The effect has since been questioned, as it has been shown to be difficult to diagnose and may have been confused with traumatic injury associated with direct collisions (Rollins et al. 2012).

The ideal objective of an investigation of the impacts to Large Bent-winged Bat as a result of collision would be to numerically quantify potential risk of collisions, in particular whether or how they might affect the overall population of the subspecies. However, unlike the situation for diurnal birds, there is no available technique to accurately record or measure numbers of bats or bat flights for particular species in an environment such as within the subject land. As a consequence there are also limited available methods to model or forecast potential numbers of collisions that might occur, and none relevant to Australian microbat species. Records of bat echolocation calls are the best available method to consistently determine species of bats present and to provide a representation of the variable activity of a given species. But it should be noted that bat calls may not be an accurate surrogate measure of bat flight activity and that detection and recording of echolocation calls are subject to a variety of limitations.

The following SAI assessment has been prepared to address the potential operational impacts to Large Bent-winged Bat as a result of the project.

Table 104 SAI assessment for Large Bent-winged Bat

Information required (BAM Section 9.1.2)	Response
1. The action and measures taken to avoid the direct and indirect impact on the species at risk of an SAI.	<p>Addressed above in Section 4.2 collectively for SAI bat species.</p>
2a. Evidence of rapid decline. (Principle 1)	<p>i. Evidence of rapid decline in the population of the species in NSW in the past 10 years or three generations (whichever is longer), or ii. decline in population of the species in NSW in the past 10 years or three generations (whichever is longer) as indicated by: an index of abundance appropriate to the species; decline in geographic distribution and/or habitat quality; exploitation; effect of introduced species, hybridisation, pathogens, pollutants, competitors or parasites.</p> <p>Species and ecological communities that have undergone large reductions or are likely to undergo large reductions in the future are considered to be at greater risk of extinction than those that have undergone or are likely to undergo smaller reductions (NSW TSSC 2018). To be considered under this principle, the species should have an observed, estimated, inferred, suspected or projected population of $\geq 80\%$ in 10 years or three generations (whichever is longer) (DPIE 2019).</p> <p>Prior to taxonomic revision that resulted in <i>Miniopterus schreibersii</i> being formally split into three sub-species (Cardinal & Christidis 2000), one study estimated established colonies ranging from 15,000 to 200,000 Bent-winged Bats (Dwyer & Hamilton-Smith 1965). More recently, Large Bent-winged Bat has been described as forming colonies of up to 150,000 individuals (DPIE 2021a).</p> <p>There are up to five maternity colonies of Large Bent-winged Bats known in NSW (Mills 2020, NPWS 2011), with three large colonies, that include the , Church Cave, in the Wee Jasper Nature Reserve which been studied for over 60 years. Population estimates have been sporadic and considerably variable. Annual population counts since 2008 saw numbers vary from as low as 16,200 in 2011 to a high of 23,600 in 2016 (Mills 2020). It was noted in 2014 that the population appeared to have been increasing by about 10% each year, and using advanced technology recent counts recorded around 40,000 females and juveniles (Mills 2020). The average annual population change (plus or minus) at the Wee Jasper maternity site is estimated to be 6-7%, apparently correlated with seasonal rainfall (Mills & Pennay 2017). Compared to the closely related Southern Bent-winged Bat which has experienced documented population declines (Lumsden & Jemison 2015), the population of the more common and widely distributed Large Bent-wing Bat appears to be more stable. The current estimate for the Church Cave maternity population is 19,000 which is about 5% lower than the long term average of 20,000, with the population considered to be stable after 14 years of monitoring (D.Mills per. com).</p> <p>The population at Drum Cave (Bungonia) is thought to be similar, though the fluctuation rate is uncertain (D.Mills pers com). Historic data for another maternity cave, Willi Willi cave (near Kempsey) estimated a population of 25,650 individuals (P.D Dwyer 1966a). Current efforts to establish regular monitoring at this site are in the early stages so the population size is currently unknown and whether this number has remained stable over time is yet to be determined. Another maternity colony at Kanagara Boyd NP is known, however little data is</p>

Information required (BAM Section 9.1.2)	Response
	<p>available on the roost numbers (NPWS 2011).</p> <p>Whilst available data documents population counts at only one of the three better studied NSW maternity sites, there is no evidence to suggest that the species has experienced rapid population declines of $\geq 80\%$ in the last 10 years within the scientific literature. Important maternity sites continue to be utilised over a significant time period, and show stable population fluctuations of 6-7% with possible increases being observed since the implementation of regular annual monitoring and advanced technologies to improve the accuracy of population counts.</p> <p>Similarly, a recent viral survey study between Large Bent-winged Bat and the related Southern Bent-winged Bat found that greater numbers of herpesvirus positive Southern Bent-winged Bats may be an indicator of a population labouring under some sort of stress, which was not observed for the comparatively stable population of Large Bent-winged Bat (Holz et al. 2018).</p> <p>Based on the available information, it is unlikely that a decline in the species is being experienced that would trigger consideration of the species under SAIL Principle 1.</p>
<p>2b. Evidence of small population size. (Principle 2)</p>	<p>i. An estimate of the species' current population size in NSW, and ii. An estimate of the decline in the species' population size in NSW in three years or one generation (whichever is longer), and iii. Where such data is available, an estimate of the number of mature individuals in each subpopulation, or the percentage of mature individuals in each subpopulation, or whether the species is likely to undergo extreme fluctuations.</p> <p>Species with small population sizes are highly vulnerable to any event which impacts and further reduces their population size due to the time-lag between developmental impacts and the realisation of ecological benefits from improvements in habitat condition at stewardship offset sites (DPIE 2019).</p> <p>To be considered under this principle a species must have a very small population size which would lead it to be considered critically endangered under the <i>IUCN Red List Categories and Criteria</i> (IUCN 2012). Specifically the species has a known population size that is either:</p> <ul style="list-style-type: none"> • Fewer than 50 mature individuals independent of whether there are any threats. • Fewer than 250 mature individuals and the species has an observed, estimated or projected continuing decline: <ul style="list-style-type: none"> – of at least 25% in three years or one generation (whichever is longer), or – where the number of mature individuals in each subpopulation is < 50, or – the percentage of mature individuals in one subpopulation is 90–100%, or – the population is subject to extreme fluctuations in the number of individuals. <p>'Population' means the total number of mature individuals in New South Wales, 'Subpopulations' are geographically or otherwise distinct groups in the total population (DPIE 2019, IUCN Standards and Petitions Committee 2019). The species is not currently listed as critically endangered under the NSW BC Act or Commonwealth EPBC Act.</p> <p>Reported sizes of maternity roosts for the species range from 100 to 150,000 individuals (DPIE 2021a, Dwyer & Hamilton-Smith 1965). Despite several studies at the Church Cave maternity roost occurring within the last 60 years, it remains difficult to ascertain whether or not the population has undergone a dramatic decline (Mills 2020). Population estimates of 60,000 were reported in the 1963-1964 season though it is unclear whether this was an aberration given later counts ranged from 12,500 in 1973/74 to 35,000 in 2000 (Mills 2020). The most recent data collected at Church Cave, placed the population of females and</p>

Information required (BAM Section 9.1.2)	Response
	<p>juveniles at around 40,000 (Mills 2020), with the long-term population average of mature individuals remaining stable at approximately 20,000 after 14 years of annual monitoring, with annual variation in the order of 6-7% (D.Mills pers com, Mills and Pennay 2017). In contrast, the closely related Southern Bent-wing Bat showed declines at the maternity site in Naracoorte from 35,000 in 2001 to 20,000 in 2009, a stark contrast compared to estimates in the 1950's and 1960's of 100,000-200,000 bats (Lumsden and Jemison 2015). A similar pattern of decline for Southern Bent-wing Bat was also observed at the other known maternity site in Warrnambool which declined from 15,000 to 10,000 bats over the same time period.</p> <p>Population monitoring of Large Bent-winged Bat from Church Cave has documented no such dramatic declines over the last 14 years of consistent monitoring and is therefore indicative that the population of Large Bent-winged Bat is likely to be relatively stable. In particular, it has not been subject to any significant declines over the last three years. The size of the maternity colony at Bungonia is assumed to be similar to that of Church Cave (D.Mills pers com), with estimates in 2010 placing the adult population at around 20,000 (Mills, Pennay, & Spate 2010). Combined with the three other large maternity sites known for the species (Willi Caves in NSW, Nargun's Cave in Victoria and Riverton Cave in Queensland), the total population size is well above the threshold of 250 mature individuals of the species and likely to be in the order of 100-150,000 individuals assuming population sizes are similar amongst the five major maternity roosts.</p> <p>Similarly, extreme fluctuations have not been observed, with annual variation in the order of 6-7 % at Church Cave over 14 years of monitoring (D.Mills pers com, Mills & Pennay 2017). Although similar data is not available for other maternity sites, it is assumed that if the species was under stress this would be evident across all maternity caves and declines or extreme fluctuations would also be detected at Church Cave.</p> <p>Within NSW the Saving Our Species program has identified seven priority management sites across the state, none of which occur within the same IBRA subregion as the development footprint (DPIE 2015). The closest of these is a volcanic cave (Euglah) in the Mount Kaputar National Park, north of Tamworth, approximately 135 kilometres to the north-west of the development footprint. A second priority area is located at the known maternity site at Willi Caves, approximately 140 kilometres to the north-east of the development footprint (DPIE 2015, Dwyer 1966).</p> <p>Various studies on seasonal activities of Large Bent-winged Bat have recorded the movement from maternity to overwinter roosts in March-April, and returning in September-November (north-eastern NSW) (Williams 2019). Overwintering roosts are selected during the cooler months when insects are in low numbers, allowing bats to reduce their energy expenditure by entering torpor (Hoye & Hall 2008a). Each maternity colony is thought to act as a focal point for a particular area, with bats thought to disperse within 300 kilometres of the nearest maternity site (P.D Dwyer 1966a).</p> <p>There is no evidence that any maternity sites are located within the development footprint or close proximity of the site and the likelihood of any new and previously undiscovered sites is considered to be low given extensive previous research efforts on the species (P.D Dwyer 1966a, Dwyer & Hamilton-Smith 1965, Mills & Pennay 2017). A geomorphological assessment of the assessment area was commissioned which found that the landscape within and surrounding the development footprint was formed by a range of volcanic activity resulting in a basalt lithology (Environmental Geosurveys Pty Ltd 2021). The extent to which potential microbat habitat is present and suitable is a function initially of lithology and</p>

Information required (BAM Section 9.1.2)	Response
	<p>rock structure modified over time by geological and environmental processes. The assessment found that whilst the basalt lithology present at the development footprint may support opportunities for microbat roosts, no substantial caves were likely to be present, and that no data was found to suggest that the development footprint and immediate surrounds geomorphologically stand out from the surrounding landscape in one way or another (Environmental Geosurveys Pty Ltd 2021).</p> <p>It is therefore unclear which maternity cave the bats occurring within the subject land would be associated with. Sites at Willi Willi, Kanagara Boyd NP, Riverton, Wee Jasper and Bungonia are all within three hundred kilometres (DPIE 2021), though it may be likely that bats would gravitate towards areas with similar climatic conditions, therefore reducing the likelihood of the bats using the Willi Willi site. Dwyer (1966) considered it probable that different climatic circumstances between geographic areas may result in different patterns of winter dispersion and found a strong bias for individuals from Euglah in the nearby Mt Kapatur National Park moving to the maternity colony at Riverton for example, rather than the more coastal sub tropical Willi Willi cave or temperate sites at Bungonia and Wee Jasper. Historic records estimate the population at Riverton to be around 15-20,000 bats, slightly smaller than estimates of 25,000 at Willi Willi (Dwyer 1965). Current monitoring at Willi Willi cave is in its infancy so it is unclear what the current population is there or at Riverton (D.Mills pers comm).</p> <p>Large Bent-wing Bats have been tracked making nightly forays of 50 – 60 kilometres from the roost (Mills & Pennay 2017), with the closely related Southern Bent-wing Bat recently being found to fly 72 kilometres between caves in a matter of hours (van Harten et al. 2022). During the periods when young are dependent on their mothers for milk, lactating females are likely to make regular trips to and from the roost of 23 – 25 kilometres (Henry et al. 2002). The Project site is assumed to be beyond the range of nightly foraging associated with bats being present at the maternity cave, and is more likely to be utilised during periods of dispersal to and from the nearest maternity cave to locate overwintering or staging roost sites.</p> <p>Due to the rural location of the assessment area, where historical survey coverage is lower, the overall size of the local population is unknown. Given the absence of any known maternity sites in close proximity it is assumed that the population is transient and associated with seasonal dispersal patterns that may result in the species being absent for several months of the year during breeding. Bats associated with the site are unlikely to be occurring in large numbers, as would be expected in close proximity to a maternity roost where large numbers of females would be making regular nightly foraging trips to support their young. Utilisation of the site may also be reduced during the cooler months, when insect activity is also likely to be low and the bats are thought to enter torpor to reserve energy (Hoye and Hall 2008).</p> <p>This assessment considers the local population to include all individuals within 100 kilometres of the development footprint, which is based on a balance between the maximum recorded overnight foraging distance of 65 kilometres (Churchill 2008), and the several hundred kilometre distances the species is known to travel between maternity and non-breeding roosts (Hoye & Hall 2008).</p> <p>Previous records for the species are located approximately 5 kilometres south-west of the wind farm development footprint. This includes one record in February 2003 at Timor Caves, a known non-breeding roost location for the species (NPWS 2011), where approximately 2,000 individuals were recorded (DPIE 2021). Seven separate detection events were also</p>

Information required (BAM Section 9.1.2)	Response
	<p>recorded slightly west of the cave using acoustic recorders in 2008 (DPIE 2021). Ultrasonic detectors were deployed at Timor Caves, as part of the current assessment, for three nights in March 2020, however the species was not recorded by this survey. Presumably, bats had not yet arrived as part of the Autumn dispersal to over-wintering caves. It is also unclear what the sex-ratio or maturity status of individuals using these caves is.</p> <p>The species was recorded during the targeted surveys supporting the current assessment. These surveys were conducted from February 2020 to May 2020 and included deployment of units at ground-level as well as deployment on three meteorological masts at heights of 2 metres, c.30 metres and c.60 metres. Additional acoustic/ultrasonic survey was also undertaken in spring 2019 between 19 – 21 November 2019 (three nights) Across the 24 deployed ultrasonic detector units deployed in 2020 and the three nights of spring surveys in 2019, the total mean call for the species per night was 10.3 calls, indicating individuals of the species were utilising areas within the development footprint on most nights. Although it is not possible to infer abundance from acoustic data, these activity levels are considered relatively low, consistent with the hypothesis that the species would be occurring in small numbers at the site.</p> <p>The highest number of recorded individuals near the site was 2000 bats at Timor Caves in 2003. If we assume this represents the maximum local population, this would be equivalent to 2% of an assumed conservative overall population estimate of 100,000. However, acoustic data suggests the actual number utilising the site may be much less than this, given the average of detection rate of 10.3 calls per night attributable to this species. There is no evidence available to indicate the species has undergone a decline in the last three years. Based on the available information, the Large Bent-winged Bat does not meet the thresholds required for consideration of the species under SAIL Principle 2.</p>
<p>2c. Evidence of limited geographic range for the threatened species. (Principle 3)</p>	<p>The geographic range of a species is measured by its area of occupancy, which represents the area of suitable habitat currently occupied by the taxon (IUCN Standards and Petitions Committee 2019). Species that are known to have a very limited geographic distribution are generally known to:</p> <ul style="list-style-type: none"> • Have an area of occupancy of $\leq 10 \text{ km}^2$. • Have an extent of occurrence of $\leq 100 \text{ km}^2$. • Have at least two of the following three conditions: <ul style="list-style-type: none"> – Are severely fragmented or only known from one location. – Continuing decline. – Extreme fluctuations. • Inhabit less than or equal to three locations in NSW (DPIE 2019). <p>i. Extent of occurrence.</p> <p>The Large Bent-winged Bat is distributed along the east coast of Australia from Cape York in northern Queensland to Castlemaine in Victoria, east of the Great Dividing Range and all along the eastern coast of NSW (Churchill 2008). This extent of occurrence is significantly larger than the threshold detailed above. The western most point of the study area corresponds with the western most range of the species, which is confined to the subtropical coastal belt in the southern part of its range.</p> <p>ii. Area of occupancy.</p> <p>Similarly the area of occupancy is also large with the occurrence of Large Bent-winged Bat being mostly contiguous across its known extent within NSW, as evidenced by Bionet records for the species (DPIE 2015, DPIE 2021b). This area is significantly larger than the</p>

Information required (BAM Section 9.1.2)	Response
	<p>threshold detailed above.</p> <p>iii. Number of threat-defined locations (geographically or ecologically distinct areas in which a single threatening event may rapidly affect all species occurrences).</p> <p>Large Bent-winged Bat breeding habitat requirements are highly specific. Unlike most Australian bats, which roost in trees, Large Bent-winged Bats rely on caves to roost and raise their young (Mills & Pennay 2017). It is understood that the species forms discrete populations centred on these caves with individuals returning to the same cave to birth and rear young (DPIE 2021b). Such caves have highly specific requirements, and every year the entire female population of the species is concentrated in these maternity roosts to give birth and raise their young (Mills and Pennay 2017).</p> <p>Following breeding, individuals disperse to wintering roosts within about a 300 kilometre range of the maternity cave (DPE 2019).</p> <p>The Saving Our Species strategy for Large Bent-winged Bat (DPIE 2015) identifies seven priority management sites across NSW for the species, linked to the locations of known roosting sites. These sites are:</p> <ul style="list-style-type: none"> • Kwiambale in Inverell LGA. • Mount Kaputar in Narrabri LGA. • Willi Willi Cave in Kempsey LGA. • Yessabah in Kempsey LGA. • Church Cave in Yass Valley LGA. • Drum Cave, Bungonia. • Dip Cave. <p>Three of these sites, Willi Willi Cave, Church Cave and Drum Cave are known large maternity sites for the species in NSW, with another at Kanagara Boyd NP however less detail is available about this site. An additional two maternity sites are known outside of NSW with one each in Victoria (Nargun's Cave in East Gippsland) and Queensland (Riverton).</p> <p>None of the above sites occur within the same IBRA subregion as the development footprint (DPIE 2015). The closest of these is a volcanic over-wintering cave in the Mount Kaputar National Park, north of Tamworth, approximately 135 kilometres to the north-west of the development footprint. A second priority area includes the maternity roost at Willi Caves, approximately 140 kilometres to the north-east of the development footprint.</p> <p>There are no known maternity roosts within the assessment area. It is assumed that individuals utilising the assessment area are associated with post-breeding dispersal to over-wintering caves, primarily associated with the Timor Caves karst system where 80 caves have been identified that could form potential roosting sites (Rutledge 2008). There are also many (dozens) of historic mines and unnamed pits in the region (AUSGIN 2021) associated with the nearby town of Nundle's rich mining history, which may also provide potential roosting habitat to the local population.</p> <p>The primary threat facing the species is disturbance and damage to maternity roosting sites. This is one of the key reasons the species is listed as "vulnerable" in the BC Act (Mills & Pennay 2017). Any stochastic event occurring at one of these caves (such as a bushfire, collapse, disease or human intervention) particularly during the period of occupancy, could result in the loss of a third or more of the species population in NSW. Impacts to these maternity caves would represent the largest threat-defined locations for the species.</p> <p>A second period of vulnerability is during wintering, when individuals must conserve their energy reserves to survive (Hamilton-Smith 1970). Any disturbance to bats during this time could reduce fat reserves available to each bat and consequently reduce chances of survival. Although over-wintering bats generally group in smaller colonies, it is unclear how many such locations exist and therefore the status of their long-term protection and availability for</p>

Information required (BAM Section 9.1.2)	Response
	<p>use is unclear. Disturbance during hibernation over winter is acknowledged as a major threat to the species and National Park Management actions include discouraging visitors from entering caves and disturbing hibernating bats (NSW NPWS 2021).</p> <p>Another emerging threat to the species is the potential introduction of the fungal disease white-nose syndrome (Holz et al. 2019), which has decimated bat populations in North America. A recent risk assessment considers it 'likely' that the pathogen causing the disease (<i>Pseudogymnoascus destructans</i>) will come into contact with bats in the coming decade in Australia (Holz et al 2019). Cave-roosting bats are particularly at risk, and the clustering behaviour and high humidity of winter roosts used by the Large Bent-winged Bat leaves them more susceptible to developing white-nose syndrome in the event it enters Australia (Turbill & Welbergen 2020). Should the disease become established in Australia, it is probable that Large Bent-winged Bat populations would incur some impacts although more research is required to determine the extent and severity (Turbill & Welbergen 2020).</p> <p>iv. Whether the species' population is likely to undergo extreme fluctuations.</p> <p>Simultaneous population count data across all known maternity caves would provide the best indicator of whether the species experiences extreme fluctuations across its range. In the absence of such data, available counts from annual population monitoring at the Wee Jasper maternity cave suggest that the population of Large Bent-winged Bats is stable and fluctuates by 6-7 % (plus or minus), possibly in response to rainfall (Mills & Pennay 2017, D.Mills pers com).</p> <p>The available information on the geographic distribution of the species indicates that the species does not meet the thresholds for consideration under SAI Principle 3.</p>
<p>2d. Evidence that the species is unlikely to respond to management. (Principle 4)</p>	<p>This principle encompasses two components, firstly whether there are any particular traits of the species which limits its' response to management, and secondly whether there are any key threatening processes affecting the species which cannot be effectively managed. Furthermore in select areas where essential habitat components cannot be readily re-created (such as caves or cliff lines used by threatened species) such impacts can be deemed irreplaceable (DPIE 2019).</p> <p>i. Known reproductive characteristics severely limit the ability to increase the existing population on, or occupy new habitat on, a biodiversity stewardship site.</p> <p>The species requires maternity cave sites with specific temperature and humidity regimes in order to breed successfully. The species forms discrete populations based on these structures which individuals return to annually in order to birth and rear young (DPIE 2021b). These features are unlikely to be replicated successfully on a stewardship site, where such features are not already naturally occurring. There few such sites (three or four in NSW) known to host large maternity colonies for the species within its range.</p> <p>ii. The species is reliant on abiotic habitats which cannot be restored or replaced on a biodiversity stewardship site.</p> <p>In addition to maternity roost sites, the species also requires specific roosting habitats outside of the breeding period. During the cooler months when insect numbers are low, bats select cool areas located within caves, mines, tunnels, drains and bridges (Hoye and Hall). Whilst man-made structures can be replicated, the preferred primary roosting habitat of caves is unlikely to be replicated successfully on a stewardship site, where such features are not already naturally occurring.</p> <p>iii. Life history traits and/or ecology is known but the ability to control key threatening processes at a biodiversity stewardship site is currently negligible.</p> <p>The primary threat to the species is loss or degradation of roosting habitat and maternity roost sites. Where suitable roosting habitats occur within a biodiversity stewardship site,</p>

Information required (BAM Section 9.1.2)	Response
	<p>effective management of such features can be readily achieved.</p> <p>Secondary threats include loss of suitable foraging resources in proximity to roosting sites due to vegetation clearing and inappropriate fire regimes. Both of these can be effectively managed at a biodiversity stewardship site.</p> <p>Another emerging threat is the potential for populations to experience declines with the introduction of the fungal pathogen causing white-nose syndrome. Controlling this on a biodiversity stewardship site may be possible with strict hygiene controls and restricted access to caves.</p> <p>Given the specialised breeding and roosting habitat requirements for the species, it is unlikely such habitats can be readily replicated at a biodiversity stewardship site. As such consideration of this species under SAI Principle 4 is warranted.</p>
3. Where the TBDC indicates data is 'unknown' or 'data deficient' for a species, the assessor must record this in the BDAR or BCAR.	<p>The TBDC does not state that data is 'unknown' or 'data deficient' for this species.</p>
4a. The impact on the species' population	<p>i. An estimate of the number of individuals (mature and immature) present in the subpopulation on the subject land (the site may intersect or encompass the subpopulation) and as a percentage of the total NSW population, and,</p> <p>ii. An estimate of the number of individuals (mature and immature) to be impacted by the proposal and as a percentage of the total NSW population.</p> <p>Large Bent-winged Bat is a highly mobile species, with some individuals documented dispersing up to 300 kilometres from known maternity caves (DPIE 2021). While it is assumed that connected vegetation is preferred by the species for movement, the existence of many records of the species within urbanised areas, and its utilisation of man-made structures as roosting sites, suggests that the species does not rely on specialised dispersal or movement habitat.</p> <p>There are no known maternity caves within the assessment area, and it is considered unlikely that any new or previously undiscovered maternity sites would occur. The highest number of recorded individuals near the site was in February 2003, where there is a record of 2000 Large Bent-winged Bats associated with Timor Caves in 2003 (Bionet). Assuming this represents the maximum local population (given the paucity of other capture or observation data), this would be equivalent to <2% of an assumed conservative overall population estimate of 100,000. However, acoustic data suggests the actual number utilising the site may be much less than this, given the average of 10.3 calls per night attributable to this species, possibly in the order of 10's or 100's of bats rather than 1000's.</p> <p>Furthermore, it is expected that individuals utilising the site would be doing so in response to seasonal migrations to over-wintering roosts, where they are expected to enter torpor and consequently reduce foraging activity for several weeks or months. Activity at the site is likely to therefore be highest in Autumn and Spring, presumably when bats are migrating to and from over-wintering caves such as those at the nearby Timor Cave system.</p> <p>In designing the wind farm layout the spacing of the turbines was considered to limit interactions with fauna and an essentially linear row of turbines was designed. This is different from most wind energy facilities in Australia in which turbines are scattered across</p>

Information required (BAM Section 9.1.2)	Response
	<p>a site. In the more usual scattered or 'clustered' array, microbats and avian fauna have a high probability of encountering multiple turbines in a given flight. The configuration of Hills of Gold turbines is such that an individual is likely to encounter multiple turbines only in the rare event that it flies directly along the row of turbines. The average distance of turbines spacing across the entire site was approximately 423 metres (from rotor hub to rotor hub), which has been increased by recent layout updates. Only five turbines considered to present a moderate risk of collision of barrier effects occur are located less than 400 metres from the adjacent turbine, with a number of gaps between turbines ranging from 1.2 kilometres wide to 1.6 kilometres. These gaps provide movement corridors, allowing for migrating and foraging bats to pass through the landscape, with the risk of barrier effects assessed as low overall. Current turbine spacing and layout is also considered to reduce the chance of collision and possible barotrauma.</p> <p>Although Large Bent-wing Bats are known to fly within rotor-swept height (up to around 100 metres elevation), they have been found to be nine times more likely to be recorded closer to ground level (Mills & Pennay 2017). Activity has also been found to be seven times greater at forested sites compared to cleared sites, suggesting a preference for foraging at vegetated sites, presumably in response to insect abundance (Mills & Pennay 2017). The local population is most likely to be utilising the known non-breeding roosting habitats afforded by Timor Caves (NPWS 2011). North-south movement over the windfarm footprint is most likely to be associated with foraging flights which would be conducted at canopy level, out of turbine blade range (see BDAR Section 7.2 for details regarding turbine blade buffers). The proposed layout also retains areas of preferred foraging habitat in steeper areas of terrain, with more densely vegetated gullies preserved.</p> <p>Available data suggests that the size of the local population of Large Bent-winged Bats is estimated to represent a maximum of around 2% of the population. Given the absence of any known maternity roosts in close proximity to the subject land, it is assumed that the local population is present on a seasonal basis during Autumn and Spring migrations to and from over-wintering caves. When present, these individuals are likely to reside within the Timor Cave complex, although other areas of suitable habitat may support small colonies of bats on a transient basis as staging caves or for the purposes of over-wintering if climatic conditions are favourable. The project site is within the assumed nightly foraging range of bats roosting at Timor Caves, and acoustic data supports the hypothesis that bats utilise the site for foraging in low numbers (averaging 10.3 calls per night during monitoring). Furthermore, foraging activity is likely to be concentrated in vegetated areas below rotor-swept height, given that previously studies have shown the species to be seven times more likely to be recorded in forested areas than cleared areas and nine times more likely to be recorded at ground level than at height (Mills & Pennay 2017).</p> <p>Globally, collision with wind turbines is the leading cause of multiple mortality events in bats (van Harten et al. 2022). Collision data in Victoria has found the majority of mortalities at wind farms to be the White-striped Freetail Bat (<i>Austronomus australis</i>) (Moloney, Lumsden, & Smales 2019). The closely related Southern Bent-wing Bat was recorded as a mortality on just eight occasions from monitoring across 15 operational windfarms from 2003 - 2018. Although the actual number of bats killed may be higher once carcass persistence rates and searcher efficiency rates are factored into overall mortality estimates, the current data suggests that the number of collisions is likely to be low. Given that the Large Bent-winged Bat is closely related to the Southern Bent-winged Bat and shows similar morphological adaptations and flight characteristics, it could be inferred that the species may have similar</p>

Information required (BAM Section 9.1.2)	Response
	<p>collision rates.</p> <p>Overall, although collision mortality is likely to be low and may occur only during some circumstances associated with seasonal movement patterns, prior to implementation of the project's mitigation strategy, it is considered a moderate risk that repeated loss of individuals as a result of collision mortality may cause change to the local abundance of this species in the short term (up to 5 years). However given the local population is estimated to represent <2% of the species' NSW population, with individuals utilising the site estimated to be in the 10's or 100's of bats rather than 1000's and only present on a transient seasonal basis, potential impacts associated with collision with turbines are considered likely to equate to a very small percentage of the total NSW population of the species.</p> <p>Efforts committed to by the Proponent to minimise the potential for collision with turbines, and hence minimise the risk of the project resulting in an operational SAIL to the species are presented above in Section 0.</p>
<p>4b. Impact on geographic range.</p>	<p>i. The area of the species' geographic range to be impacted by the proposal in hectares, and a percentage of the total AOO, or EOO within NSW.</p> <p>The impacts to available habitat for Large Bent-winged Bat as a result of the project is likely to represent <0.001 % of the extent of occurrence for the species, along the east coast of Australia from Cape York in northern Queensland to Castlemaine in Victoria, east of the Great Dividing Range and all along the eastern coast of NSW (Churchill 2008). Similarly it is likely to represent <0.001 % of the occurrence for the species with BioNet records indicating the species is primarily contiguous across its known range.</p> <p>ii. The impact on the subpopulation as either: all individuals will be impacted (subpopulation eliminated); OR impact will affect some individuals and habitat; OR impact will affect some habitat, but no individuals of the species will be directly impacted.</p> <p>The proposed impacts are unlikely to affect potential habitat for the species. Direct mortality to individuals in the sub-population present in the area on a seasonal or transient basis may occur as a result of collisions with turbines.</p> <p>Available data is scarce regarding the size of the sub-population. Nearby records are predominantly acoustic observations, from which it is not possible to infer abundance. The most informative record was in February 2003 where 2000 bats were reportedly observed in Timor Caves (Bionet). Main Cave is considered an important roosting site for the species (NPWS 2011), though it is unclear when and how the bats use the cave over time.</p> <p>The best available data suggests that the local population may include up to 2000 bats (<2% of the total population), that probably undertake seasonal movements to and from the area between maternity and other roosts. Acoustic data further suggests that activity levels for Large Bent-winged Bat are relatively low, averaging 10.3 calls per night across the survey period. Previous research also suggests that when present, foraging is likely to be concentrated at ground-level rather than at rotor-swept height and within vegetated areas. This hypothesis is supported by collision data from Victorian windfarms that found the number of likely mortalities of Southern Bent-winged Bat to be low, compared with recognised high-flying species that forage well above the canopy such as the White-striped Freetail Bat.</p> <p>Overall, although the project may result in direct mortality to a small number of individuals, it is considered unlikely that this would cause a significant impact to the subpopulation such that it would be eliminated.</p>

Information required (BAM Section 9.1.2)	Response
	<p>iii. To determine if the persisting subpopulation that is fragmented will remain viable, estimate the habitat area required to support the remaining population, and habitat available within dispersal distance, and distance over which genetic exchange can occur and pollination distance for the species.</p> <p>Approximately 2.54 million hectares of native vegetation occurs within a 100 kilometre buffer zone surrounding the development footprint. These areas represent quality foraging habitat for the species which typically hunts in forest areas, catching moths and other flying insects above the tree canopy (DPIE 2021c). This represents the foraging habitat available with the range of the local population. Only 132.43 hectares of native vegetation occurs within the development footprint, a fraction of the vegetation within the locality. Of this vegetation approximately 37.92 hectares (28.6 %) is in high condition providing good habitat values for foraging native fauna, including Large Bent-winged Bat (<i>note these areas will be updated on final BDAR review</i>). This foraging habitat would be modified as a result of the Project.</p> <p>Large Bent-winged Bat is a highly mobile species, capable of travelling several hundreds of kilometres between roost sites (Hoye and Hall 2008), even up to 1300 kilometres (Churchill 2008). The species uses a broad range of habitats for foraging including rainforests, wet and dry sclerophyll forests and vine thicket, hunting in timbered areas, catching beetles, moths and flies above the canopy trees (Churchill 2008, DPIE 2021). While it is assumed that connected vegetation is preferred by the species for movement, the existence of many records of the species within urbanised areas, and its utilisation of man-made structures as roosting sites, suggests that the species does not rely on specialised dispersal or movement habitat.</p> <p>The local population is expected to occur on a seasonal and transient basis in response to seasonal movements, which are unlikely to be impacted as a result of the Project such that the sub-population would become fragmented or unviable. The available foraging habitat to these individuals, which are capable of travelling large distances, possibly over 70 km in a single night (van Harten et al. 2022), is unlikely to be significantly impacted as a result of the Project and no impacts on genetic exchange are anticipated given that the population is unlikely to become fragmented as a result of the Project.</p> <p>iv. To determine changes in threats affecting remaining subpopulations and habitat if the proposed impact proceeds, estimate changes in environmental factors including changes to fire regimes (frequency, severity); hydrology, pollutants; species interactions (increased competition and effects on pollinators or dispersal); fragmentation, increased edge effects, likelihood of disturbance; and disease, pathogens and parasites. Where these factors have been considered elsewhere in relation to the target species the assessor may refer to the relevant sections of the BDAR or BCAR.</p> <p>Changes to fire regimes</p> <p>The risk of fire as a result of sparks from machinery during proposed works is unlikely, but could increase the risk of fire occurring in foraging habitat or nearby potential roost sites. This risk will be managed by implementing appropriate mitigation measures such as spark dampeners, water spraying or the close proximity of fire-fighting gear such as extinguishers within the project's CEMP.</p> <p>Upon completion of works, the proposal is unlikely to result in changes to frequency or intensity of fire regimes within the locality.</p>

Hydrology

Changes to hydrology as a result of the proposal are discussed in the BDAR. These changes are unlikely to significantly impact on the foraging or roosting behaviour of Large Bent-winged Bat.

Pollutants

The project does not involve the use of any pesticides, nor is it likely to substantially increase the levels of pesticides within the environment.

Species interactions

Introduced predators such as feral cats and foxes can negatively impact the species by preying on bats as they exit caves, sometimes taking significant numbers. One study reported 476 Bent-winged bats (prior to the taxonomic subspecies change) being predated (P.D Dwyer 1966b). Black Rats have also been reported in maternity caves and are likely to prey upon young (P.D Dwyer 1966b, Lumsden & Jemison 2015). Such predators were recorded during camera trap surveys, are likely to already present within the broader locality, and are unlikely to increase as a result of the proposal.

Fragmentation

Large Bent-winged Bat is a highly mobile species capable of dispersing across breaks in habitats. While it is assumed that connected vegetation is preferred by the species for movement, the existence of many records of the species within urbanised areas suggests that the species does not rely on specialised dispersal or movement habitat. Thus, the loss of approximately 37.92 hectares (*note these areas will be updated on final BDAR review*) of native vegetation from the buffer area is unlikely to impact the movement ecology of the local population. As such population fragmentation will not occur as a result of the proposal.

Increased edge effects

The proposed works that occur within the vicinity of the identified potential Large Bent-winged Bat habitats are occurring within a previously disturbed area as a result of existing agricultural and industrial practices. As such these areas are already subject to some edge effects. Whilst transport of weeds is possible during construction, these will be mitigated through the application of appropriate weed control measures to be detailed within the CEMP. This will ensure the existing edge effects are not exacerbated by the proposal. These edge effects are also unlikely to significantly impact the species, given the availability of foraging resources within the locality.

Likelihood of disturbance

Disturbance will occur in the form of direct impacts to native vegetation that may remove a small amount of potential foraging habitat. There would be no disturbance to roosting or breeding habitat.

Operational impacts may also result in disturbance to the seasonal or foraging movements of the bats, including the potential for mortalities due to collisions with turbines. However, the potential for substantial numbers of mortalities associated with turbine collision is considered to be low. This is due to the likely small size of the local population, the seasonal and transient nature of their likely occurrence in the area, the foraging preferences shown by the species suggesting they are more likely to forage at ground level and in association with vegetated areas, and the seasonal habits of the species entering torpor and therefore showing reduced periods of activity that would render the species less likely to encounter turbines while present in the area. Although disturbance to Large Bent-winged Bat (including collisions) is possible as a result of the Project, these impacts are likely to be restricted to a small number of individuals. Based on the factors above, it is considered unlikely that disturbance (including a small number of possible mortalities), would result in impacts at the population level.

Information required (BAM Section 9.1.2)	Response
	<p>In designing the wind farm layout the spacing of the turbines was considered to limit interactions with fauna and an essentially linear row of turbines was designed. This is different from most wind energy facilities in Australia in which turbines are scattered across a site. In the more usual scattered or 'clustered' array, microbats and avian fauna have a high probability of encountering multiple turbines in a given flight. The configuration of Hills of Gold turbines is such that an individual is likely to encounter multiple turbines only in the rare event that it flies directly along the row of turbines. The average distance of turbines spacing across the entire site was approximately 423 metres (from rotor hub to rotor hub), which has been increased by recent layout updates. Only five turbines considered to present a moderate risk of collision of barrier effects occur are located less than 400 metres from the adjacent turbine, with a number of gaps between turbines ranging from 1.2 kilometres wide to 1.6 kilometres. These gaps provide movement corridors, allowing for migrating and foraging bats to pass through the landscape, with the risk of barrier effects assessed as low overall. Current turbine spacing and layout is also considered to reduce the chance of collision and possible barotrauma.</p> <p>Overall, although there is potential for disturbance to the seasonal or foraging movements of the species, the likelihood of this disturbance representing a significant impact to the local population is considered to be low.</p> <p>Disease, pathogens and parasites</p> <p>An emerging threat to Australian bats, particularly cave-roosting species, is the fungal disease white-nose syndrome. To date there have been no cases of white-nose syndrome recorded in Australia however, a recent risk assessment considers it 'likely' that the pathogen causing the disease (<i>Pseudogymnoascus destructans</i>) will come into contact with Australian bats in the coming decade (Holz et al 2019).</p> <p>Cave-roosting bats are particularly at risk, and the clustering behaviour and high humidity of winter roosts used by the Large Bent-winged Bat leaves them more susceptible to developing white-nose syndrome in the event it enters Australia (Turbill & Welbergen 2020). Should the disease become established in Australia, it is probable that Large Bent-winged Bat populations would incur some impacts although more research is required to determine the extent and severity (Turbill & Welbergen 2020).</p> <p>The project would not contribute to any increased risk in the introduction of white-nose syndrome.</p>
<p>5. The assessor may also provide new information that can be used to demonstrate that the principle identifying the species as at risk of an SAIL, is inaccurate.</p>	<p>Not applicable.</p>

Little Bent-winged Bat SAI assessment

Little Bent-winged Bat is a member of the Miniopteridae family (Churchill 2008). The species occurs along the east coast of Australia from northern Queensland to north of Batemans Bay in NSW, and is listed as Vulnerable under the BC Act.

Little Bent-winged Bat uses a broad range of habitats including moist eucalypt forest, rainforest, vine thicket, wet and dry sclerophyll forest, *Melaleuca* spp. swamps, dense coastal forests and banksia scrub (Churchill 2008, DPIE 2019). Little Bent-winged bats prefer well-timbered areas where they feed primarily in the shrub and canopy layers. Their diet consists primarily of beetles, moths, flies and spiders (Churchill 2008).

Little Bent-winged Bat is considered a species credit species under the BAM, specifically in relation to the species' breeding habitat. This breeding habitat is defined in the Atlas of NSW Wildlife (DPIE 2021) as any:

'Cave, tunnel, mine, culvert or other structure known or suspected to be used for breeding including species records in BioNet with microhabitat code 'IC – in cave'; observation type code 'E nest-roost'; with numbers of individuals >500; or from the scientific literature'.

Impacts to these breeding habitat features, and the area within a 100 metre radius buffer around an accurate GPS point location centred on the cave / feature entrance, are what are considered to be potentially serious and irreversible for the species (DPIE 2021).

Amendments to the project design have since resulted in impacts to such areas, that previously occurred within the project footprint, being successfully avoided. However, following correspondence with BCS (April 2022), the species is still considered at risk of an SAI as a result of the potential operational impacts of the project. In particular, the risk of mortalities due to collisions with wind turbines.

As described in the previous SAI for Large Bent-winged Bat and Large-eared Pied Bat, it is impossible to numerically quantify potential risk of collisions for Little Bent-winged Bat.

The following SAI has been prepared to address the potential operational impacts to Little Bent-winged Bat as a result of the project.

Table 105 SAI assessment for Little Bent-winged Bat

Information required (BAM Section 9.1.2)	Response
1. The action and measures taken to avoid the direct and indirect impact on the species at risk of an SAI.	Addressed above in Section 4.2 collectively for SAI bat species.
2a. Evidence of rapid decline. (Principle 1)	<p>i. Evidence of rapid decline in the population of the species in NSW in the past 10 years or three generations (whichever is longer), or</p> <p>ii. decline in population of the species in NSW in the past 10 years or three generations (whichever is longer) as indicated by: an index of abundance appropriate to the species; decline in geographic distribution and/or habitat quality; exploitation; effect of introduced species, hybridisation, pathogens, pollutants, competitors or parasites.</p> <p>Species and ecological communities that have undergone large reductions or are likely to undergo large reductions in the future are considered to be at greater risk of extinction than those that have undergone or are likely to undergo smaller reductions (NSW TSSC 2018).</p>

Information required (BAM Section 9.1.2)	Response
	<p>To be considered under this principle, the species should have an observed, estimated, inferred, suspected or projected population of $\geq 80\%$ in 10 years or three generations (whichever is longer) (DPIE 2019).</p> <p>Numerical data quantifying the overall size of the Little Bent-winged Bat population (and therefore evidence of any declines) is scarce. The best available method of estimating the population size of cave-dwelling bats is to conduct population counts at known maternity roosts, where the population is concentrated during the breeding season. Ideally, concurrent counts would occur across the known range of these roosts to provide an overall population estimate. In the absence of such data, we must rely on the evidence available to assess the population size and any likely declines.</p> <p>Maternity roosts for Little Bent-winged Bat are rare with only seven known sites within Australia (Augusteyn, Matthews, & Richards 2022). The few documented sites are normally situated in limestone cave systems, although Little Bent-winged Bat is also known to roost in abandoned mines, tunnels, stormwater drains and occasionally buildings (Churchill 2008). Maternity roosts for species the range in size from 3,000 to 4,000 individuals (Dwyer 1968) up to an estimated 139,000 individuals observed at the largest known maternity roost for the species at Mt Etna in Queensland (Augusteyn, Matthews, & Richards 2022). Recent monitoring has found fluctuations in the population at the Mt Etna roost from around 84,000 in 2015 to a peak of around 139,000 in 2018. The current estimate is around 121,000, and the population at this site is considered to be stable (Augusteyn, Matthews, & Richards 2022).</p> <p>In NSW, the largest maternity colony for Little Bent-winged Bat is Willi Willi Cave, which has historically consisted of around 4000 individuals (1800 young) (Dwyer 1968). This colony co-occurs with a large maternity colony of Large Bent-winged Bat. More recent count data for Little Bent-winged Bat at Willi Willi Cave is unavailable however, Large Bent-winged Bat population monitoring has indicated that the population of this species has remained stable over the past 14 years of monitoring (Mills 2020). Given that Little Bent-winged Bat is thought to be dependent upon the heat generated by the larger Large Bent-winged Bat maternity colonies in the southern extent of its range in order to successfully rear its young (Dwyer 1968), it could be inferred that stability in the Large Bent-winged Bat population coupled with stable population counts of Little Bent-winged Bat colonies in other parts of its range indicate that this population is also likely to be stable.</p> <p>There is no available data to indicate that the population has experienced rapid declines of $\geq 80\%$ in the last 10 years within the scientific literature. Available population estimates indicate the population is stable and important maternity nesting sites continue to be utilised over a significant time period. It is unlikely that a decline in the species is being experienced that would trigger consideration of the species under SAI Principle 1.</p>
2b. Evidence of small population size. (Principle 2)	<p>i. An estimate of the species' current population size in NSW, and</p> <p>ii. An estimate of the decline in the species' population size in NSW in three years or one generation (whichever is longer), and</p> <p>iii. Where such data is available, an estimate of the number of mature individuals in each subpopulation, or the percentage of mature individuals in each subpopulation, or whether the species is likely to undergo extreme fluctuations.</p> <p>Species with small population sizes are highly vulnerable to any event which impacts and further reduces their population size due to the time-lag between developmental impacts and the realisation of ecological benefits from improvements in habitat condition at</p>

Information required (BAM Section 9.1.2)	Response
	<p>stewardship offset sites (DPIE 2019).</p> <p>To be considered under this principle a species must have a very small population size which would lead it to be considered critically endangered under the <i>IUCN Red List Categories and Criteria</i> (IUCN 2012). Specifically the species has a known population size that is either:</p> <ul style="list-style-type: none"> • Fewer than 50 mature individuals independent of whether there are any threats. • Fewer than 250 mature individuals and the species has an observed, estimated or projected continuing decline: <ul style="list-style-type: none"> – of at least 25% in three years or one generation (whichever is longer), or – where the number of mature individuals in each subpopulation is <50, or – the percentage of mature individuals in one subpopulation is 90–100%, or – the population is subject to extreme fluctuations in the number of individuals. <p>‘Population’ means the total number of mature individuals in New South Wales, ‘Subpopulations’ are geographically or otherwise distinct groups in the total population (DPIE 2019, IUCN Standards and Petitions Committee 2019). The species is not currently listed as critically endangered under the NSW BC Act or Commonwealth EPBC Act.</p> <p>The most recent population estimates come from monitoring of the largest known maternity site for the species in Queensland, which considered the population to be stable at around 121,000 individuals, representing 80% of the population (Augusteyn, Matthews, & Richards 2022).</p> <p>The only known nursery colony for Little Bent-winged Bat in NSW occurs at Willi Willi Bat Cave, which also represents the southernmost maternity roost of the species (Dwyer 1968). This nursery colony was historically estimated to include a peak population size of 6850 individuals (Dwyer 1968). Dwyer (1968) also included colony estimates at three non-breeding roosting caves (Carrai, Yessabah and Big Hill), with numbers tending to be around 500 to 600 individuals. Little Bent-winged Bats are known to make substantial seasonal movements to over-wintering caves, with up to 6000 bats known to utilise Balickera Tunnel near Newcastle as a winter roost, some 300 km south of the maternity cave at Willi Willi (Eco Logical 2021). More contemporary estimates of population sizes are not available within the literature for NSW, though it can be inferred that the state supports less than 20% of the total population of Little Bent-winged Bat.</p> <p>In contrast to other threatened bat species such as the Southern Bent-winged Bat where significant declines have been observed (Lumsden & Jemison 2015), there are no documented declines in the population of Little Bent-winged Bat over the last three years or generation in the scientific literature.</p> <p>Given the reported sizes of the individual colonies, coupled with the large breeding congregations that are reported to occur at maternity roost sites, the species does not meet the thresholds required for consideration of the species under SAIL Principle 2.</p>
2c. Evidence of limited geographic range for the threatened species. (Principle 3)	<p>The geographic range of a species is measured by its area of occupancy, which represents the area of suitable habitat currently occupied by the taxon (IUCN Standards and Petitions Committee 2019). Species that are known to have a very limited geographic distribution are generally known to:</p> <ul style="list-style-type: none"> • Have an area of occupancy of $\leq 10 \text{ km}^2$. • Have an extent of occurrence of $\leq 100 \text{ km}^2$. • Have at least two of the following three conditions: <ul style="list-style-type: none"> – Are severely fragmented or only known from one location. – Continuing decline.

Information required (BAM Section 9.1.2)	Response
	<ul style="list-style-type: none"> – Extreme fluctuations. • Inhabit less than or equal to three locations in NSW (DPIE 2019). <p>i. Extent of occurrence.</p> <p>The species extent of occurrence is very large as it occurs along the east coast of Australia, ranging from north of Batemans Bay up to Cape York in Queensland (DPIE 2021a).</p> <p>ii. Area of occupancy.</p> <p>Similarly the area of occupancy is also large with the occurrence of Little Bent-winged Bat being mostly contiguous across its known extent within NSW, as evidenced by Bionet records for the species (DPIE 2015, DPIE 2021b). This area is significantly larger than the threshold detailed above.</p> <p>iii. Number of threat-defined locations (geographically or ecologically distinct areas in which a single threatening event may rapidly affect all species occurrences).</p> <p>The Little Bent-winged Bat is a cave-dwelling species, congregating in the summer months at maternity caves and dispersing during autumn and winter (Churchill 2008). Such caves have highly specific requirements, even more so for this species which is thought to rely on the larger Large Bent-winged Bat to facilitate the ideal thermal properties required for successfully rearing young (Dwyer 1968).</p> <p>The primary threat facing the species is disturbance and damage to maternity roosting sites. Any stochastic event occurring at one of these caves (such as a bushfire, collapse, disease or human intervention) particularly during the period of occupancy, could result in the loss of a large proportion of the population in NSW. Impacts to these maternity caves would represent the largest threat-defined locations for the species.</p> <p>The Saving Our Species program for Little Bent-winged Bat recognises the only known maternity site in NSW for the species at Willi Willi Cave in northern NSW as a key management site (DPIE 2015b).</p> <p>The Saving Our Species strategy also lists two priority management areas for the species, one north of Newcastle and one near Byron Bay. These sites are associated with significant non-breeding roosts known for the species (Dwyer 1968, Eco Logical 2021). These roosts are important as they provide cooler sites that assist bats in conserving energy during the cooler months (Dwyer 1968). During the non-breeding season the Little Bent-winged Bat population is thought to scatter and form smaller colonies in cooler caves where they remain active for large parts of the winter, unlike Large Bent-winged Bat which enters longer periods of torpor (Dwyer 1968). Disturbance during hibernation over winter is acknowledged as a major threat to the species and national park management actions include discouraging visitors from entering caves and disturbing hibernating bats (NSW NPWS 2021). It is unclear how many non-breeding caves occur in NSW however, none have been documented in the assessment area.</p> <p>Another emerging threat to the species is the potential introduction of the fungal disease white-nose syndrome (Holz et al. 2019), which has decimated bat populations in North America. A recent risk assessment considers it 'likely' that the pathogen causing the disease (<i>Pseudogymnoascus destructans</i>) will come into contact with bats in the coming decade in Australia (Holz et al 2019). Cave-roosting bats are particularly at risk, and the clustering behaviour and high humidity of winter roosts used by the Little Bent-winged Bat leaves them susceptible to developing white-nose syndrome in the event it enters Australia (Turbill & Welbergen 2020). Should the disease become established in Australia, it is probable that Little Bent-winged Bat populations would incur some impacts although more research is required to determine the extent and severity (Turbill & Welbergen 2020). Conditions in the</p>

Information required (BAM Section 9.1.2)	Response
	<p>southern part of the species range are likely to be better suited to <i>P.destructans</i> growth (Turbill & Welbergen 2020), so the large geographic distribution of the species may provide it with some resilience in the event of the disease becoming established.</p> <p>iv. Whether the species' population is likely to undergo extreme fluctuations.</p> <p>Simultaneous population count data across all known maternity caves would provide the best indicator of whether the species experiences extreme fluctuations across its range. In the absence of such data, we can infer that the population appears to be stable based on count data from the largest known maternity colony for the species in Queensland (Augusteyn, Matthews, & Richards 2022). Persistence at known roost sites over long time period similarly suggests that the species is stable within its range. There is no evidence available to suggest the Little Bent-winged Bat is likely to undergo extreme fluctuations in the population.</p> <p>The available information on the geographic distribution of the species indicates that the species does not meet the thresholds for consideration under SAI Principle 3.</p>
<p>2d. Evidence that the species is unlikely to respond to management. (Principle 4)</p>	<p>This principle encompasses two components, firstly whether there are any particular traits of the species which limits its' response to management, and secondly whether there are any key threatening processes affecting the species which cannot be effectively managed. Furthermore in select areas where essential habitat components cannot be readily re-created (such as caves or cliff lines used by threatened species) such impacts can be deemed irreplaceable (DPIE 2019).</p> <p>i. Known reproductive characteristics severely limit the ability to increase the existing population on, or occupy new habitat on, a biodiversity stewardship site.</p> <p>The species requires maternity cave sites with specific temperature and humidity regimes in order to breed successfully. The species forms discrete populations based on these structures which individuals return to annually in order to birth and rear young (DPIE 2021b, Dwyer 1968). Furthermore, it is believed the temperature of some caves is dependent upon the mixed breeding congregations formed with the Large Bent-winged Bat, which is a larger species that gathers at maternity colonies in even greater numbers than Little Bent-winged Bat. It is thought that the colonisation of the Little Bent-winged Bat, which is essentially a tropical species, into the southern regions of its distribution has been dependent on the establishment of maternity colonies in association with the larger Large Bent-winged Bat (DPIE 2021b, Dwyer 1968).</p> <p>These features are unlikely to be replicated successfully on a stewardship site unless already naturally occurring.</p> <p>ii. The species is reliant on abiotic habitats which cannot be restored or replaced on a biodiversity stewardship site.</p> <p>In addition to maternity roost sites, the species also requires specific roosting habitats in proximity to foraging resources in the form of caves, derelict mines, storm-water tunnels, buildings and other man-made structures (DPIE 2021b). Whilst man-made structures can be replicated, the species' preferred primary roosting habitat of caves is unlikely to be replicated successfully on a stewardship site, where such features are not already naturally occurring.</p> <p>iii. Life history traits and/or ecology is known but the ability to control key threatening processes at a biodiversity stewardship site is currently negligible.</p> <p>The primary threat to the species is loss or degradation of roosting habitat and maternity</p>

Information required (BAM Section 9.1.2)	Response
	<p>roost sites. Where suitable roosting habitats occur within a biodiversity stewardship site, effective management of such features can be readily achieved.</p> <p>Another emerging threat is the potential for populations to experience declines with the introduction of the fungal pathogen causing white-nose syndrome. Controlling this on a biodiversity stewardship site may be possible with strict hygiene controls and restricted access to caves.</p> <p>Secondary threats include loss of suitable foraging resources in proximity to roosting sites due to vegetation clearing and inappropriate fire regimes. Both of these can be effectively managed at a biodiversity stewardship site.</p> <p>Given the specialised breeding and roosting habitat requirements for the species, it is unlikely such habitats can be readily replicated at a biodiversity stewardship site. As such consideration of this species under SAI Principle 4 is warranted.</p>
3. Where the TBDC indicates data is 'unknown' or 'data deficient' for a species, the assessor must record this in the BDAR or BCAR.	<p>The TBDC does not state that data is 'unknown' or 'data deficient' for this species.</p>
4a. The impact on the species' population	<p>i. An estimate of the number of individuals (mature and immature) present in the subpopulation on the subject land (the site may intersect or encompass the subpopulation) and as a percentage of the total NSW population, and,</p> <p>ii. An estimate of the number of individuals (mature and immature) to be impacted by the proposal and as a percentage of the total NSW population.</p> <p>The Little Bent-winged Bat is a highly mobile species, known to disperse up to 300 kilometres from known maternity caves (DPIE 2021). While it is assumed that connected vegetation is preferred by the species for movement, the existence of many records of the species within urbanised areas, and its utilisation of man-made structures as roosting sites, suggests that the species does not rely on specialised dispersal or movement habitat.</p> <p>Little Bent-winged Bat is essentially a tropical bat whose range extends into subtropical northern NSW, likely supported by the presence of Large Bent-winged Bat colonies providing suitable conditions for successful breeding by Little Bent-winged Bat (Churchill 2008, Dwyer 1968). It is thought that dependence upon Large Bent-winged Bat nursery colonies may result in localised populations of Little Bent-winged Bat associated with this species (Dwyer 1968). Its distribution is known to be more limited than that of the Large Bent-winged Bat which occurs abundantly in many roosts of the Northern Tablelands and North-western Slopes (Dwyer 1968). In contrast, the Little Bent-winged Bat is restricted to the subtropical coastal region other than a few colonies located on the coastal escarpment (Dwyer 1968). This is reflected in the Bionet records for the species, which are overwhelmingly concentrated around coastal NSW.</p> <p>In terms of the local population size, there is no evidence of maternity roost sites within the subject land for Little Bent-winged Bat or for Large Bent-winged Bat with which the species is thought to rely on for successful breeding in NSW and it is considered unlikely that any new or previously undiscovered maternity sites would occur based on extensive previous research in local caves (Dwyer 1968, P.D Dwyer 1966a, Dwyer & Hamilton-Smith 1965). The western extent of the study area corresponds with the known western extent of the species'</p>

Information required (BAM Section 9.1.2)	Response
	<p>range (Australasian Bat Society 2022). This is consistent with the small number of documented records for the species in the locality. Two historical records for Little Bent-winged Bat are located approximately 14 km south-west of the wind farm development footprint. These records were based on ultrasonic recording that were made in December 2016 and December 2018 (DPIE 2021). The records occur within the Timor Caves locality, which is a known non-breeding roost location for Large Bent-winged Bat (NPWS 2011, Dwyer 1968), although there are no records of Little Bent-winged Bat roosting in this area. Historical survey coverage is likely to be low in the area due to its rural location, resulting in further uncertainty around the size of the local population. This assessment considers the local population to include all individuals within 100 km of the development footprint, which is a balance between the recorded maximum overnight distance of 60 kms (Dwyer 1968), and the over 300 km distances the species is known to travel between maternity and non-breeding roost locations (Eco Logical 2021, Hoyer & Hall 2008b).</p> <p>Based on the species distribution, previous records in the locality, and relatively low levels of activity recorded as part of the current assessment, Little Bent-winged Bat is assumed to occur in the subject land in small numbers on a seasonal and transient basis. The majority of the population is assumed to concentrate on maternity caves during the summer months, and are therefore most likely only to be present in the locality following dispersal during autumn and winter to non-breeding caves. As there are no known maternity or non-breeding caves present within the locality, it is assumed that Little Bent-winged Bats are utilising the subject land for foraging as part of a broad foraging range around non-breeding roosts.</p> <p>This hypothesis is supported by acoustic surveys undertaken during the current assessment. These surveys were conducted from February 2020 to May 2020 and included deployment of units at ground-level as well as deployment on three meteorological masts at heights of 2 metres, c.30 metres and c.60 metres. Additional acoustic/ultrasonic survey was also undertaken in spring 2019 between 19 – 21 November 2019 (three nights) Across the 24 deployed ultrasonic detector units deployed in 2020 and the three nights of spring surveys in 2019, the total mean call for the species per night was 2.6 calls, indicating a low activity within the development footprint on an infrequent basis. The maximum number of calls recorded in a single night was 43.</p> <p>Although activity data cannot be used to infer abundance for microbats, the low call activity is suggestive of a small number of bats likely to be present at any given time, probably in the order of 10's of individuals at most.</p> <p>The size of the NSW population is unknown however, the maximum recorded number of Little Bent-winged Bats occurring at the only known roost site was around 6000 individuals (Dwyer 1968). Similarly, Balickera Tunnel, the largest and southernmost non-breeding cave for the species, is known to support up to 6000 individuals (Eco Logical 2021). If we assume that up to 100 individuals occur locally at the subject land, this would represent <2% of the NSW population and <0.1% of the overall population (assumed to be >100,000).</p> <p>It is difficult to quantify the number of individuals likely to be impacted. In designing the wind farm layout the spacing of the turbines was considered to limit interactions with fauna and an essentially linear row of turbines was designed. This is different from most wind energy facilities in Australia in which turbines are scattered across a site. In the more usual scattered or 'clustered' array, microbats and avian fauna have a high probability of encountering multiple turbines in a given flight. The configuration of Hills of Gold turbines is such that an individual is likely to encounter multiple turbines only in the rare event that it flies directly</p>

Information required (BAM Section 9.1.2)	Response
	<p>along the row of turbines. The average distance of turbines spacing across the entire site was approximately 423 metres (from rotor hub to rotor hub), which has been increased by recent layout updates. Only five turbines considered to present a moderate risk of collision of barrier effects occur are located less than 400 metres from the adjacent turbine, with a number of gaps between turbines ranging from 1.2 kilometres wide to 11.6 kilometres. These gaps provide movement corridors, allowing for migrating and foraging bats to pass through the landscape, with the risk of barrier effects assessed as low overall. Current turbine spacing and layout is also considered to reduce the chance of collision and possible barotrauma.</p> <p>Although Little Bent-winged Bats exhibit fast flight and may fly within rotor-swept height, they are likely to concentrate foraging activity in densely vegetated areas where they fly between the shrub and canopy layers to locate spiders, beetles, moths and other prey (Churchill 2008). As there are no known maternity or non-breeding caves present within the locality, it is assumed that Little Bent-winged Bats are utilising the study area for foraging as part of a broad foraging range around non-breeding roosts.</p> <p>Globally, collision with wind turbines is the leading cause of multiple mortality events in bats (van Harten et al. 2022). Collision data in Victoria has found the majority of mortalities at wind farms to be the White-striped Freetail Bat (<i>Austronomus australis</i>) (Moloney, Lumsden, & Smales 2019). The cave-roosting Southern Bent-winged Bat, which similarly migrates between maternity and non-breeding caves, was recorded as a mortality on just eight occasions from monitoring across 15 operational windfarms from 2003 - 2018. Although the actual number of bats killed may be higher once carcass persistence rates and searcher efficiency rates are factored into overall mortality estimates, the current data suggests that the number of collisions is likely to be low. Given that the Little Bent-winged Bat shows similar behavioural patterns and flight characteristics to Southern Bent-wing Bat, it could be inferred that the species may have similar collision risk.</p> <p>Overall, collision mortality is likely to be low given the design of the wind farm and the small number of individuals presumed to be present at any one time (<0.1% of the total population), combined with the known distribution, movement and foraging patterns for the species.</p> <p>Efforts committed to by the Proponent to minimise the potential for collision with turbines, and hence minimise the risk of the project resulting in an operational SAI to the species are presented above in Section 0.</p> <p>Notwithstanding the above, a conservative approach has been taken in considering there to be a moderate risk that repeated loss of individuals as a result of collision mortality may cause change to the local abundance of this species in the short term (up to 5 years).</p>
4b. Impact on geographic range.	<p>i. The area of the species' geographic range to be impacted by the proposal in hectares, and a percentage of the total AOO, or EOO within NSW.</p> <p>The impacts to available habitat for Little Bent-winged Bat as a result of the proposed works are likely to represent <0.001 % of the extent of occurrence for the species, along the east coast of Australia from Cape York in northern Queensland down along the eastern coast of NSW to Batemans Bay (Churchill 2008). Similarly it is likely to represent <0.001 % of the occurrence for the species with BioNet records indicating the species is primarily contiguous across its known range.</p> <p>ii. The impact on the subpopulation as either: all individuals will be impacted (subpopulation eliminated); OR impact will affect some individuals and habitat; OR</p>

Information required (BAM Section 9.1.2)	Response
	<p>impact will affect some habitat, but no individuals of the species will be directly impacted.</p> <p>The proposed impacts are unlikely to affect potential habitat for the species. Direct mortality to individuals in the sub-population present in the area only on a seasonal or transient basis may occur as a result of collisions with turbines.</p> <p>As described above, available data is scarce regarding the size of the local population or the population in NSW. Only two historical records for Little Bent-winged Bat are documented in the area, approximately 14 kilometres south-west of the wind farm development footprint. Acoustic surveys conducted during the current assessment found a low call rate for this species, averaging 2.6 calls per night, indicating a low activity within the development footprint on an infrequent basis. The maximum number of calls recorded in a single night was 43. Although activity data cannot be used to infer abundance for microbats, the low call activity is suggestive of a small number of bats likely to be present at any given time, probably in the order of 10's of individuals at most.</p> <p>The distribution of the species, known movement patterns, the scarcity of acoustic records in the area and the absence of any known roost sites, provides further evidence that the local population is likely to be low, likely only to be occurring on a seasonal or transient basis in response to seasonal movements between maternity and non-breeding roosts.</p> <p>The best available data suggests that up to 100 individuals may occur locally at the study area, which would represent <2% of the NSW population and <0.1% of the overall population (assumed to be >100,000). Acoustic data suggests that activity levels for Little Bent-winged Bat are low, and likely to be concentrated in well-timbered areas where the species is known to forage.</p> <p>Overall, although the project has the potential to result in direct mortality to a small number of individuals, however it is considered unlikely that this would cause a significant impact to the subpopulation such that it would be eliminated.</p> <p>iii. To determine if the persisting subpopulation that is fragmented will remain viable, estimate the habitat area required to support the remaining population, and habitat available within dispersal distance, and distance over which genetic exchange can occur and pollination distance for the species.</p> <p>Approximately 2.54 million hectares of native vegetation occurs within a 100 kilometre buffer zone surrounding the development footprint. These areas represent quality foraging habitat for the species which typically hunts in well-timbered areas flying between the shrub layer and canopy to catch beetles, moths, flies and spiders (DPIE 2021c, Churchill 2008). This represents the foraging habitat available with the range of the local population. Only 132.43 hectares of native vegetation occurs within the development footprint, a fraction of the vegetation within the locality. Of this vegetation approximately 37.92 hectares (28.6 %) is in high condition providing good habitat values for foraging native fauna, including Little Bent-winged Bat (<i>note these areas will be updated on final BDAR review</i>). This foraging habitat would be modified as a result of the Project.</p> <p>Little Bent-winged Bat is a highly mobile species, capable of travelling several hundreds of kilometres between roost sites (Dwyer 1968). The species uses a broad range of habitats for foraging including moist eucalypt forest, rainforest, vine thicket, wet and dry sclerophyll forest, Melaleuca swamps, dense coastal forests and banksia scrub (DPIE 2021c). While it is assumed that connected vegetation is preferred by the species for movement, the existence of many records of the species within urbanised areas, and its utilisation of man-made structures as roosting sites, suggests that the species does not rely on specialised dispersal</p>

Information required (BAM Section 9.1.2)	Response
	<p>or movement habitat.</p> <p>The local population is expected to occur on a seasonal and transient basis in response to seasonal movements, which are unlikely to be impacted as a result of the Project such that the sub-population would become fragmented or unviable. The foraging habitat available to these individuals, which are capable of travelling large distances, is unlikely to be significantly impacted as a result of the Project and no impacts on genetic exchange are anticipated given that the population is unlikely to become fragmented as a result of the Project.</p> <p>iv. To determine changes in threats affecting remaining subpopulations and habitat if the proposed impact proceeds, estimate changes in environmental factors including changes to fire regimes (frequency, severity); hydrology, pollutants; species interactions (increased competition and effects on pollinators or dispersal); fragmentation, increased edge effects, likelihood of disturbance; and disease, pathogens and parasites. Where these factors have been considered elsewhere in relation to the target species the assessor may refer to the relevant sections of the BDAR or BCAR.</p> <p>Changes to fire regimes</p> <p>The risk of fire as a result of sparks from machinery during proposed works is unlikely, but could increase the risk of fire occurring in foraging habitat or nearby potential roost sites. This risk will be managed by implementing appropriate mitigation measures such as spark dampeners, water spraying or the close proximity of fire-fighting gear such as extinguishers within the project's CEMP.</p> <p>Upon completion of works, the proposal is unlikely to result in changes to frequency or intensity of fire regimes within the locality.</p> <p>Hydrology</p> <p>Changes to hydrology as a result of the proposal are discussed in BDAR. These changes are unlikely to significantly impact on the foraging or roosting behaviour of Little Bent-winged Bat.</p> <p>Pollutants</p> <p>The project does not involve the use of any pesticides, nor is it likely to substantially increase the levels of pesticides within the environment.</p> <p>Species interactions</p> <p>Introduced predators such as feral cats and foxes can negatively impact the species by preying on bats as they exit caves, sometimes taking significant numbers. One study reported 476 Bent-winged bats (prior to the taxonomic subspecies change) being predated (P.D Dwyer 1966b). Black Rats have also been reported in maternity caves and are likely to prey upon young (P.D Dwyer 1966b, Lumsden & Jemison 2015). Such predators were recorded during camera trap surveys, are likely to already present within the broader locality, and are unlikely to increase as a result of the proposal.</p> <p>Fragmentation</p> <p>Little Bent-winged Bat is a highly mobile species capable of dispersing across breaks in habitats. While it is assumed that connected vegetation is preferred by the species for movement, the existence of many records of the species within urbanised areas suggests that the species does not rely on specialised dispersal or movement habitat. Thus, the loss of approximately 37.92 hectares (<i>note these areas will be updated on final BDAR review</i>) of native vegetation from the buffer area is unlikely to impact the movement ecology of the local population. As such population fragmentation will not occur as a result of the proposal.</p>

Information required (BAM Section 9.1.2)	Response
	<p>Increased edge effects</p> <p>The proposed works that occur within the vicinity of the identified potential Little Bent-winged Bat habitats are occurring within a previously disturbed area as a result of existing agricultural and industrial practices. As such these areas are already subject to some edge effects. Whilst transport of weeds is possible during construction, these will be mitigated through the application of appropriate weed control measures to be detailed within the CEMP. This will ensure the existing edge effects are not exacerbated by the proposal. These edge effects are also unlikely to significantly impact the species, given the availability of foraging resources within the locality.</p> <p>Likelihood of disturbance</p> <p>Disturbance will occur in the form of direct impacts to native vegetation that may remove a small amount of potential foraging habitat. There would be no disturbance to roosting or breeding habitat.</p> <p>Operational impacts may also result in disturbance to the seasonal or foraging movements of the bats, including the potential for mortalities due to collisions with turbines. However, the potential for substantial numbers of mortalities associated with turbine collision is considered to be low given the likely small size of the local population and the seasonal and transient nature of their likely occurrence in the area that would render the species less likely to encounter turbines while present in the area. Although disturbance to Little Bent-winged Bat (including collisions) is possible as a result of the Project, these impacts are likely to be restricted to a small number of individuals. Based on the factors above, it is considered unlikely that disturbance (including a small number of possible mortalities) would result in impacts at the population level.</p> <p>In designing the wind farm layout the spacing of the turbines was considered to limit interactions with fauna and an essentially linear row of turbines was designed. This is different from most wind energy facilities in Australia in which turbines are scattered across a site. In the more usual scattered or 'clustered' array, microbats and avian fauna have a high probability of encountering multiple turbines in a given flight. The configuration of Hills of Gold turbines is such that an individual is likely to encounter multiple turbines only in the rare event that it flies directly along the row of turbines. The average distance of turbines spacing across the entire site was approximately 423 metres (from rotor hub to rotor hub), which has been increased by recent layout updates. Only five turbines considered to present a moderate risk of collision or barrier effects occur are located less than 400 metres from the adjacent turbine, with a number of gaps between turbines ranging from 1.2 kilometres wide to 1,600 metres (1,400 metres at turbine height). These gaps provide movement corridors, allowing for migrating and foraging bats to pass through the landscape, with the risk of barrier effects assessed as low overall. Current turbine spacing and layout is also considered to reduce the chance of collision and possible barotrauma.</p> <p>Disease, pathogens and parasites</p> <p>An emerging threat to Australian bats, particularly cave-roosting species, is the fungal disease white-nose syndrome. To date there have been no cases of white-nose syndrome recorded in Australia however, a recent risk assessment considers it 'likely' that the pathogen causing the disease (<i>Pseudogymnoascus destructans</i>) will come into contact with Australian bats in the coming decade (Holz et al 2019).</p>

Information required (BAM Section 9.1.2)	Response
	<p>Cave-roosting bats are particularly at risk, and if the disease becomes established in Australia it is probable that Little Bent-winged Bat populations would incur some impacts although more research is required to determine the extent and severity (Turbill & Welbergen 2020). Conditions in the southern part of the species range are likely to be better suited to <i>P.destructans</i> growth (Turbill & Welbergen 2020), so the large geographic distribution of the species may provide it with some resilience in the event of the disease becoming established.</p> <p>The project would not contribute to any increased risk in the introduction of white-nose syndrome.</p>
<p>5. The assessor may also provide new information that can be used to demonstrate that the principle identifying the species as at risk of an SAIL, is inaccurate.</p>	<p>Not applicable.</p>

Large-eared Pied Bat SAI assessment

Large-eared Pied Bat is listed as Vulnerable under both the EPBC Act and BC Act. It is a member of the Vespertilionidae family (Churchill 2008), that was not formerly described until 1966 (Dwyer 1966). As a result, little is known about the species and it is considered data-deficient. The current distribution of Large-eared Pied Bat is poorly known however, records for the species occur from Shoalwater Bay, north of Rockhampton, Queensland, through to Ulladulla in southern NSW (DAWE 2020, DPIE 2021a). The species has most commonly been recorded in dry sclerophyll forests and woodlands however they do also occur in sub-alpine woodland as well as the edges of rainforest, wet sclerophyll forest, and *Callitris* sp. dominated forests, and areas with sandstone outcropping. (Churchill 2008). Sandstone cliffs and fertile wooded valley habitat within close proximity of each other is thought to be critical habitat for the species

Large-eared Pied Bat is considered a full species credit species under the BAM as it cannot be reliably predicted to occur on a site based on vegetation and other landscape features. Identified habitat constraints for the species include areas within two kilometres of rocky areas containing caves, overhangs, escarpments, outcrops, or crevices, or within two kilometres of old mines or tunnels (DPIE 2021b). The SAI threshold for the species specifically relates to potential breeding habitat and the presence of breeding habitat. The Atlas of NSW Wildlife identified these areas as follows:

Potential breeding habitat is PCTs associated with the species within 100m of rocky areas containing caves, or overhangs or crevices, cliffs or escarpments, or old mines, tunnels, culverts, derelict concrete buildings.

Amendments to the project design, as outlined in the BDAR, have avoided direct impacts to the species' potential breeding habitat, however following correspondence with BCS (April 2022), the species is still considered at risk of an SAI as a result of the operational impacts of the project. In particular, the risk of mortalities due to collisions with wind turbines.

As described in the previous SAI assessment for Large Bent-winged Bat, it is impossible to numerically quantify potential risk of collisions for Large-eared Pied Bat.

The following SAI has been prepared to address the potential operational impacts to Large-eared Pied Bat as a result of the project.

Table 106 SAI assessment for Large-eared Pied Bat

Information required (BAM Section 9.1.2)	Response
1. The action and measures taken to avoid the direct and indirect impact on the species at risk of an SAI.	Addressed above collectively for SAI bat species.
2a. Evidence of rapid decline. (Principle 1)	<p>i. Evidence of rapid decline in the population of the species in NSW in the past 10 years or three generations (whichever is longer), or</p> <p>ii. decline in population of the species in NSW in the past 10 years or three generations (whichever is longer) as indicated by: an index of abundance appropriate to the species; decline in geographic distribution and/or habitat quality; exploitation; effect of introduced species, hybridisation, pathogens, pollutants, competitors or parasites.</p> <p>Species and ecological communities that have undergone large reductions or are likely to undergo large reductions in the future are considered to be at greater risk of extinction than</p>

Information required (BAM Section 9.1.2)	Response
	<p>those that have undergone or are likely to undergo smaller reductions (NSW TSSC 2018). To be considered under this principle, the species should have an observed, estimated, inferred, suspected or projected population of $\geq 80\%$ in 10 years or three generations (whichever is longer) (DPIE 2019).</p> <p>Large-eared Pied Bat is a data-deficient species (DPIE 2021c), the former distribution of which is poorly known. The Species Profile and Threats Database (SPRAT) (DAWE 2020) indicates that whilst it has been suggested that there have been large declines in suitable habitat, it is not possible to evaluate past declines in the species extent of occurrence due for the following reasons:</p> <ul style="list-style-type: none"> • It was only formally described in 1966. • Like most insectivorous bats it is nocturnal and unobtrusive so opportunistic observations are uncommon. • Targeted surveys utilising appropriate techniques to record this species have only taken place since the 1990s (DAWE 2020). <p>Extensive surveying has been undertaken to determine maternity roost sites for the species, however only four sites have been recorded across the species' range, one of which was permanently flooded in 1976 and another one was abandoned in 2009 (DERM 2011, DAWE 2020, TSSC 2012).</p> <p>The National Recovery Plan for the Large-eared Pied Bat states that it has not yet been determined whether any specific populations of the Large-eared Pied Bat are at a higher level of threat than others, and that a better understanding of distribution, population size, roost preferences and threats is required before such populations can be identified (DERM 2011). The fact that the preferred habitat for the species comprised of suitable foraging and roosting habitat in close proximity to each other, occurs so rarely in the landscape suggests the species may have always been uncommon (DERM 2011). However, preferential clearing of fertile forests and woodlands preferred by the species has 'almost certainly' reduced the amount of available habitat.</p> <p>Based on the available information it is likely that the species has experienced some declines associated with habitat loss, and the destruction and interference of maternity and other roosts (TSSC 2012, DERM 2011). Due to the data-deficient nature of the species it is not possible to state the scale of decline in the last 10 years.</p> <p>Given the data-deficient nature of this species, consideration of the species under SAI Principle 1 is warranted.</p>
2b. Evidence of small population size. (Principle 2)	<p>i. An estimate of the species' current population size in NSW, and</p> <p>ii. An estimate of the decline in the species' population size in NSW in three years or one generation (whichever is longer), and</p> <p>iii. Where such data is available, an estimate of the number of mature individuals in each subpopulation, or the percentage of mature individuals in each subpopulation, or whether the species is likely to undergo extreme fluctuations.</p> <p>Species with small population sizes are highly vulnerable to any event which impacts and further reduces their population size due to the time-lag between developmental impacts and the realisation of ecological benefits from improvements in habitat condition at stewardship offset sites (DPIE 2019).</p> <p>To be considered under this principle a species must have a very small population size which would lead it to be considered critically endangered under the <i>IUCN Red List Categories and Criteria</i> (IUCN 2012). Specifically the species has a known population size that is either:</p>

Information required (BAM Section 9.1.2)	Response
	<ul style="list-style-type: none"> • Fewer than 50 mature individuals independent of whether there are any threats. • Fewer than 250 mature individuals and the species has an observed, estimated or projected continuing decline: <ul style="list-style-type: none"> – of at least 25% in three years or one generation (whichever is longer), or – where the number of mature individuals in each subpopulation is <50, or – the percentage of mature individuals in one subpopulation is 90–100%, or – the population is subject to extreme fluctuations in the number of individuals. <p>‘Population’ means the total number of mature individuals in New South Wales, ‘Subpopulations’ are geographically or otherwise distinct groups in the total population (DPIE 2019, IUCN Standards and Petitions Committee 2019). The species is not currently listed as critically endangered under the NSW BC Act or Commonwealth EPBC Act.</p> <p>There is currently insufficient data to estimate the abundance or population trends of the Large-eared Pied Bat (DAWE 2020). The species appears to exist in a number of small populations throughout its known range, with colonies seldom containing more than 50 individuals (TSSC 2012, DERM 2011). Four maternity roosts have been documented in NSW, though one of them has been abandoned and another flooded in 1976 (DAWE 2020). It is unclear what impact the absence of these sites has had (if any) on the population.</p> <p>The national listing advice for the species states that there is insufficient data to accurately determine the total number of mature individuals (TSSC 2012). The cryptic and often inaccessible nature of the species’ preferred roosting habitat makes it difficult to determine the number and location of roost sites in order to more accurately assess the size of the population.</p> <p>Only two studies have been undertaken on maternity caves, one north-west of Coonabarabran in central New South Wales (Pennay 2008), and one approximately 200 kilometres away in Copeton in 1962-1963 (Dwyer 1966). Both of these studies reported small population sizes, ranging from 14-40 mature adults with females typically supporting two pups. The displayed breeding behaviours at the two sites was broadly consistent despite the period of 39 years between observations (Dwyer 1966, Pennay 2008, TSSC 2012).</p> <p>Experts have suggested the species is unlikely to undergo extreme natural fluctuations in population numbers or extent of occurrence or area of occupancy, though the justification for such assertions has not been published (DAWE 2020).</p> <p>The total number of mature adults in each subpopulation is low (reportedly <50 individuals), this appears to be stable with similar subpopulation results being observed at two maternity caves approximately 200 kilometres apart over a 39 year period time lapse (Dwyer 1966, Pennay 2008, TSSC 2012).</p> <p>A total of 1,967 BioNet Atlas records for Large-eared Pied Bat exist across NSW (DPIE 2021b). Threatened bats generally have a lower detection rate than other species, and capture rates for Large-eared Pied Bat are estimated at just 1.2% of all bat captures in its range each year (OEH 2018). The species is readily identifiable from acoustic recordings due to the frequency it typically calls at and the unique pattern of call alternation it produces (Pennay, Law, & Reinhold 2004). Acoustic detection rates are also likely to be low for this species, which produces soft calls and therefore is only likely to be detected within about 10 m of the microphone (Williams & Thomson 2018).</p> <p>Although definitive data on total population numbers is not available, the number and distribution of records for the species across NSW over time, suggest that it is unlikely the Large-eared Pied Bat population across NSW consists of less than 250 mature individuals.</p>

Information required (BAM Section 9.1.2)	Response
	<p>However, given the data-deficient nature of this species, consideration of the species under SAI Principle 2 is warranted.</p>
<p>2c. Evidence of limited geographic range for the threatened species. (Principle 3)</p>	<p>The geographic range of a species is measured by its area of occupancy, which represents the area of suitable habitat currently occupied by the taxon (IUCN Standards and Petitions Committee 2019). Species that are known to have a very limited geographic distribution are generally known to:</p> <ul style="list-style-type: none"> • Have an area of occupancy of $\leq 10 \text{ km}^2$. • Have an extent of occurrence of $\leq 100 \text{ km}^2$. • Have at least two of the following three conditions: <ul style="list-style-type: none"> – Are severely fragmented or only known from one location. – Continuing decline. – Extreme fluctuations. • Inhabit less than or equal to three locations in NSW (DPIE 2019). <p>i. Extent of occurrence.</p> <p>Large-eared Pied Bat is distributed along the east-coast of Australia, ranging from Shoalwater Bay, north of Rockhampton, Queensland, through to Ulladulla in southern NSW (DPIE 2021a, DAWE 2020). Whilst the range of the species is large, it has been suggested that the species is restricted within its range, and does not occur continuously (DAWE 2020). This may be due to the lack of available roosts throughout its range and has likely resulted in the populations of the species in north-east NSW and south-east Queensland, Shoalwater Bay and Blackdown Tablelands to become isolated from each other, with little interaction (DAWE 2020).</p> <p>The Saving Our Species strategy for this species is currently listed as data-deficient, with no priority management areas currently identified for the species (DPIE 2021c). Known breeding locations are extremely limited within NSW. Five locations are known to have been used for breeding within NSW, including:</p> <ul style="list-style-type: none"> • A mine tunnel at Copeton which was used for breeding until flooded by dam waters in 1976 (Dwyer 1966). • A sandstone cave near Coonabarabran, NSW (Pennay 2008). • Capture of lactating females adjacent to sandstone cliffs in Ulan, NSW (Fly By Night 2005). • Observations of a small group of lactating females and dependent young in a disused gold mine near Barraba, NSW (DERM 2011). • Anecdotal observations of small groups of females and young bats in the sandstone Pilliga region, NSW (DERM 2011). <p>The species extent of occurrence is approximately $570,000 \text{ km}^2$, based on the distribution range in Hoye and Dwyer (1995, DAWE 2020).</p> <p>ii. Area of occupancy.</p> <p>The area of occupancy is approximately $9,120 \text{ km}^2$. This is calculated from the extent of occurrence and the detection rate of echolocation calls of 1.6 % at 3,154 site across the range of Large-eared Pied Bat (DAWE 2020).</p>

Information required (BAM Section 9.1.2)	Response
	<p>iii. Number of threat-defined locations (geographically or ecologically distinct areas in which a single threatening event may rapidly affect all species occurrences).</p> <p>The primary threat facing the species is disturbance and damage to primary nursery sites by animals and humans (DAWE 2020). The species requires cave environments of a specific structure (arch caves with dome roofs and indentations for holding) in order to breed successfully. These physical characteristics are very uncommon and their scarcity presumably poses a limiting factor in the distribution of the species (DERM 2011, Pennay 2008). The species exhibits high fidelity to these maternity cave sites, and only four such roosting sites have been formerly recorded in NSW (DAWE 2020), two of which are no longer in use. The closest of these known breeding sites to the study area is the disused gold mine near Barraba, with the town of Barraba located approximately 145 kilometres away from the development footprint.</p> <p>Sandstone cliffs and fertile wooded valley habitat within close proximity of each other should also be considered habitat critical to the survival of the species (DERM 2011). The combination of high-fertility forest near suitable roosting habitat is rare in the landscape, and clearing, harvesting or destruction of vegetation in proximity to roosts has the potential to affect foraging resources, which could be particularly detrimental to pregnant and lactating females (DERM 2011).</p> <p>Unlike other cave-roosting bat species, it appears that Large-eared Pied bat form multiple, small maternity colonies in suitable cave environments, rather than congregating in single large nursery caves like Large Bent-winged Bat or Little Bent-winged Bat. This may make the species less susceptible to single, threatening events although this is difficult to assess thoroughly without sufficient data on the number and distribution of suitable roost sites currently available or in use.</p> <p>Another emerging threat to the species is the potential introduction of the fungal disease white-nose syndrome (Holz et al. 2019), which has decimated bat populations in North America. A recent risk assessment considers it 'likely' that the pathogen causing the disease (<i>Pseudogymnoascus destructans</i>) will come into contact with bats in the coming decade in Australia (Holz et al 2019). Cave-roosting bats are particularly at risk and it is probable that Large-eared Pied Bat populations would incur some impacts should the disease become established in Australia, although more research is required to determine the extent and severity (Turbill & Welbergen 2020).</p> <p>iv. Whether the species' population is likely to undergo extreme fluctuations.</p> <p>Expert opinion is that Large-eared Pied Bat is unlikely to undergo extreme natural fluctuations in population numbers, extent of occurrence or area of occupancy, however the justification for this opinion has not been published (DAWE 2020).</p> <p>The available information on the geographic distribution of the species indicates that the species does not meet the thresholds for consideration under SAI Principle 3. Although there are only three known roost sites for the species in NSW, based on its distribution and the cryptic nature of roost sites, it is assumed that more exist that have not yet been documented in the literature.</p>
<p>2d. Evidence that the species is unlikely to respond to management.</p>	<p>This principle encompasses two components, firstly whether there are any particular traits of the species which limits its' response to management, and secondly whether there are any key threatening processes affecting the species which cannot be effectively managed. Furthermore in select areas where essential habitat components cannot be readily re-</p>

Information required (BAM Section 9.1.2)	Response
(Principle 4)	<p>created (such as caves or cliff lines used by threatened species) such impacts can be deemed irreplaceable (DPIE 2019).</p> <p>i. Known reproductive characteristics severely limit the ability to increase the existing population on, or occupy new habitat on, a biodiversity stewardship site.</p> <p>The species requires highly specific maternity roosting sites consisting of arched cave environments with dome roofs and indentations for holding in order to breed successfully. These physical characteristics are uncommon and their scarcity presumably poses a limiting factor on the distribution of the species (DERM 2011, Pennay 2008). The species exhibits high site fidelity to these maternity cave sites, and only four such roosting sites have been formerly recorded in NSW (DAWE 2020). Additionally, the species appears to require fertile wooded valley habitat in close proximity to roost sites. These features are unlikely to be replicated successfully on a stewardship site, where such features are not already naturally occurring.</p> <p>ii. The species is reliant on abiotic habitats which cannot be restored or replaced on a biodiversity stewardship site.</p> <p>In addition to maternity roost sites, the species also requires sandstone cliff/escarpments habitats (roosting habitat) in close proximity to fertile woodland valley habitats (foraging habitat), particularly box gum woodlands or river/rainforest corridors (DAWE 2020). These features are unlikely to be replicated successfully on a stewardship site, where such features are not already naturally occurring.</p> <p>iii. Life history traits and/or ecology is known but the ability to control key threatening processes at a biodiversity stewardship site is currently negligible.</p> <p>The primary threat to the species is loss or degradation of roosting habitat and maternity roost site as a result of animals and human activity. Where suitable roosting habitats occur within a biodiversity stewardship site, effective management of such features can be readily achieved.</p> <p>Secondary threats include loss of suitable foraging resources in proximity to roosting sites due to vegetation clearing and inappropriate fire regimes. Both of these can be effectively managed at a biodiversity stewardship site.</p> <p>Another emerging threat is the potential for populations to experience declines with the introduction of the fungal pathogen causing white-nose syndrome. Controlling this on a biodiversity stewardship site may be possible with strict hygiene controls and restricted access to caves.</p> <p>Given the highly specialised breeding and roosting habitat requirements for the species, it is unlikely such habitats can be readily replicated at a biodiversity stewardship site. As such consideration of this species under SAI Principle 4 is warranted.</p>
3. Where the TBDC indicates data is 'unknown' or 'data deficient' for a species, the assessor must record this in the BDAR or BCAR.	The TBDC does not state that data is 'unknown' or 'data deficient' for this species.
4a. The impact on the species' population	i. An estimate of the number of individuals (mature and immature) present in the subpopulation on the subject land (the site may intersect or encompass the subpopulation) and as a percentage of the total NSW population, and,

Information required (BAM Section 9.1.2)	Response
	<p>ii. An estimate of the number of individuals (mature and immature) to be impacted by the proposal and as a percentage of the total NSW population.</p> <p>There is very little population data available for this species. The size of the population throughout its range is uncertain, and in NSW only two maternity caves are currently known for the species, both consisting of <50 individuals. Historical survey coverage in the project's assessment area (and subject land particularly) is low due to its rural location. The absence of local information is compounded by the lack of existing scientific literature of the species' population biology, making it difficult to estimate both the size of the local population and the NSW population as a whole. This assessment utilises the available information regarding the roosting and foraging behaviour of the species to determine the likely local population relevant to, and potentially impacted by, the Project.</p> <p>Large-eared Pied Bat is known to form small colonies (DERM 2011). It has been suggested that Large-eared Pied Bat does not occur continuously within its range, likely due to the lack of available roosts that may have resulted in colonies or populations becoming isolated from each other. The largest known populations of the Large-eared Pied Bat occurs in areas dominated by sandstone escarpments, which in NSW is in the Sydney basin and northwest slopes (DERM 2011). This is assumed to be where the species density would be highest. Large-eared Pied Bat has been found to preferentially roost in clifflines facing south-west to north-west, which is thought to confer thermoregulatory benefits to the roosting bats (Williams & Thomson 2018). All day roosts have been found to occur primarily in flat, vertical sandstone rockfaces overlooking preferred foraging areas with only small fissures and no cave or cavern development (Williams & Thomson 2018). None of these habitat types were observed within the subject land. The underlying geology is volcanic, comprised of basalt rock formations (Environmental Geosurveys Pty Ltd 2021). No significant caves in the subject land have been identified, although there remains potential for smaller caves, cracks and crevices to occur that may provide roosting opportunities for Large-eared Pied Bat.</p> <p>Two of the four known roost sites (Copeton and Barraba) for the species, are associated with mine tunnels. The closest known breeding site to the project is a disused gold mine near Barraba, with the town of Barraba located approximately 145 kilometres away from the development footprint, well outside the known commuting range for this species. Observations have also been made of Large-eared Pied Bats roosting in Fairy Martin nests, which may be available in and around the study area (Churchill 2008).</p> <p>Potential roosting sites may occur around the nearby town of Nundle, where there are many (dozens) of historic mines and unnamed pits associated with the town's rich mining history (AUSGIN 2021). Studies have shown that the species prefers roosting locations within 700 metres of foraging habitats, with the maximum distance travelled between a known roost and the edge of preferred foraging areas being 2.5 kilometres (Williams & Thomson 2018). It is unlikely that individuals utilising available habitat around Nundle would be making regular use of the subject land for foraging.</p> <p>In the absence of known roosting sites, acoustic records are the best available information on the local presence of Large-eared Pied Bat. Historic records occur approximately 14 kilometres south-west of the wind farm development footprint around the Timor Cave area in December 2018 (DPIE 2021b). Targeted surveys supporting the current assessment also recorded Large-eared Pied Bat. These surveys were conducted from February 2020 to May 2020 and included deployment of units at ground-level as well as deployment on three meteorological masts at heights of 2 metres, c.30 metres and c.60 metres. Additional acoustic/ultrasonic survey was also undertaken in spring 2019 between 19 – 21 November</p>

Information required (BAM Section 9.1.2)	Response
	<p>2019 (three nights) across the 24 deployed ultrasonic detector units deployed in 2020 and the three nights of spring surveys in 2019, the total mean call for the species per night was 1.0 call, indicating a low level of activity within the development footprint on an infrequent basis. The maximum number of calls attributable to this species in one night was 19. Due to limitations of acoustic detections, it is difficult to infer abundance from this data.</p> <p>Given the absence of preferred roosting habitat (sandstone cliff faces) within 2.5 km of the development footprint, it is assumed that the local population is likely to be small, possibly utilising disused mine shafts in the area or other unidentified cracks or crevices. There are large tracts of intact native vegetation within and adjacent to the study area such as the Ben Halls Gap National Park, Crawney Pass National Park, Wallabadah Nature Reserve and the Nundle and Hanging Rock State Forests which are connected to the development footprint. These areas may provide foraging or dispersal pathways for Large-eared Pied Bat, assuming suitable roost sites are available.</p> <p>This hypothesis is supported by the acoustic data recorded during the current assessment that found mean call activity attributable to this species to be low. Large-eared Pied Bat is known to show high fidelity to preferred foraging locations, and it is assumed that acoustic detection rates would have been higher if the species was making regular use of foraging habitats within the subject land. Based on the available data, it is assumed that the study area does not represent an area of high population density for Large-eared Pied Bat. This assessment conservatively estimates that the local population of Large-eared Pied bat would consist of one colony of 50 individuals or fewer. It is unclear what percentage of the NSW population this would represent, given the absence of population data for the species.</p> <p>Large-eared Pied Bat wing morphology predicts it is most likely to show slow, highly agile flight that allows it to forage most successfully around but not within vegetation clutter. This is supported by recent observations that found the species to show slow, fluttery flight, and forage in the mid to upper canopy along the forest edge or around tree crowns on the slopes (Williams & Thomson 2018). Preferred foraging areas were found to contain diverse vegetation types (grassland, dry and wet forest) and nearby water bodies to provide conditions that support increased invertebrate density and diversity (Williams & Thomson 2018). Recent studies found Large-eared Pied Bat activity to be 24 times greater at the forest edge than in grassland and forest (Williams & Thomson 2018), suggesting strong preferences for this species to forage at ground level in association with forest edges.</p> <p>Furthermore, the species is thought to show strong fidelity to preferred foraging locations, commuting less than 3km between roosts and foraging habitat. The species is thought to preferentially forage in and around forest edges, and has been observed to be adversely affected by high wind speeds where it has been apparently 'blown' off course (Williams & Thomson 2018). The species is known to be capable of commuting at fast speed between roost sites and preferred foraging areas (Williams & Thomson 2018). As some detections for this species occurred at height, it is possible that these represented commuting rather than foraging bats.</p> <p>Foraging behaviour, wing morphology and short commuting distances shown by this species suggest that collision risk to Large-eared Pied Bat is likely to be low. In addition to the consideration that has been made with regards to the spacing of wind turbines, leading to preservation of movement corridors, it is unlikely that the movement of individuals would be significantly restricted by the proposed development.</p> <p>Overall, although collision mortality is likely to be low, prior to the implementation of the project's mitigation strategy it is considered a moderate risk that loss of individuals as a</p>

Information required (BAM Section 9.1.2)	Response
	<p>result of collision mortality may cause change to the local abundance of the species in the short term (up to 5 years). However, given the estimated low numbers of individuals present within the subject land, which equate to a commensurately low percentage of the species' NSW population, potential impacts associated with collision with turbines are considered likely to equate to a very small percentage of the total NSW population of the species. Efforts committed to by the Proponent to minimise the potential for collision with turbines, and hence minimise the risk of the project resulting in an operational SAIL to the species are presented above in Section 4.2.</p>
4b. Impact on geographic range.	<p>i. The area of the species' geographic range to be impacted by the proposal in hectares, and a percentage of the total AOO, or EOO within NSW.</p> <p>The area to be impacted represents <0.001 % of the 570,000 km² extent of occurrence for the species, and <0.001 % of the 9,120 km² area of occurrence for the species, based on the estimates provided in the species' SPRAT profile (DAWE 2020b).</p> <p>However, it should be noted that these measures do not account for the highly specialised breeding habitats required by the species, which are very uncommon across the species' area of occurrence.</p> <p>ii. The impact on the subpopulation as either: all individuals will be impacted (subpopulation eliminated); OR impact will affect some individuals and habitat; OR impact will affect some habitat, but no individuals of the species will be directly impacted.</p> <p>The proposed impacts are unlikely to affect potential habitat for the species. No roosting habitat will be impacted.</p> <p>Direct mortality to individuals in the sub-population is considered unlikely given that the size of the local population is likely to be small, the species is not known to commute over large distances and the foraging behaviour of the species shows strong preferences for foraging around vegetated forest edges.</p> <p>Additionally, the species is known to roost in close proximity to foraging sites (<3 kilometres). Preferred roosting habitat (sandstone escarpments) does not occur in the subject land, and it is assumed that the species either utilises unidentified cracks and crevices, disused mine shafts or potentially fairy martin nests to roost in the locality where present. Acoustic data recorded low activity in the site.</p> <p>Overall it is considered unlikely that the project would result in direct mortality to individuals of this species.</p> <p>iii. To determine if the persisting subpopulation that is fragmented will remain viable, estimate the habitat area required to support the remaining population, and habitat available within dispersal distance, and distance over which genetic exchange can occur and pollination distance for the species.</p> <p>It has been suggested that Large-eared Pied Bat does not occur continuously within its range and that colonies or populations already occur in isolation from each other (DERM 2011). Individual colonies have been shown to forage in close proximity (<3km) from roosts, and it is unclear over what distance individuals are capable of dispersing. The project would not result in any impacts to roosting habitat. It is therefore unlikely to result in fragmentation of the existing subpopulation.</p> <p>Approximately 2.54 million hectares of native vegetation occurs within a 100 kilometre buffer zone surrounding the development footprint. These areas represent quality foraging habitat for the species which typically hunts along forest edges. Only 132.43 hectares of</p>

Information required (BAM Section 9.1.2)	Response
	<p>native vegetation occurs within the development footprint, a fraction of the vegetation within the locality. Of this vegetation approximately 37.92 hectares (28.6 %) is in high condition providing good habitat values for foraging native fauna, including Large-eared Pied Bat (<i>note these areas will be updated in the final BDAR review</i>). This foraging habitat would be modified as a result of the Project but is considered unlikely to impact the viability of the species within the locality.</p> <p>iv. To determine changes in threats affecting remaining subpopulations and habitat if the proposed impact proceeds, estimate changes in environmental factors including changes to fire regimes (frequency, severity); hydrology, pollutants; species interactions (increased competition and effects on pollinators or dispersal); fragmentation, increased edge effects, likelihood of disturbance; and disease, pathogens and parasites. Where these factors have been considered elsewhere in relation to the target species the assessor may refer to the relevant sections of the BDAR or BCAR.</p> <p>Changes to fire regimes</p> <p>The risk of fire as a result of sparks from machinery during proposed works is unlikely, but could increase the risk of fire occurring in foraging habitat or nearby potential roost sites. This risk will be managed by implementing appropriate mitigation measures such as spark dampeners, water spraying or the close proximity of fire-fighting gear such as extinguishers within the project's CEMP.</p> <p>Upon completion of works, the proposal is unlikely to result in changes to frequency or intensity of fire regimes within the locality.</p> <p>Hydrology</p> <p>Changes to hydrology as a result of the proposal are discussed in the BDAR. These changes are unlikely to significantly impact on the foraging or roosting behaviour of Large-eared Pied Bat.</p> <p>Pollutants</p> <p>The project does not involve the use of any pesticides, nor is it likely to substantially increase the levels of pesticides within the environment.</p> <p>Species interactions</p> <p>The impact area and surrounds likely support several pest animal species. The nature of the Project is unlikely to result in an increase in feral animal activity in the area, or alter the existing disturbances to roosting sites that may already be exhibited by feral species within the locality.</p> <p>Fragmentation</p> <p>The project will not result in the removal of any roosting habitat. Large-eared Pied Bat is capable of commuting across breaks in habitats and preferentially forages along forest edges. Thus, the loss of approximately 37.92 hectares of native vegetation from the buffer area is unlikely to impact the movement ecology of the local population. As such population fragmentation is unlikely to occur as a result of the proposal.</p> <p>Increased edge effects</p> <p>The proposed works that occur within the vicinity of the potential Large-eared Pied Bat habitats are occurring within a previously disturbed area as a result of existing agricultural and industrial practices. As such these areas are already subject to some edge effects. Whilst transport of weeds is possible during construction, these will be mitigated through the application of appropriate weed control measures to be detailed within the CEMP. This will ensure the existing edge effects are not exacerbated by the proposal. These edge effects are</p>

Information required (BAM Section 9.1.2)	Response
	<p>also unlikely to significantly impact the species, given the availability of foraging resources within the locality.</p> <p>Likelihood of disturbance</p> <p>Disturbance will occur in the form of direct impacts to native vegetation that may remove a small amount of potential foraging habitat. Operational impacts, including the potential for mortalities due to collisions with turbines, are considered low for this species. This is due to the likely small size of the local population, the species is not known to commute over large distances and the foraging behaviour of the species shows strong preferences for foraging around vegetated forest edges.</p> <p>Additionally, the species is known to roost in close proximity to foraging sites (<3 kilometres). Preferred roosting habitat (sandstone escarpments) does not occur in the study area, and it is assumed that the species either utilises unidentified cracks and crevices, disused mine shafts or potentially fairy martin nests to roost in the locality where present.</p> <p>Acoustic data recorded low activity in the site, suggesting the study area does not support preferred foraging habitat for the species.</p> <p>Overall it is considered unlikely that the project would result in direct mortality to individuals of this species.</p> <p>Disease, pathogens and parasites</p> <p>An emerging threat to Australian bats, particularly cave-roosting species, is the fungal disease white-nose syndrome. To date there have been no cases of white-nose syndrome recorded in Australia however, a recent risk assessment considers it 'likely' that the pathogen causing the disease (<i>Pseudogymnoascus destructans</i>) will come into contact with Australian bats in the coming decade (Holz et al 2019).</p> <p>Cave-roosting bats are particularly at risk, leaving them more susceptible to developing white-nose syndrome in the event it enters Australia (Turbill & Welbergen 2020). Should the disease become established in Australia, it is probable that Large-eared Pied Bat populations would incur some impacts although more research is required to determine the extent and severity (Turbill & Welbergen 2020).</p> <p>The project would not contribute to any increased risk in the introduction of white-nose syndrome.</p>
<p>5. The assessor may also provide new information that can be used to demonstrate that the principle identifying the species as at risk of an SAIL, is inaccurate.</p>	<p>Not applicable.</p>

Appendix F Geomorphology, ecology and potential microbat roosting habitat (Environmental Geosurveys Pty Ltd)

HILLS OF GOLD WIND FARM

GEOMORPHOLOGY AND GEOLOGY

&

POTENTIAL MICROBAT ROOSTING HABITAT

Prepared by: ENVIRONMENTAL GEOSURVEYS Pty Ltd (Neville Rosengren)

Prepared for: BIOSIS Pty Ltd.

VERSION 2: April 2021

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1 INTRODUCTION

1.1 Scope and Purpose

This short report was prepared in response to a letter brief (e-mail March # 2021) from Callan Wharfe (Senior Ecologist, Biosis Pty Ltd) for a desktop assessment of the geology and geomorphology at and enclosing the proposed Hills of Gold Wind Farm (HGWF) in the Mount Royal and Liverpool Ranges 60 km south-east of Tamworth in northeast New South Wales (Figure 1).

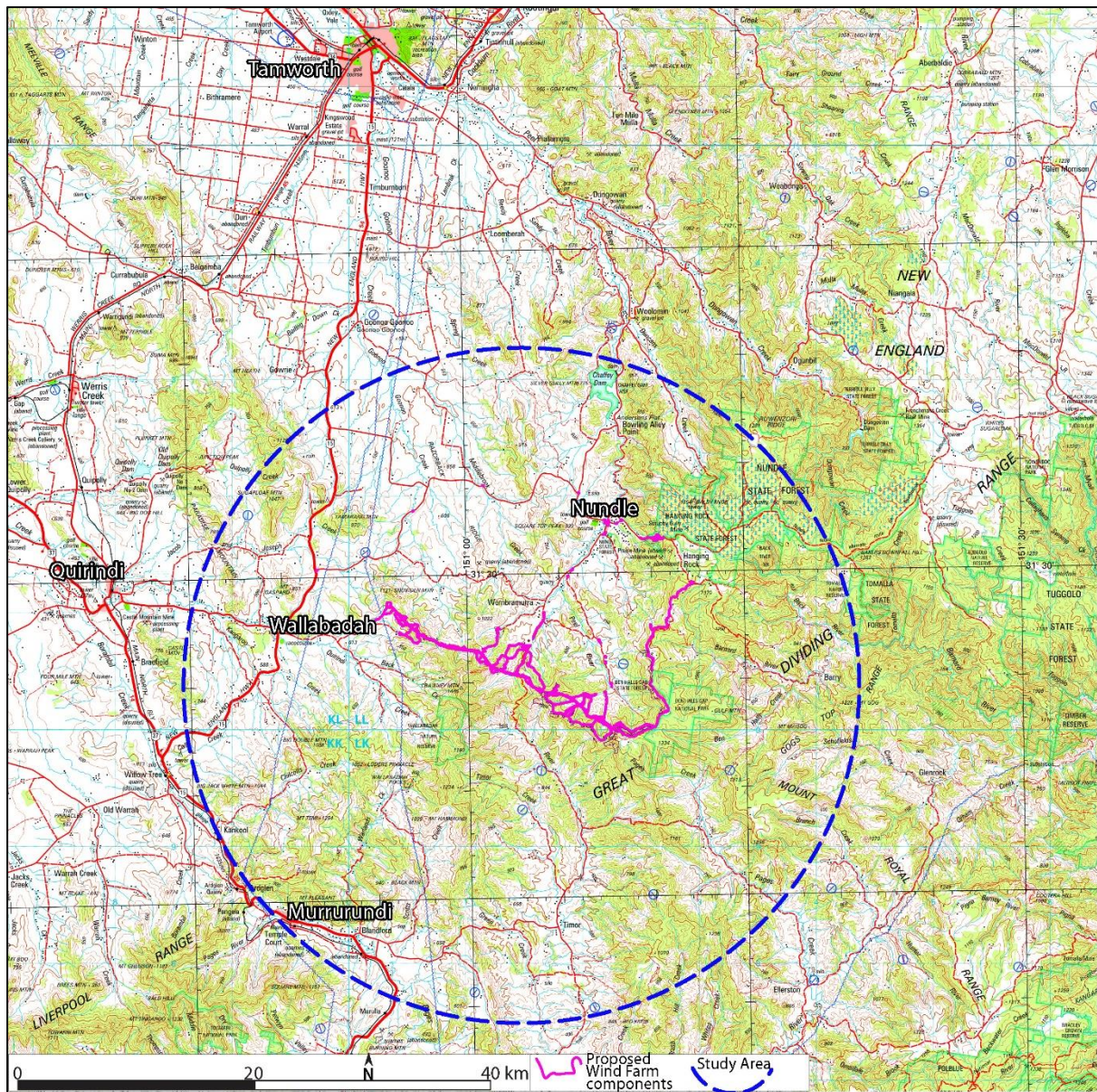


Figure 1. Location of proposed wind farm and extended study area. Base map part of Tamworth 1:250,000 Topographical Map. (Map Source: Geoscience Australia).

The purpose of the report—as requested in the letter brief—is to provide an opinion on the likelihood of the geology and landform of the area as potential microbat roosting habitat. Such habitat are cavities large enough to provide bat colony roosting space and with external access dimensions sufficient to allow flight entry and exit. The cavities may be

inherent as *in situ* attributes of the geology or produced by subaerial and subsurface weathering and denudation (erosion), including rock piles accumulated from landslides or other mass movement processes. Habitat created by human activities such as mining and quarry sites may also provide habitat.

1.2 Study Area

The study area includes precinct of the proposed HOGWF (turbines, transmission lines, roads and tracks), and a region extending approximately 30 km beyond these (Figure 1). The HOGWF turbines will be located on ridges and spurs from east of Crawney Pass to the south of Hanging Rock (Figure 2). Associated components will extend north and west of the turbine field.

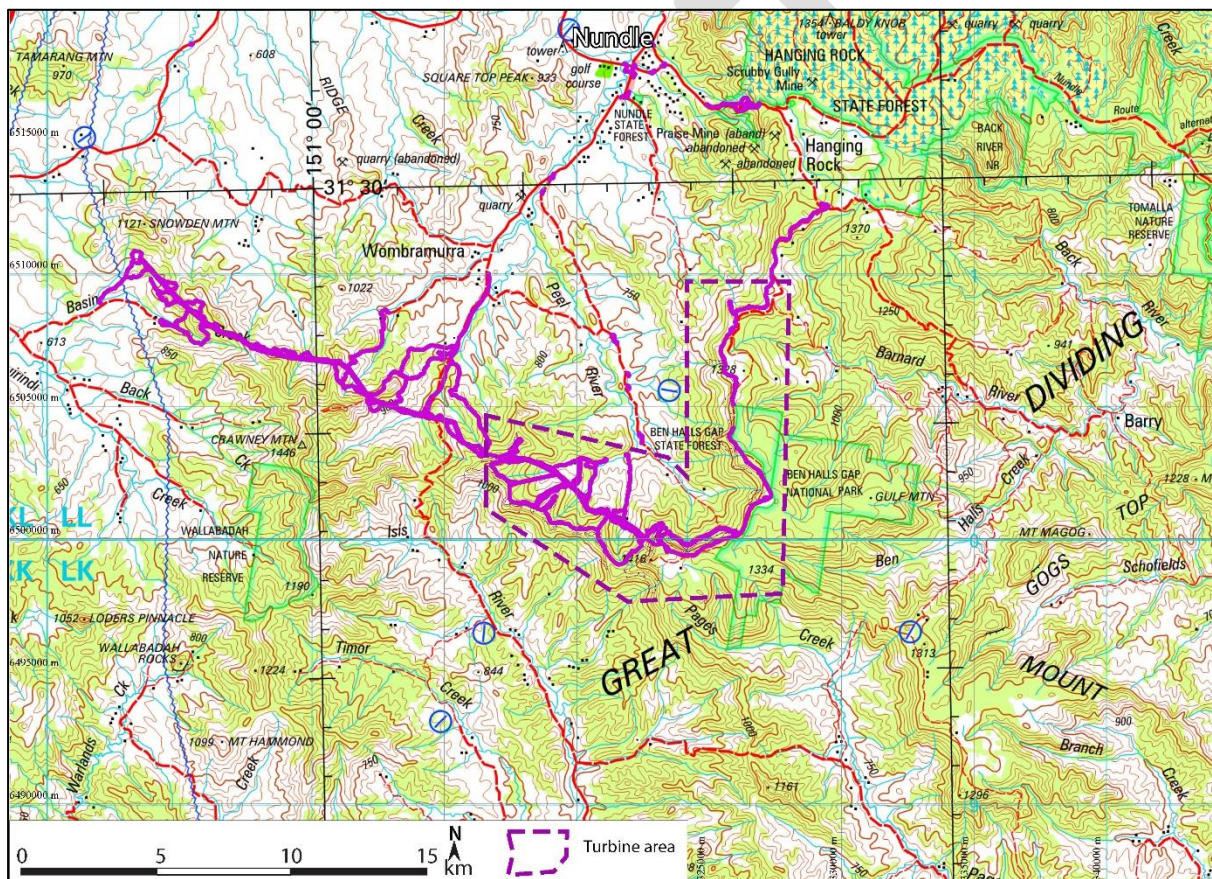


Figure 2. Topography and location of components of proposed wind farm. Base map part of Tamworth 1:250,000 Topographical Map. (Map Source: Geoscience Australia).

1.3 Methodology

This report was prepared from literature, spatial data (topography, geology) and aerial photography. Topographic data (Point Cloud LiDAR) and orthophoto imagery supplied by Someva and Biosis Pty Ltd, was supplemented by data downloaded from Geoscience Australia, Geological Survey of New South Wales, the SIX Maps mapping tool, and other products from Spatial Services New South Wales.

The GIS data was displayed on Global Mapper V22.1 LiDAR module. A DEM of the immediate study area (turbines and access tracks) was constructed from LiDAR Keypoints files with a drape of the orthophotograph and overlay of New South Wales digital Seamless Geology 2020. DEMs of turbine locations were created from LiDAR point cloud files. Combinations of slope shading and elevation shading at screen scales between 1:500 and 1:1,000 were used to identify potential habitat based on photo-interpretation of rock outcrop and scree slopes.

To provide context for potential habitat surrounding the proposed HOGWF site by identifying areas of outcrop, a DEM with 30 km buffer from the turbine sites was constructed from 5 metre LiDAR data obtained from Geoscience Australia ELVIS portal. An overlay of aerial photography of this envelope was accessed via Global Mapper allowing display of the DEM, photography, and geology. This image was viewed at screen scale approximately 1:4,000. Initially areas of slope $>20^{\circ}$ were identified on the DEM and then the photograph layer displayed to discern visible outcrop. These were outlined and the layer exported and displayed in Figure 8 in the report. Areas of steep slope where outcrop was possible—but ground could not be seen due to tree cover or shadow—were also selected as potential habitat.

2 GEOLOGY AND GEOMORPHOLOGY

2.1 Context: New England Orogen

The study area lies in the southern margin of the New England Orogen (NEO) the youngest and easternmost of the eastern Australian Tasmanides¹. The geological history of the NEO is complex resulting in a wide range of rock types and structures in northeast New South Wales and southeast Queensland. The basement rocks of the NEO developed from the Cambrian—540 million years ago (Ma) to the Carboniferous—320 Ma, and over time have been altered by multiple episodes of deformation and igneous intrusions, and further sedimentation during the Permian (300 Ma) to (Triassic-Jurassic 250 to 16 Ma) prior to the break-up of Gondwana (Scheibner 1976) (Figure 3).

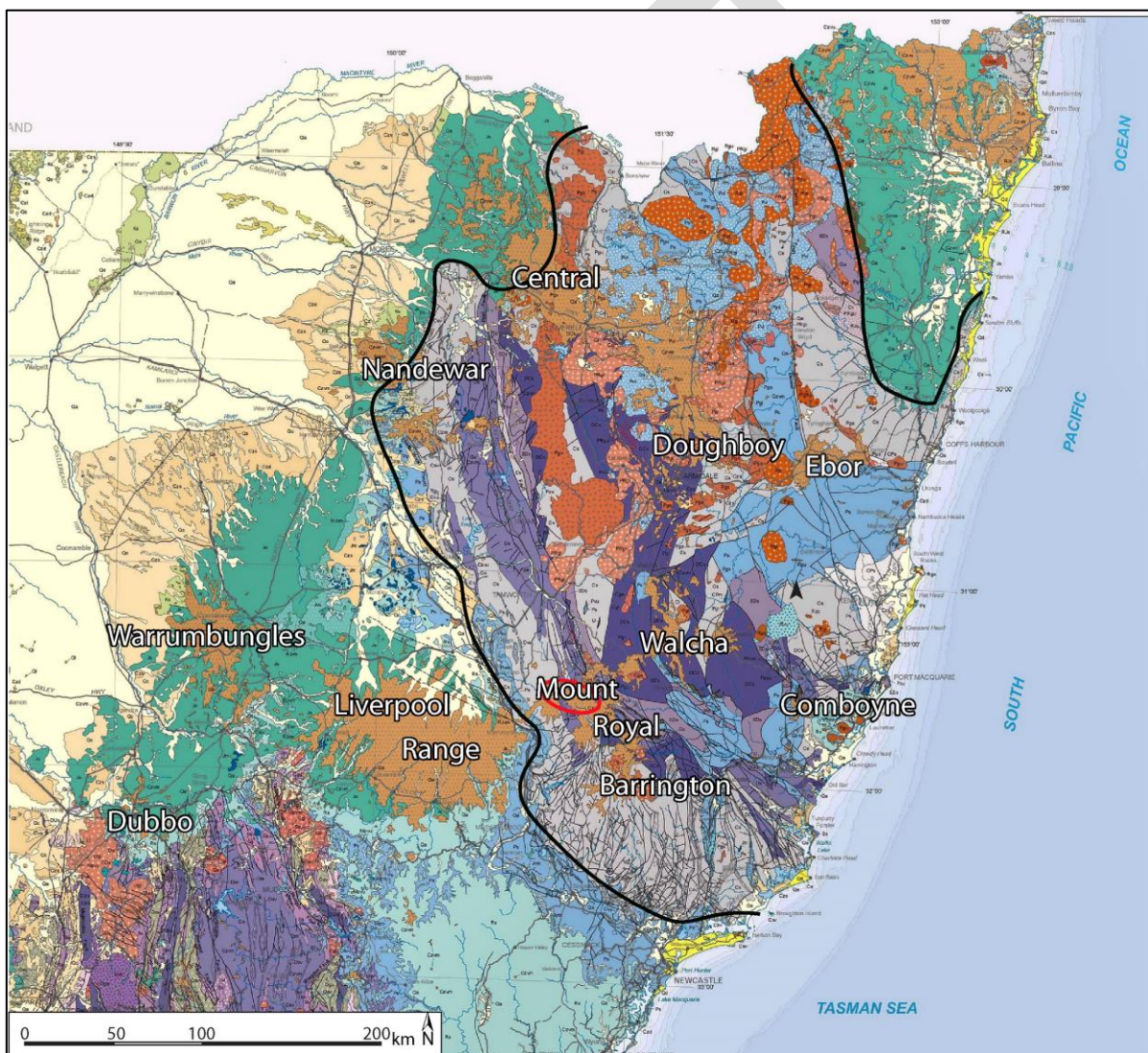


Figure 3. Surface geology and volcanic provinces of the New England Orogen (heavy black line) in northeast New South Wales. Proposed HOGWF shown as red ellipse at Mount Royal. (Province names after Kutson *et al.* (1989) From: Surface Geology of New South Wales 1:1 500 000, Planning Industry and Environment NSW 2009).

¹ The Tasmanides are geological units that form the basement rocks of eastern Australia. The New England Orogen is also referred to as the New England Fold Belt.

The exposed basement geology includes stratified silicic and calcareous sedimentary beds, granitoids and interbedded volcanics previously subjected to and altered by metamorphism, large-scale folding and faulting (Rosenbaum *et al.* 2005). Widespread volcanism in the Cenozoic covered much of the surface of the NEO, but the areal extent of volcanics has been reduced by denudation.

2.2 Cenozoic and Mesozoic Volcanics

Across much of the NEO the basement rocks are unconformably overlain by late Mesozoic and Cenozoic volcanic rocks (less than 80 Ma). Extensive remnants of eruption points (the volcano) and lava flows are conspicuous features of the modern landscape of northeast New South Wales (Figure 3, Figure 4).

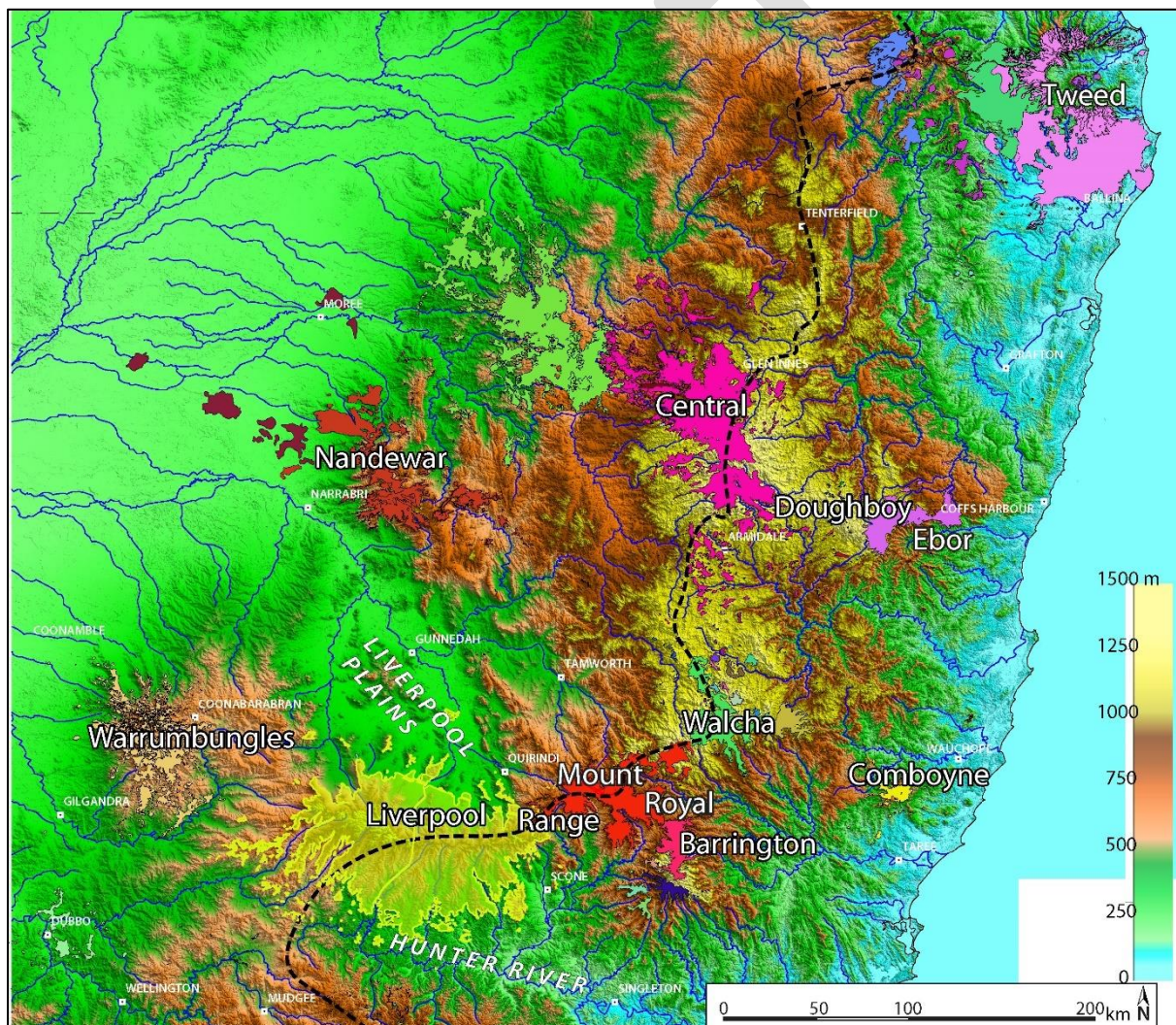


Figure 4. Digital elevation model northeast NSW showing Cenozoic volcanic provinces and position of the Great Divide (broken black line). (Volcanic province names from Knutson *et al.* (1989) and Seamless Geology NSW 2018. Base map 1 sec DEM: Geoscience Australia 2013).

Volcanoes are classified according to a range of overlapping characteristics including

- 1, *eruption source*—localised vents and fissures or numerous widespread eruption locations;
- 2, *eruption styles*—effusive with coherent lava flows or explosive with fragmental-

pyroclastic deposits; 3, *geochemistry*—fraction of silica that determines mineralogy; 4, *spatial and temporal distribution*—the extent, duration and the continuity of eruptions; 5, *geomorphic representation*—position in present landscape and preservation of initial volcanic characteristics.

Wilkinson (1969) grouped the volcanic rocks of north-eastern NSW into eight geographical provinces and recognised the petrographic (compositional) associations that occurred in some of these. Later studies, supported by radiometric dating (Wellman and McDougall 1974a, 1974b) refined the distinction based on geochemical affinity and age, and defined eleven volcanic provinces ranging from 70 Ma to 12 Ma. Recent mapping (Colquhoun *et al.* 2020) extends this to 12 provinces (Figure 3, Figure 4).

Wellman and McDougall (1974) and Knutson (in Johnson 1989) further classified the volcanic provinces into central volcanoes and lava fields. *Central volcanoes* are large single vent or clusters of close-spaced vents with frequent basaltic and felsic (higher silica) lava flows building a thick pile referred to as a shield volcano up to 100 km across. The Tweed, Warrumbungle, Nandewar, Ebor, Comboyne, Liverpool and Barrington provinces were identified as central (shield) volcanoes. Paine (1983) showed the Barrington Tops shield volcano to be at least 400 m thick and consist of 20 lava flows. *Lava fields* are provinces with eruptions from a diffuse pipe and dyke swarm producing generally thin lava flows, but in places with lava piles hundreds of metres thick. Wellman and McDougall (1974a) identified the Central Province as a lava field.

2.3 Volcanic Terrain

The original lava thickness was determined by the lava viscosity, rate of effusion and pre-volcanic relief, with greatest thickness adjacent to the eruption point and in now-buried valleys and depressions. The thickness and distribution to-day is a function of denudation and tectonic history. Three groups of terrain occur across the volcanic provinces: planar surfaces, basalt margins, sub-basaltic (exhumed) terrain.

Planar surfaces are the undulating plains formed on the denuded surfaces of lava flows. The surface is the product of weathering and erosion over millions of years and does not represent the original post-eruptive landform materials or morphology. Much of the volcanic terrain occurs along the Great Divide—a continuous topographically variable watershed in the Eastern Uplands of Australia—commonly referred to as the Great Dividing Range. The Great Divide trends north-south along the northern New England Fold Belt to 60 km south of Walcha where it turns southwest and west as the Liverpool Range headwater

tributaries of the Hunter River and Peel River with ridge and plateau summits rising to over 1400 metres (Figure 4). As the volcanism preceded and accompanied the uplift of much of the Eastern Uplands of Australia, volcanic rocks occur at a range of elevations, in places as broad elevated plateau and elongated ridges on the crest and flanks of the Great Divide.

Basalt margins are now the eroded edges of the original lava flows and are either continuous enclosures of the largest lava bodies across or adjacent to the Great Divide or outlying detached remnants—typically as hill capping. The basalt is geologically unconformable with the underlying older rocks and the base is commonly marked by a defined slope change or escarpment with locally precipitous slopes.

Sub-basaltic terrain surfaces are defined by the range in elevations of the basalt base contact with pre-volcanic rock. For example, a granite outcrop along the Main Divide at Morrisons Gap at an elevation of 1,260 m ASL is surrounded by basalt that has a base elevation around 900 m ASL in Nundle Creek, illustrating around 300 m of local relief. Paine (1983) mapped a series of bedrock benches occurring as plateau surfaces or flattened ridge crests at the base of the basalt around Barrington volcano, demonstrating pre-volcanic relief of about 1,000 m.

Weathering profiles are locally preserved in the buried-exhumed basement rocks. The basement rocks are exposed as narrow zones striking north-north-east locally defining strike ridges and valley alignment and spacing. There is variable rock outcrop and locally thick accumulations of colluvial rock debris.

2.3.1 Walcha, Mount Royal Range, Liverpool Range, Barrington Volcanic Provinces

The Walcha, Mount Royal Range, Liverpool Range and Barrington volcanic provinces are a geographically continuous surface of volcanic rocks extending 150 km south and west of Walcha and bracketing the Main Divide (Figure 5). Wellman and McDougall (1974), and Schön (1989) showed the Liverpool Range Volcanics Province as all the Liverpool Ranges south and west of Whites Sugarloaf to west of Murrurundi to Coolah Tops (Figure 6). The most recent geological mapping (Colquhoun *et al.* 2020) separates the Liverpool Range volcanics to be only west of Murrurundi—Liverpool East Basalt and Liverpool West Basalt—and recognises a new volcanic sub-province—Mount Royal Volcanics—as all the volcanics south of Whites Sugarloaf along the Main Divide to Crawney Mountain and the high southwest Liverpool Range ridge of Lagoon Mountain, Mount Temi and Mount Helen.

The newly-defined Mount Royal Volcanic complex encloses the ridges and hills of the proposed HOGWF turbine field and associated infrastructure (Figure 5).

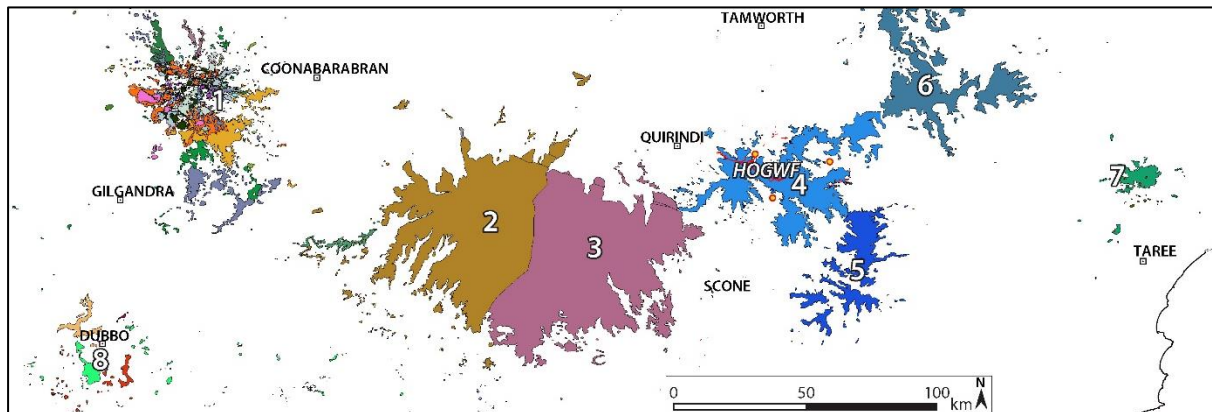


Figure 5. Distribution of volcanic units approximately 100 km radius of proposed HOGWF. 1, Warrumbungle Volcanic Complex; 2, Liverpool Range Volcanic Complex – Liverpool West Basalt; 3, Liverpool Range Volcanic Complex – Liverpool East Basalt; 4, Mount Royal Volcanic Complex; 5, Barrington Volcanic Complex; 6, Walcha Volcanic Complex; 7, Comboyne Volcano; 8, Dubbo Volcanic Complex.

The relationship between volcanic outcrop and topography in the region of the HOGWF is shown in Figure 6. The Mount Royal Volcanic province accumulated over Palaeozoic sediment at elevations between 600 m and 900 m ASL. The remnant planar basalt surface along the divide reaches elevations to nearly 1,500 m along the Great Divide indicating lava thickness in places of almost 1,000 metres, in accordance with the figure of 850 m determined by Wellman and McDougall (1974b).

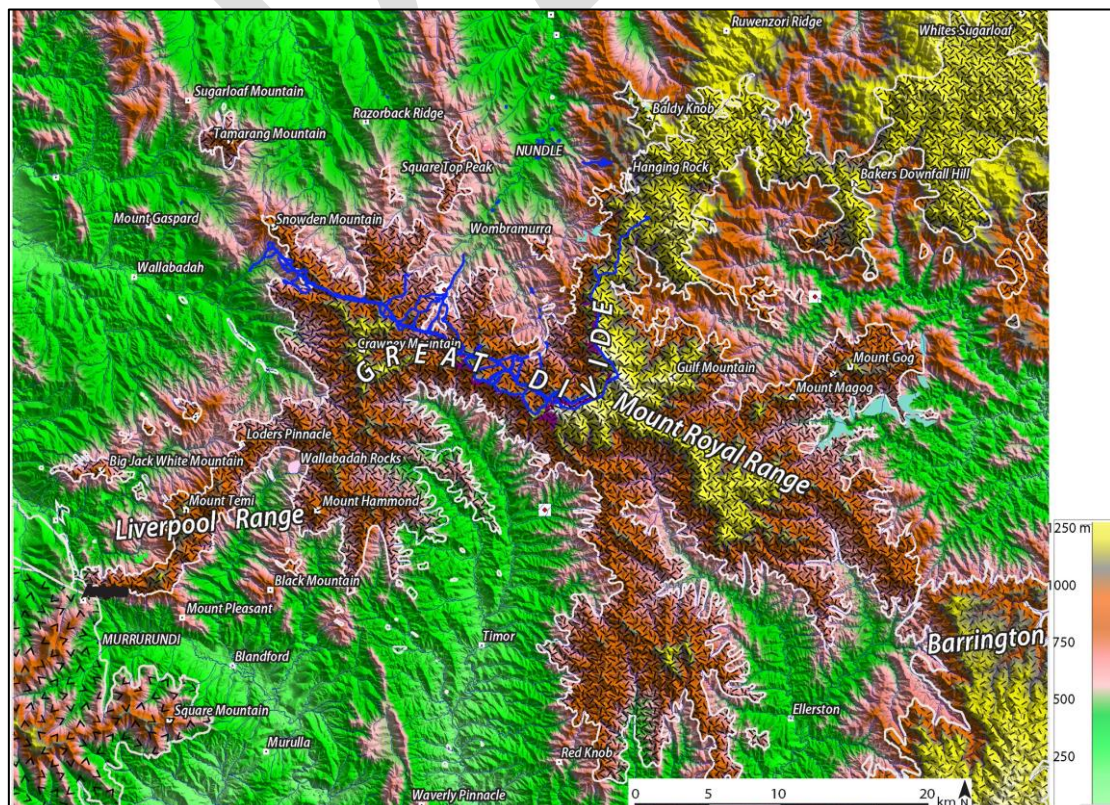


Figure 6. Mount Royal Range and Great Divide volcanic rock distribution. Mount Royal Volcanic Province extends from Whites Sugarloaf to Murrumbidgee. Base of basalt shown by white lines. (Base: 5 m DEM).

3 POTENTIAL FOR MICROBAT HABITAT

3.1 Potential Habitat Types

The occurrence of suitable microbat habitat is determined by the presence and persistence of spaces in a coherent rock mass or in accumulations of detached rock clasts. The extent to which either of these niches is present and suitable is a function initially of lithology and rock structure modified over time by geological and environmental processes that can increase or decrease the available space. As noted in Section 1.1 above, suitable space may be due to inherent attributes of the rock mass or produced by subsequent processes.

A broad initial classification of types and origins of habitat in rock groups that are likely to crop out within flight distance of the HOGWF is outlined in Table 1.

TABLE 1. Potential Microbat Habitat by Geology

TYPE	TYPE OF ROCK	SPACE	ORIGIN of SPACE	OPPORTUNITY	CONSTRAINT	MODIFICATION
Inherent structure	Fissile sedimentary. Slate, phyllite, shale, laminated mudstone	Narrow fissure	Partings along planar surfaces -bedding, cleavage, faults, shrinkage joints.	Multiple and close-spaced planes	Narrow, linear, discontinuous, unstable, soft, smooth surface.	Inherently soft rock. Change by failure along slip planes. Weathering.
Inherent structure & weathering	Interbedded, cross-bedded	Variable, overhang, small cavern	Selective failure and weathering of different lithologies.	Overhangs or small cavern	Potentially unstable roof overhangs.	Enlarged by weathering and collapse.
Secondary structure	Sandstone, coarse-grained igneous.	Broad fissure to cavernous	Tension – stress during deformation, expansion by pressure release on unloading.	Small, cavernous. Rough-textured surface. Dry. Stable.	Wet.	Brittle failure.
Inherent structure and weathering	Conglomerate	Small caverns	Partial disaggregation of rock mass by dislodgement of large clasts	Small, cavernous. Rough-textured surface. Dry. Stable.	Wet.	Enlarged by weathering and collapse.
Inherent structure	Basalt lava flows	Wide range of fissures and caverns	Flow cooling and crusting. Volatile emission. Cooling/contraction fractures.	Wide range of shape and size of caverns and overhangs. Stable.	Wet. Locally unstable due to roof collapse.	Local collapse
Secondary processes	Basalt lava flows	Solution caverns (rare)	Solution of amygdaloidal lava.	Limited occurrence.	Large entrance opening, limited dark space.	Flooding
Secondary processes	Carbonate	Enlarged fissures and caverns.	Weathering – solution and precipitation.	Wide range of habitat.		Flooding, sediment and precipitation infill

TYPE	TYPE OF ROCK	SPACE	ORIGIN of SPACE	OPPORTUNITY	CONSTRAINT	MODIFICATION
Inherent properties and secondary processes	Intrusive igneous (granitic)	Enlarged fissure, overhang, small caverns.	Weathering enlarging joints developed by stress release and expansion	Typically vertical and fluted fissures to overhang and caverns.	Wide openings and limited dark space.	Accumulation of colluvium and surface wash debris. Vegetation growth.
Secondary processes	Basalt, sandstone, conglomerate.	Irregular caverns	Stacking and arrangement of fallen talus and landslide rock blocks.	Newly developed.	Irregular space. Limited entry size or limited dark space. Unstable.	Slope movement. Infill by fines, Vegetation growth.

3.2 Known Habitat Occurrence

Biosis supplied the location of three known cave roosting sites in proximity to the HOGWF areas—Timor Caves, Travelling Stock Route and Barry Cave (Figure 7). No details of the cave site were provided. The caves were located on the Seamless Geology map and the distribution of that geological formation displayed across the study area.

3.2.1 Timor Caves

Timor Caves are developed in the Timor Limestone Member of the Yarrimie Formation located in the valley of Isis River. Cave and geological details are provided by Connolly and Francis (1979). Limestone of the Yarrimie Formation crops out across nearly 900 ha along the Isis River but caves are known from only two small localities.

3.2.2 Barry Caves

Barry Cave is located on the crest of a spur 250 m north of a bend in the Barnard River in an area mapped as Myra beds of the Woolomin Group. The cave is not in limestone but in low-grade metamorphic geology described as slate, phyllite, chert, jasper and metabasalts. Rocks of the Myra Group crop out in NW-SE strike in -wide belts and are exposed here as the Mount Royal Range Volcanics have been eroded. The cave is described in Allen *et al.* (1986).

3.2.3 Glenrock

Glenrock karst area is 30 km southeast of the HOGWF. Over 100 caves are known in Silver Gully Formation limestone along a ridge and slopes between Orham Creek and Spring Creek (tributaries of Barnard River) (Allen et al. (1986).

3.2.4 Travelling Stock Route

Travelling Stock Route Cave is marked as adjacent to an outcrop and block scatter of rocks shown as Yarrimie Formation mudstone, chert, siliceous siltstone, limestone, and

conglomerate. It is possible the cave is developed in the Crawney Limestone Member “coralline biohermal limestone” as that formation is mapped 140 metres from the cave location. The Crawney Limestone Member is mainly covered by Liverpool Range and Royal Range basalt and crops out along strike for a short distance south of the of the marked cave location. No detail of cave geomorphology is available for this site.

3.3 Potential Habitat Occurrence

Advice from an academic and expert speleologist—Dr Susan White (La Trobe University) is that the foraging range of bats is significantly larger than previously thought and may be at least 75 km (Appendix A). Potential roost sites in basalt and other lithologies across this wider area e.g., up to 75 km foraging range must be considered.

3.3.1 Biosis Desktop Assessment

A desktop study by Biosis identified several discontinuous areas of potential habitat along the upper slopes on both sides of the Main Divide and some higher ridges between Crawney Pass and Wardens Brook. All were areas of surficial basalt geology. It was assessed that the likelihood for unreported caves was low given the areas were not remote and there was generally good surface visibility. No assessment was made of areas of non-basalt geology beyond these limits.

3.3.2 This Desktop Review

As shown in Table 1, potential habitat can occur in a wide range of geological materials. Some habitat entrances will not be discernible on aerial photography or airborne LiDAR, and small outcrop with potential habitat sites can be masked by medium to tall or other closely-spaced vegetation. After examining Point Cloud LiDAR and orthophotography of the HOGWF precinct and immediately adjacent terrain the conclusion of this desktop review is that while these techniques can be applicable for exposed ground surfaces, they cannot be conclusive in identifying (or dismissing) the likelihood of potential geological habitat.

It is unlikely that large caverns—such as those formed by dissolution of limestone at Timor and amygdaloidal basalt as at Coolah Tops—have remained undetected in the immediate and adjacent study area. The area is not remote and has been mapped in some geological detail by expert groups (Government and academic). It is an area visited by trampers and speleologists and there are known caves (see 3.2 above and Appendix A). The limited area of limestone constrains the potential for large habitat caves. Habitat in other rock type is also constrained by limited outcrop of suitable structural and weathering forms.

The undulating plateau and ridge terrain of the proposed HOGWF has no extended rock escarpments and limited outcrop of fractured basalt as vertical or inclined columnar structures. Detailed ground survey would be needed to define the extent of such outcrop but the evidence from the available data point to discontinuous and limited habitat in the immediate vicinity. Talus deposits appear to be inactive and substantially vegetated at the altitude range of the HOGWF, reducing the likelihood of habitat cavity flight entry being preserved.

3.3.3 Potential Habitat in Extended Area

The New England Orogen has limited areas of karst and cave development (Figure 7).

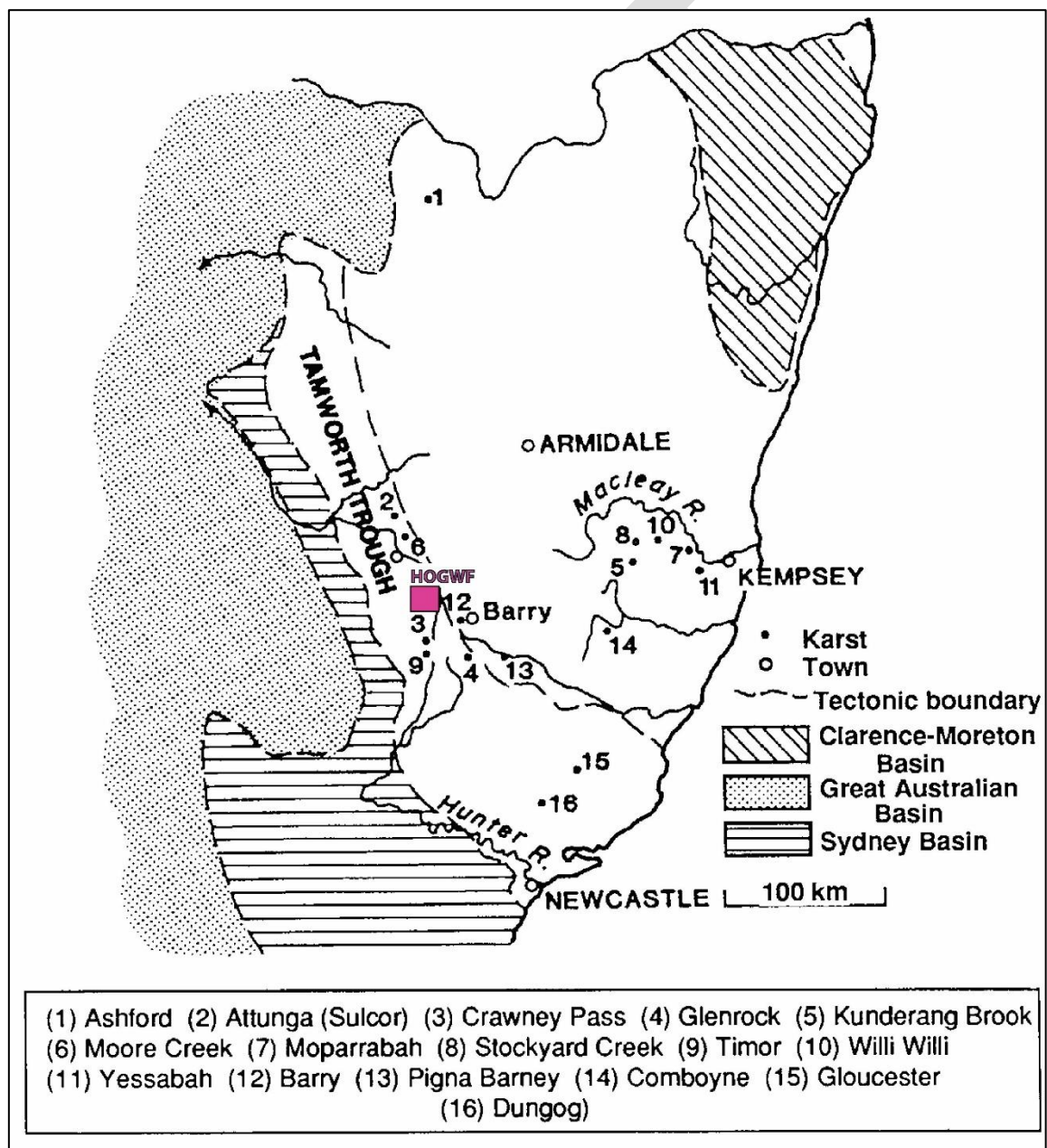


Figure 7. Karsts of the New England Fold Belt (Orogen). (From Osborne and Branagan 1988).

Dense crystalline Devonian limestone has developed multiple caverns and other karst terrain at Timor and Glenrock. The cave at Barry is unusual in being developed in non-carbonate rock.

As outlined in section 1.3 above, an envelope of approximately 30 km around the HOGWF was examined for potential habitat (Figure 7).

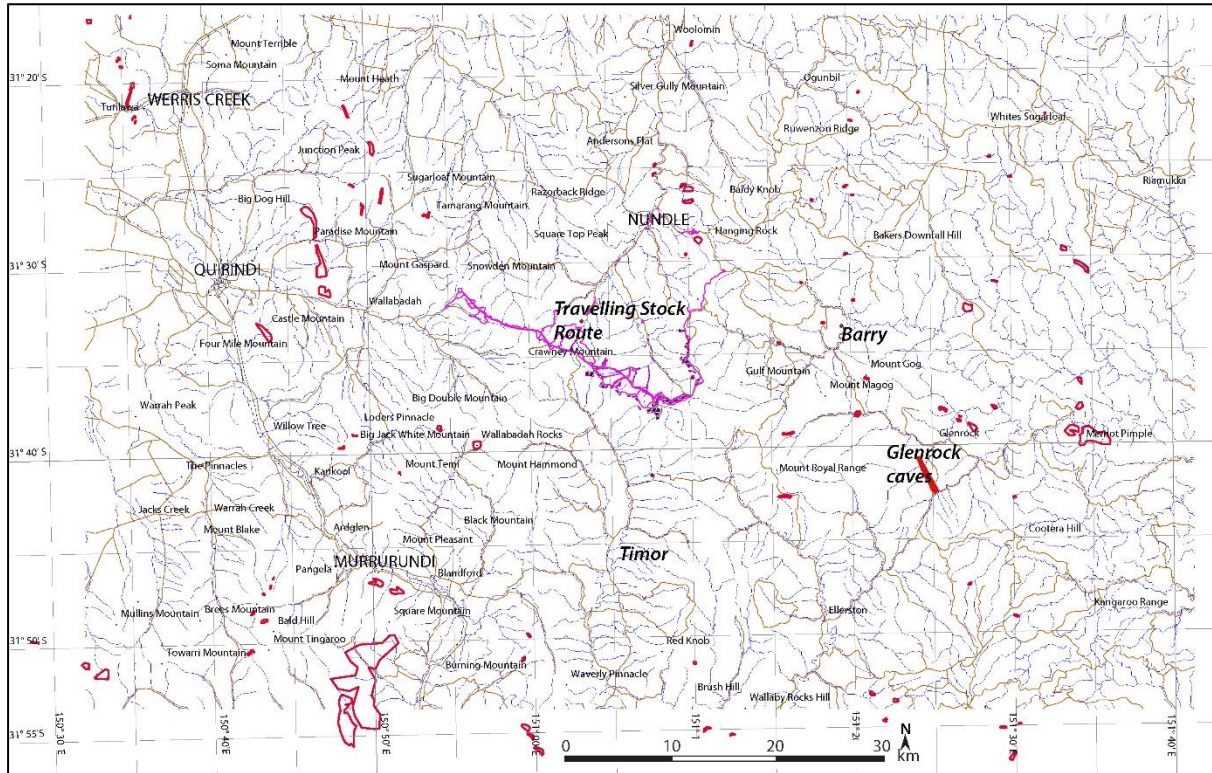


Figure 8. Potential habitat (red outlines) determined from LiDAR 5 m DEM and aerial photography. Four known cave areas shown are Glenrock, Barry, Timor and Travelling Stock Route.

The extent of basalt exposure in valleys and at the margins of the several lava fields within foraging range means smaller cavities formed in this rock may occur. Outcrop of fractured and weathered rock and downslope accumulations of blocky scree as potential habitat sites are also possibly widespread. As the basalts are hundreds of metres total thickness and are comprised of multiple flow units, there are numerous discontinuities in a vertical exposure. The discontinuities may be intra-flow e.g., separation and inflation planes of flow tongues and lobes, or more substantially defined by time-separation of upper flows with lithological differences and weathering developing stepped profiles with broad overhands and access to cavities.

Basalt relict blockstreams and limited areas of exposed scree with minimal vegetation cover at higher altitudes east and south of the HOGWF area are habitat potential, but assessment of extent and distribution of these is beyond the scope of the present study.

3.3.3.1 Habitat in Other Lithologies

The structure of the basement rocks exposed north and south of the main lava fields produce narrow ridges aligned along the NNW strike direction with a core of resistant beds forming the ridge crest and upper slopes. Isolated and groups of such ridges with crestal outcrop occur across parts of the Liverpool Plains and on the lower and mid-slopes of the Main Divide, Liverpool Range and Mount Royal Range. The wide range of lithology including carbonate and close-bedded sandstone-mudstone units have the inherent and secondary properties to develop potential habitat sites.

3.4 Landform Diversity

The geological evolution of northeast NSW has resulted in a diverse and dynamic landscape. Key elements in producing this diversity are the tectonic and volcanic history generating a range of slope forms developed by palaeo and on-going landforming processes on diverse geological materials. Cenozoic climate changes produced weathering regimes where episodes of intense rock decomposition alternated with mechanical breakdown in subsequent cold climate. Uplift of the Eastern Uplands and the rejuvenation of river systems has developed a complex drainage system.

4 CONCLUSION

Diverse terrain and lithology and dynamic geomorphology of the New England Orogen make high potential for microbat roosting sites across landscapes at all elevations in a range of rock types.

In the immediate area of the HOGWF, outcrops of fractured basalt may provide localised habitat, However, terrain and geology of this precinct provides limited opportunity for extensive habitat. While several large solution caverns in limestone and basalt occur in surrounding terrain, these are localised and there is a low probability that similar unreported large habitat sites occur. It is extremely unlikely there are basalt caverns of the dimension to accommodate a large colony. There is also a low possibility that unknown caves occur in the Devonian crystalline limestone, as these outcrops have been searched on several occasions (Allen *et al.* (1986).

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Caves containing bats in the Tamworth area, NSW

There are no caves with bats recorded very close to Tamworth but caves containing bats, both as large populations and occasional roosting occur within the flight ranges of bats. Over the past few years with better monitoring capabilities, the range of foraging bats is now known to be significantly larger than previously thought and now needs to be *at least* 75 Km radius from major bat roosts and maternity caves (pers comm; Emmi van Harten 2021).

The following places have been recorded with bats (Matthews 1985) that are within the potential range of wind towers in the Tamworth area. There may be more recent information but the major bat caves are listed in the table below.

There are other areas with caves but no bat populations are recorded for them in the information I can access. If there are now other caves known with extensive bat populations in the area, the best contact is Kempsey Speleological Society (KSS) who can be contacted via PO Box 31, Kempsey, NSW 2440. The contact person is Glen Bowman: email bowmac@live.com.au. KSS may also probably know the status of the populations. If further contacts are needed I can possibly provide them but KSS would probably be the best local knowledge.

The other major bat caves in the NE NSW and SE Queensland are further away but significant bat maternity sites occur at Riverton (just over the border into Queensland) at Riverton Main Cave (4RN-1) and at Ashford at Ashford Main Cave (2AS-1). Other smaller roosting sites might also occur in these areas.

Reference: Matthews, P.G. 1985 Australian Karst Index (Australian Speleological Federation Inc).

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Cave Area	District	Cave Number	Cave Name	Significant number of bats	Occasional roosting
Kunderang Brook	East of Kempsey	2KB-1	Youdales Cave, Hut Cave	Yes	
		2KB2			Yes
Timor (Incl Isaacs Ck; Isis R; Allston)	Timor	2TR-2	Belfry Cave		Yes
		2TR-4	Helictite Cave		Yes
Stockyard Creek	West of Kempsey	2SC-5	Carrai Bat Cave	Yes	
		2SC-7		Yes	
		2SC-9			Yes
Moparabah	West of Kempsey	2MP-1	Moparabah Cave; Main Cave	Yes	
Yessabah	WSW of Kempsey	2YE-1	Yessabah Bat Cave	Yes	
Willi Willi	West of Kempsey	2WW-1	Willi Willi Bat Cave	Yes	
		2WW-4	Possum Cave	Yes	
Moore Creek	North of Tamworth	2MC-1	Moore Creek		Yes
Sulcor	North of Tamworth	2S-4	Bullock Hole		Yes

Appendix G Offset credit summary reports

Proposal Details

Assessment Id	Proposal Name	BAM data last updated *
00020779/BAAS18138/20/00020780	Hills of Gold Wind Farm - Nandewar-Peel IBRA split	19/12/2022
Assessor Name	Report Created	BAM Data version *
Callan Wharfe	23/01/2023	56
Assessor Number	BAM Case Status	Date Finalised
BAAS18138	Finalised	23/01/2023
Assessment Revision	Assessment Type	
6	Major Projects	

* Disclaimer: BAM data last updated may indicate either complete or partial update of the BAM calculator database. BAM calculator database may not be completely aligned with Bionet.

Ecosystem credits for plant communities types (PCT), ecological communities & threatened species habitat

Zone	Vegetation zone name	TEC name	Current Vegetation integrity score	Change in Vegetation integrity (loss / gain)	Area (ha)	Sensitivity to loss (Justification)	Species sensitivity to gain class	BC Act Listing status	EPBC Act listing status	Biodiversity risk weighting	Potential SAIL	Ecosystem credits
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Blakely's Red Gum - Yellow Box grassy tall woodland on flats and hills in the Brigalow Belt South Bioregion and Nandewar Bioregion												
19	599_High	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	99.9	99.9	0.81	PCT Cleared - 80%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	51

BAM Credit Summary Report

20	599_Mode rate	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	99.9	99.9	0.5	PCT Cleared - 80%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	31
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BAM Credit Summary Report

21	599_Low	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	99.9	99.9	3.7	PCT Cleared - 80%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	229
											Subtotal	311
River Oak - Rough-barked Apple - red gum - box riparian tall woodland (wetland) of the Brigalow Belt South Bioregion and Nandewar Bioregion												
22	84_Low	Not a TEC	99.3	99.3	0.07	PCT Cleared - 40%	High Sensitivity to Gain			1.50		3
											Subtotal	3
River Oak moist riparian tall open forest of the upper Hunter Valley, including Liverpool Range												
3	486_Mode rate	Not a TEC	99.1	99.1	1.8	PCT Cleared - 40%	High Sensitivity to Gain			1.50		67

BAM Credit Summary Report

4	486_Low	Not a TEC	99.1	99.1	0.66	PCT Cleared - 40%	High Sensitivity to Gain			1.50		25
5	486_DNG	Not a TEC	69.2	69.2	0.08	PCT Cleared - 40%	High Sensitivity to Gain			1.50		2
											Subtotal	94
Rough-barked Apple - Blakely's Red Gum open forest of the Nandewar Bioregion and western New England Tableland Bioregion												
24	538_Low	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	99.7	99.7	0.06	Population size	High Sensitivity to Gain	Critically Endangered Ecological Community	Not Listed	2.50	True	4
											Subtotal	4

BAM Credit Summary Report

Silvertop Stringybark - Forest Ribbon Gum very tall moist open forest on basalt plateau on the Liverpool Range, Brigalow Belt South Bioregion												
6	490_Low	Not a TEC	98.3	98.3	1.9	PCT Cleared - 28%	High Sensitivity to Gain			1.50		69
											Subtotal	69
Silvertop Stringybark - Ribbon Gum - Rough-barked Apple open forest on basalt hills of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion												
11	540_High	Not a TEC	80.3	80.3	5.2	PCT Cleared - 56%	High Sensitivity to Gain			1.75		181
12	540_Mode rate	Not a TEC	86.1	86.1	11.6	PCT Cleared - 56%	High Sensitivity to Gain			1.75		438
13	540_Low	Not a TEC	95.9	95.9	4.3	PCT Cleared - 56%	High Sensitivity to Gain			1.75		182
14	540_DNG	Not a TEC	52.1	52.1	8.4	PCT Cleared - 56%	High Sensitivity to Gain			1.75		192
											Subtotal	993

Silvertop Stringybark - Rough-barked Apple grassy open forest of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion											
15	541_High	Not a TEC	87.3	87.3	3.6	PCT Cleared - 64%	High Sensitivity to Gain			1.75	136
16	541_Mode rate	Not a TEC	83.6	83.6	4.9	PCT Cleared - 64%	High Sensitivity to Gain			1.75	179
17	541_Low	Not a TEC	84.9	84.9	6.8	PCT Cleared - 64%	High Sensitivity to Gain			1.75	254
18	541_DNG	Not a TEC	54.7	54.7	2.5	PCT Cleared - 64%	High Sensitivity to Gain			1.75	61
										Subtotal	630

Silvertop Stringybark - Yellow Box - Apple Box - Rough-barked Apple shrub grass open forest mainly on southern slopes of the Liverpool Range, Brigalow Belt South Bioregion

7	492_High	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	93	93.0	0.01	PCT Cleared - 43%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	1
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8	492_Mode rate	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	93	93.0	0.03	PCT Cleared - 43%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	2
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9	492_Low	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	60.3	60.3	0.63	PCT Cleared - 43%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	24
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BAM Credit Summary Report

10	492_DNG	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	59.9	59.9	1.1	PCT Cleared - 43%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	41
											Subtotal	68

White Box grass shrub hill woodland on clay to loam soils on volcanic and sedimentary hills in the southern Brigalow Belt South Bioregion												
23	434_Low	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	100	100.0	0.01	Population size	High Sensitivity to Gain	Critically Endangered Ecological Community	Not Listed	2.50	True	1
											Subtotal	1

White Box grassy woodland to open woodland on basalt flats and rises in the Liverpool Plains sub-region, BBS Bioregion												
1	433_Mode rate	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	99.9	99.9	0.01	PCT Cleared - 85%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	1

BAM Credit Summary Report

2	433_Low	White Box - Yellow Box - Blakely's Red Gum Grassy Woodland and Derived Native Grassland in the NSW North Coast, New England Tableland, Nandewar, Brigalow Belt South, Sydney Basin, South Eastern Highla	99.9	99.9	0.01	PCT Cleared - 85%	High Sensitivity to Gain	Critically Endangered Ecological Community	Critically Endangered	2.50	True	1
											Subtotal	2
											Total	2175

Species credits for threatened species

Vegetation zone name	Habitat condition (Vegetation Integrity)	Change in habitat condition	Area (ha)/Count (no. individuals)	Sensitivity to loss (Justification)	Sensitivity to gain (Justification)	BC Act Listing status	EPBC Act listing status	Potential SAI	Species credits
<i>Cercartetus nanus / Eastern Pygmy-possum (Fauna)</i>									
599_High	99.9	99.9	0.56			Vulnerable	Not Listed	False	28
								Subtotal	28

<i>Chalinolobus dwyeri</i> / Large-eared Pied Bat (Fauna)									
599_Low	99.9	99.9	0.24			Vulnerable	Vulnerable	True	18
599_High	99.9	99.9	0.06			Vulnerable	Vulnerable	True	4
599_Moderate	99.9	99.9	0.05			Vulnerable	Vulnerable	True	4
								Subtotal	26
<i>Litoria booroolongensis</i> / Booroolong Frog (Fauna)									
486_Moderate	99.1	99.1	0.85			Endangered	Endangered	False	42
540_High	80.3	80.3	0.05			Endangered	Endangered	False	2
541_High	87.3	87.3	0.03			Endangered	Endangered	False	1
541_Moderate	83.6	83.6	0.03			Endangered	Endangered	False	1
599_Low	99.9	99.9	0.01			Endangered	Endangered	False	1
								Subtotal	47
<i>Ninox connivens</i> / Barking Owl (Fauna)									
433_Moderate	99.9	99.9	0.01			Vulnerable	Not Listed	False	1
433_Low	99.9	99.9	0.01			Vulnerable	Not Listed	False	1
486_Moderate	99.1	99.1	1.7			Vulnerable	Not Listed	False	83
486_Low	99.1	99.1	0.66			Vulnerable	Not Listed	False	33
490_Low	98.3	98.3	1.8			Vulnerable	Not Listed	False	88
492_High	93.0	93.0	0.01			Vulnerable	Not Listed	False	1
492_Moderate	93.0	93.0	0.03			Vulnerable	Not Listed	False	1
540_High	80.3	80.3	4.8			Vulnerable	Not Listed	False	193
540_Moderate	86.1	86.1	11.3			Vulnerable	Not Listed	False	485
540_Low	95.9	95.9	4.2			Vulnerable	Not Listed	False	202
541_High	87.3	87.3	3.5			Vulnerable	Not Listed	False	153

BAM Credit Summary Report

541_Moderate	83.6	83.6	4.2		Vulnerable	Not Listed	False	176
599_High	99.9	99.9	0.81		Vulnerable	Not Listed	False	40
599_Moderate	99.9	99.9	0.5		Vulnerable	Not Listed	False	25
599_Low	99.9	99.9	3.6		Vulnerable	Not Listed	False	180
434_Low	100.0	100.0	0.01		Vulnerable	Not Listed	False	1
538_Low	99.7	99.7	0.06		Vulnerable	Not Listed	False	3
							Subtotal	1666
<i>Petaurus norfolcensis / Squirrel Glider (Fauna)</i>								
433_Moderate	99.9	99.9	0.01		Vulnerable	Not Listed	False	1
599_High	99.9	99.9	0.56		Vulnerable	Not Listed	False	28
599_Moderate	99.9	99.9	0.4		Vulnerable	Not Listed	False	20
							Subtotal	49
<i>Phascolarctos cinereus / Koala (Fauna)</i>								
433_Moderate	99.9	99.9	0.01		Endangered	Endangered	False	1
599_Low	99.9	99.9	3.6		Endangered	Endangered	False	179
599_High	99.9	99.9	0.56		Endangered	Endangered	False	28
599_Moderate	99.9	99.9	0.4		Endangered	Endangered	False	20
433_Low	99.9	99.9	0.01		Endangered	Endangered	False	1
434_Low	100.0	100.0	0.01		Endangered	Endangered	False	1
538_Low	99.7	99.7	0.06		Endangered	Endangered	False	3
							Subtotal	233
<i>Tyto novaehollandiae / Masked Owl (Fauna)</i>								
433_Moderate	99.9	99.9	0.01		Vulnerable	Not Listed	False	1
433_Low	99.9	99.9	0.01		Vulnerable	Not Listed	False	1

BAM Credit Summary Report

599_High	99.9	99.9	0.81			Vulnerable	Not Listed	False	40
599_Moderate	99.9	99.9	0.5			Vulnerable	Not Listed	False	25
599_Low	99.9	99.9	3.6			Vulnerable	Not Listed	False	180
434_Low	100.0	100.0	0.01			Vulnerable	Not Listed	False	1
538_Low	99.7	99.7	0.06			Vulnerable	Not Listed	False	3
Subtotal									251
<i>Uvidicolus sphyrurus / Border Thick-tailed Gecko (Fauna)</i>									
599_High	99.9	99.9	0.56			Vulnerable	Vulnerable	False	28
599_Moderate	99.9	99.9	0.11			Vulnerable	Vulnerable	False	5
Subtotal									33
<i>Vespadelus troughtoni / Eastern Cave Bat (Fauna)</i>									
599_Low	99.9	99.9	0.24			Vulnerable	Not Listed	True	18
599_High	99.9	99.9	0.06			Vulnerable	Not Listed	True	4
599_Moderate	99.9	99.9	0.05			Vulnerable	Not Listed	True	4
Subtotal									26

Proposal Details

Assessment Id	Proposal Name	BAM data last updated *
00031734/BAAS18138/22/00031735	HoG - NSW North Coast - Tomalla	19/12/2022
Assessor Name	Report Created	BAM Data version *
Callan Wharfe	23/01/2023	56
Assessor Number	BAM Case Status	Date Finalised
BAAS18138	Finalised	23/01/2023
Assessment Revision	Assessment Type	
4	Major Projects	

* Disclaimer: BAM data last updated may indicate either complete or partial update of the BAM calculator database. BAM calculator database may not be completely aligned with Bionet.

Ecosystem credits for plant communities types (PCT), ecological communities & threatened species habitat

Zone	Vegetation zone name	TEC name	Current Vegetation integrity score	Change in Vegetation integrity (loss / gain)	Area (ha)	Sensitivity to loss (Justification)	Species sensitivity to gain class	BC Act Listing status	EPBC Act listing status	Biodiversity risk weighting	Potential SAI	Ecosystem credits
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Black Sallee - Snow Gum grassy woodland of the New England Tableland Bioregion												
4	507_Mode rate	Not a TEC	57.6	57.6	0.09	PCT Cleared - 87%	High Sensitivity to Gain			2.00		3
											Subtotal	3
Messmate - Mountain Gum tall moist forest of the far southern New England Tableland Bioregion												
12	931_Mode rate	Not a TEC	45	45.0	2	PCT Cleared - 85%	High Sensitivity to Gain			2.00		45
											Subtotal	45
Messmate open forest of the tableland edge of the NSW North Coast Bioregion and New England Tableland Bioregion												
13	934_High	Not a TEC	87	87.0	4.2	PCT Cleared - 71%	High Sensitivity to Gain			2.00		183
14	934_Mode rate	Not a TEC	99.7	99.7	0.19	PCT Cleared - 71%	High Sensitivity to Gain			2.00		9
15	934_DNG	Not a TEC	20.1	20.1	0.17	PCT Cleared - 71%	High Sensitivity to Gain			2.00		2
21	934_Low	Not a TEC	99.7	99.7	0.92	PCT Cleared - 71%	High Sensitivity to Gain			2.00		46

										Subtotal	240
Mountain Ribbon Gum - Messmate open forest of escarpment ranges of the NSW North Coast Bioregion and New England Tableland Bioregion											
16	954_High	Not a TEC	99.8	99.8	1.2	PCT Cleared - 50%	High Sensitivity to Gain			1.75	54
										Subtotal	54
River Oak moist riparian tall open forest of the upper Hunter Valley, including Liverpool Range											
1	486_High	Not a TEC	99.8	99.8	0.54	PCT Cleared - 40%	High Sensitivity to Gain			1.50	20
2	486_Mode rate	Not a TEC	99.8	99.8	1.4	PCT Cleared - 40%	High Sensitivity to Gain			1.50	54
										Subtotal	74
Silvertop Stringybark - Ribbon Gum - Rough-barked Apple open forest on basalt hills of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion											
5	540_High	Not a TEC	72.9	72.9	10.8	PCT Cleared - 56%	High Sensitivity to Gain			1.75	345
6	540_Mode rate	Not a TEC	78	78.0	16.7	PCT Cleared - 56%	High Sensitivity to Gain			1.75	569
7	540_Low	Not a TEC	94.9	94.9	3.7	PCT Cleared - 56%	High Sensitivity to Gain			1.75	154

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8	540_DNG	Not a TEC	44.2	44.2	5.3	PCT Cleared - 56%	High Sensitivity to Gain			1.75		103
											Subtotal	1171
Silvertop Stringybark - Rough-barked Apple grassy open forest of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion												
9	541_High	Not a TEC	79.9	79.9	7.7	PCT Cleared - 64%	High Sensitivity to Gain			1.75		269
10	541_Mode rate	Not a TEC	76.9	76.9	4.2	PCT Cleared - 64%	High Sensitivity to Gain			1.75		142
11	541_Low	Not a TEC	61.5	61.5	1.1	PCT Cleared - 64%	High Sensitivity to Gain			1.75		30
											Subtotal	441
Silvertop Stringybark - Yellow Box - Apple Box - Rough-barked Apple shrub grass open forest mainly on southern slopes of the Liverpool Range, Brigalow Belt South Bioregion												
3	492_Mode rate	Not a TEC	89.1	89.1	1.4	PCT Cleared - 43%	High Sensitivity to Gain			1.50		47
											Subtotal	47

BAM Credit Summary Report

Snow Gum - Mountain Gum - Mountain Ribbon Gum open forest on ranges of the NSW North Coast Bioregion and eastern New England Tableland Bioregion											
17	1194_High	Not a TEC	72.1	72.1	10.8	PCT Cleared - 80%	High Sensitivity to Gain			2.00	389
18	1194_Moderate	Not a TEC	65.7	65.7	8.4	PCT Cleared - 80%	High Sensitivity to Gain			2.00	277
19	1194_Low	Not a TEC	38	38.0	3.3	PCT Cleared - 80%	High Sensitivity to Gain			2.00	63
20	1194_DNG	Not a TEC	7.8	7.8	3.7	PCT Cleared - 80%	High Sensitivity to Gain			2.00	0
										Subtotal	729
										Total	2804

Species credits for threatened species

Vegetation zone name	Habitat condition (Vegetation Integrity)	Change in habitat condition	Area (ha)/Count (no. individuals)	Sensitivity to loss (Justification)	Sensitivity to gain (Justification)	BC Act Listing status	EPBC Act listing status	Potential SAIL	Species credits
<i>Cercartetus nanus</i> / Eastern Pygmy-possum (Fauna)									
934_High	87.0	87.0	2.4			Vulnerable	Not Listed	False	104
954_High	99.8	99.8	1.2			Vulnerable	Not Listed	False	61
1194_High	72.1	72.1	9.3			Vulnerable	Not Listed	False	335

BAM Credit Summary Report

								Subtotal	500
<i>Myotis macropus / Southern Myotis (Fauna)</i>									
1194_Moderate	65.7	65.7	1.4			Vulnerable	Not Listed	False	47
								Subtotal	47
<i>Ninox connivens / Barking Owl (Fauna)</i>									
486_High	99.8	99.8	0.54			Vulnerable	Not Listed	False	27
486_Moderate	99.8	99.8	1.4			Vulnerable	Not Listed	False	71
492_Moderate	89.1	89.1	1.4			Vulnerable	Not Listed	False	62
540_High	72.9	72.9	9.5			Vulnerable	Not Listed	False	347
540_Moderate	78.0	78.0	1.8			Vulnerable	Not Listed	False	69
540_Low	94.9	94.9	2.2			Vulnerable	Not Listed	False	103
541_High	79.9	79.9	7.5			Vulnerable	Not Listed	False	300
541_Moderate	76.9	76.9	1.6			Vulnerable	Not Listed	False	61
934_High	87.0	87.0	0.31			Vulnerable	Not Listed	False	13
934_Moderate	99.7	99.7	0.19			Vulnerable	Not Listed	False	9
954_High	99.8	99.8	0.78			Vulnerable	Not Listed	False	39
1194_High	72.1	72.1	2.7			Vulnerable	Not Listed	False	96
1194_Moderate	65.7	65.7	0.6			Vulnerable	Not Listed	False	20
1194_Low	38.0	38.0	0.03			Vulnerable	Not Listed	False	1
1194_DNG	7.8	7.8	0.88			Vulnerable	Not Listed	False	3
934_Low	99.7	99.7	0.91			Vulnerable	Not Listed	False	45
								Subtotal	1266
<i>Ninox strenua / Powerful Owl (Fauna)</i>									
931_Moderate	45.0	45.0	1.3			Vulnerable	Not Listed	False	28

BAM Credit Summary Report

934_High	87.0	87.0	0.31		Vulnerable	Not Listed	False	13
934_Moderate	99.7	99.7	0.19		Vulnerable	Not Listed	False	9
954_High	99.8	99.8	0.78		Vulnerable	Not Listed	False	39
1194_High	72.1	72.1	2.7		Vulnerable	Not Listed	False	96
1194_Moderate	65.7	65.7	0.6		Vulnerable	Not Listed	False	20
1194_Low	38.0	38.0	0.03		Vulnerable	Not Listed	False	1
1194_DNG	7.8	7.8	0.88		Vulnerable	Not Listed	False	3
934_Low	99.7	99.7	0.91		Vulnerable	Not Listed	False	45
							Subtotal	254
<i>Petaurus norfolcensis / Squirrel Glider (Fauna)</i>								
1194_High	72.1	72.1	5.8		Vulnerable	Not Listed	False	209
1194_Moderate	65.7	65.7	0.67		Vulnerable	Not Listed	False	22
							Subtotal	231
<i>Phascolarctos cinereus / Koala (Fauna)</i>								
507_Moderate	57.6	57.6	0.09		Endangered	Endangered	False	3
931_Moderate	45.0	45.0	1.8		Endangered	Endangered	False	40
934_High	87.0	87.0	3.3		Endangered	Endangered	False	142
934_Moderate	99.7	99.7	0.15		Endangered	Endangered	False	7
954_High	99.8	99.8	1.1		Endangered	Endangered	False	56
1194_High	72.1	72.1	10.2		Endangered	Endangered	False	366
1194_Moderate	65.7	65.7	1.6		Endangered	Endangered	False	53
1194_Low	38.0	38.0	3.3		Endangered	Endangered	False	63
934_Low	99.7	99.7	0.92		Endangered	Endangered	False	46
							Subtotal	776

<i>Tyto novaehollandiae / Masked Owl (Fauna)</i>									
931_Moderate	45.0	45.0	2			Vulnerable	Not Listed	False	45
934_High	87.0	87.0	0.31			Vulnerable	Not Listed	False	13
934_Moderate	99.7	99.7	0.19			Vulnerable	Not Listed	False	9
954_High	99.8	99.8	0.78			Vulnerable	Not Listed	False	39
1194_High	72.1	72.1	4			Vulnerable	Not Listed	False	143
1194_Moderate	65.7	65.7	0.89			Vulnerable	Not Listed	False	29
1194_Low	38.0	38.0	1.1			Vulnerable	Not Listed	False	21
1194_DNG	7.8	7.8	0.76			Vulnerable	Not Listed	False	3
934_Low	99.7	99.7	0.92			Vulnerable	Not Listed	False	46
								Subtotal	348
<i>Tyto tenebricosa / Sooty Owl (Fauna)</i>									
540_Moderate	78.0	78.0	1.8			Vulnerable	Not Listed	True	103
1194_Moderate	65.7	65.7	0.23			Vulnerable	Not Listed	True	11
								Subtotal	114

Proposal Details

Assessment Id	Proposal Name	BAM data last updated *
00031734/BAAS18138/22/00031737	HoG - New England Tablelands - Walcha Plateau	19/12/2022
Assessor Name	Report Created	BAM Data version *
Callan Wharfe	23/01/2023	56
Assessor Number	BAM Case Status	Date Finalised
BAAS18138	Finalised	23/01/2023
Assessment Revision	Assessment Type	
4	Major Projects	

* Disclaimer: BAM data last updated may indicate either complete or partial update of the BAM calculator database. BAM calculator database may not be completely aligned with Bionet.

Ecosystem credits for plant communities types (PCT), ecological communities & threatened species habitat

Zone	Vegetation zone name	TEC name	Current Vegetation integrity score	Change in Vegetation integrity (loss / gain)	Area (ha)	Sensitivity to loss (Justification)	Species sensitivity to gain class	BC Act Listing status	EPBC Act listing status	Biodiversity risk weighting	Potential SAI	Ecosystem credits
Messmate - Mountain Gum tall moist forest of the far southern New England Tableland Bioregion												
7	931_High	Not a TEC	44.4	44.4	0.83	PCT Cleared - 85%	High Sensitivity to Gain			2.00		18

BAM Credit Summary Report

8	931_Mode rate	Not a TEC	55.1	55.1	1.4	PCT Cleared - 85%	High Sensitivity to Gain			2.00		39
9	931_Low	Not a TEC	26.9	26.9	0.22	PCT Cleared - 85%	High Sensitivity to Gain			2.00		3
											Subtotal	60
Messmate open forest of the tableland edge of the NSW North Coast Bioregion and New England Tableland Bioregion												
10	934_High	Not a TEC	91.2	91.2	2.2	PCT Cleared - 71%	High Sensitivity to Gain			2.00		101
11	934_Mode rate	Not a TEC	99.6	99.6	0.13	PCT Cleared - 71%	High Sensitivity to Gain			2.00		6
12	934_Low	Not a TEC	99.6	99.6	0.4	PCT Cleared - 71%	High Sensitivity to Gain			2.00		20
13	934_DNG	Not a TEC	21.9	21.9	16.4	PCT Cleared - 71%	High Sensitivity to Gain			2.00		179
											Subtotal	306
Mountain Ribbon Gum - Messmate - Broad-leaved Stringybark open forest on granitic soils of the New England Tableland Bioregion												
1	526_High	Not a TEC	99.3	99.3	0.39	PCT Cleared - 56%	High Sensitivity to Gain			1.75		17

BAM Credit Summary Report

2	526_Mode rate	Not a TEC	99.3	99.3	0.37	PCT Cleared - 56%	High Sensitivity to Gain			1.75		16
											Subtotal	33
Silvertop Stringybark - Ribbon Gum - Rough-barked Apple open forest on basalt hills of southern Nandewar Bioregion, southern New England Tableland Bioregion and NSW North Coast Bioregion												
3	540_High_TEC	Ribbon Gum—Mountain Gum—Snow Gum Grassy Forest/Woodland of the New England Tableland Bioregion	73.3	73.3	0.49	PCT Cleared - 56%	High Sensitivity to Gain	Endangered Ecological Community	Not Listed	2.00		18
4	540_Mode rate_TEC	Ribbon Gum—Mountain Gum—Snow Gum Grassy Forest/Woodland of the New England Tableland Bioregion	82.1	82.1	0.89	PCT Cleared - 56%	High Sensitivity to Gain	Endangered Ecological Community	Not Listed	2.00		37

BAM Credit Summary Report

5	540_Low_T EC	Ribbon Gum—Mountain Gum—Snow Gum Grassy Forest/Woodlan d of the New England Tableland Bioregion	95.4	95.4	0.03	PCT Cleared - 56%	High Sensitivity to Gain	Endangered Ecological Community	Not Listed	2.00		1
											Subtot al	56
Snow Grass - Swamp Foxtail tussock grassland sedgeland of cold air drainage valleys of the New England Tableland Bioregion												
6	586_Low	Not a TEC	41.7	41.7	2.6	PCT Cleared - 75%	High Sensitivity to Gain			2.00		53
											Subtot al	53

Snow Gum - Mountain Gum - Mountain Ribbon Gum open forest on ranges of the NSW North Coast Bioregion and eastern New England Tableland Bioregion											
14	1194_High_TEC	Ribbon Gum—Mountain Gum—Snow Gum Grassy Forest/Woodland of the New England Tableland Bioregion	72.6	72.6	5.4	PCT Cleared - 80%	High Sensitivity to Gain	Endangered Ecological Community	Not Listed	2.00	198
15	1194_Moderate_TEC	Ribbon Gum—Mountain Gum—Snow Gum Grassy Forest/Woodland of the New England Tableland Bioregion	64.6	64.6	7.2	PCT Cleared - 80%	High Sensitivity to Gain	Endangered Ecological Community	Not Listed	2.00	233
16	1194_Low_TEC	Ribbon Gum—Mountain Gum—Snow Gum Grassy Forest/Woodland of the New England Tableland Bioregion	37.5	37.5	3.2	PCT Cleared - 80%	High Sensitivity to Gain	Endangered Ecological Community	Not Listed	2.00	60

BAM Credit Summary Report

17	1194_DNG _TEC	Ribbon Gum—Mountain Gum—Snow Gum Grassy Forest/Woodlan d of the New England Tableland Bioregion	7.4	7.4	1.7	PCT Cleared - 80%	High Sensitivity to Gain	Endangered Ecological Community	Not Listed	2.00		0
											Subtotal	491
											Total	999

Species credits for threatened species

Vegetation zone name	Habitat condition (Vegetation Integrity)	Change in habitat condition	Area (ha)/Count (no. individuals)	Sensitivity to loss (Justification)	Sensitivity to gain (Justification)	BC Act Listing status	EPBC Act listing status	Potential SAIL	Species credits
<i>Cercartetus nanus</i> / Eastern Pygmy-possum (Fauna)									
526_High	99.3	99.3	0.39			Vulnerable	Not Listed	False	19
931_High	44.4	44.4	0.83			Vulnerable	Not Listed	False	18
934_High	91.2	91.2	2.2			Vulnerable	Not Listed	False	101
1194_High_TEC	72.6	72.6	5.4			Vulnerable	Not Listed	False	198
Subtotal									336
<i>Chalinolobus dwyeri</i> / Large-eared Pied Bat (Fauna)									
931_High	44.4	44.4	0.28			Vulnerable	Vulnerable	True	9
931_Moderate	55.1	55.1	1.4			Vulnerable	Vulnerable	True	59

BAM Credit Summary Report

934_Moderate	99.6	99.6	0.13			Vulnerable	Vulnerable	True	10
934_Low	99.6	99.6	0.4			Vulnerable	Vulnerable	True	30
934_DNG	21.9	21.9	7.5			Vulnerable	Vulnerable	True	123
1194_High_TEC	72.6	72.6	2.7			Vulnerable	Vulnerable	True	148
1194_Moderate_TEC	64.6	64.6	4			Vulnerable	Vulnerable	True	193
1194_Low_TEC	37.5	37.5	1.9			Vulnerable	Vulnerable	True	53
1194_DNG_TEC	7.4	7.4	1.1			Vulnerable	Vulnerable	True	6
								Subtotal	631
<i>Myotis macropus / Southern Myotis (Fauna)</i>									
934_DNG	21.9	21.9	1.8			Vulnerable	Not Listed	False	20
1194_High_TEC	72.6	72.6	0.06			Vulnerable	Not Listed	False	2
1194_Moderate_TEC	64.6	64.6	0.64			Vulnerable	Not Listed	False	21
								Subtotal	43
<i>Ninox connivens / Barking Owl (Fauna)</i>									
526_High	99.3	99.3	0.39			Vulnerable	Not Listed	False	19
526_Moderate	99.3	99.3	0.35			Vulnerable	Not Listed	False	17
540_High_TEC	73.3	73.3	0.49			Vulnerable	Not Listed	False	18
540_Moderate_TEC	82.1	82.1	0.61			Vulnerable	Not Listed	False	25
540_Low_TEC	95.4	95.4	0.03			Vulnerable	Not Listed	False	1
934_High	91.2	91.2	0.48			Vulnerable	Not Listed	False	22
934_Moderate	99.6	99.6	0.13			Vulnerable	Not Listed	False	6
934_Low	99.6	99.6	0.4			Vulnerable	Not Listed	False	20

BAM Credit Summary Report

934_DNG	21.9	21.9	0.03		Vulnerable	Not Listed	False	1
1194_High_TEC	72.6	72.6	1.9		Vulnerable	Not Listed	False	68
1194_Moderate_TEC	64.6	64.6	2.6		Vulnerable	Not Listed	False	85
1194_Low_TEC	37.5	37.5	1.6		Vulnerable	Not Listed	False	30
							Subtotal	312
<i>Ninox strenua / Powerful Owl (Fauna)</i>								
526_High	99.3	99.3	0.39		Vulnerable	Not Listed	False	19
526_Moderate	99.3	99.3	0.35		Vulnerable	Not Listed	False	17
931_High	44.4	44.4	0.42		Vulnerable	Not Listed	False	9
931_Moderate	55.1	55.1	0.37		Vulnerable	Not Listed	False	10
934_High	91.2	91.2	0.48		Vulnerable	Not Listed	False	22
934_Moderate	99.6	99.6	0.13		Vulnerable	Not Listed	False	6
934_Low	99.6	99.6	0.4		Vulnerable	Not Listed	False	20
934_DNG	21.9	21.9	0.03		Vulnerable	Not Listed	False	1
1194_High_TEC	72.6	72.6	1.9		Vulnerable	Not Listed	False	68
1194_Moderate_TEC	64.6	64.6	2.6		Vulnerable	Not Listed	False	85
1194_Low_TEC	37.5	37.5	1.6		Vulnerable	Not Listed	False	30
							Subtotal	287
<i>Petaurus norfolcensis / Squirrel Glider (Fauna)</i>								
526_High	99.3	99.3	0.39		Vulnerable	Not Listed	False	19
526_Moderate	99.3	99.3	0.35		Vulnerable	Not Listed	False	17
1194_High_TEC	72.6	72.6	3.2		Vulnerable	Not Listed	False	117

BAM Credit Summary Report

1194_Moderate_TEC	64.6	64.6	6.1		Vulnerable	Not Listed	False	196
							Subtotal	349
<i>Phascolarctos cinereus / Koala (Fauna)</i>								
526_High	99.3	99.3	0.39		Endangered	Endangered	False	19
526_Moderate	99.3	99.3	0.35		Endangered	Endangered	False	17
931_High	44.4	44.4	0.83		Endangered	Endangered	False	18
931_Moderate	55.1	55.1	1.4		Endangered	Endangered	False	39
931_Low	26.9	26.9	0.12		Endangered	Endangered	False	2
934_High	91.2	91.2	2.2		Endangered	Endangered	False	101
934_Moderate	99.6	99.6	0.13		Endangered	Endangered	False	6
934_Low	99.6	99.6	0.4		Endangered	Endangered	False	20
1194_High_TEC	72.6	72.6	5.4		Endangered	Endangered	False	198
1194_Moderate_TEC	64.6	64.6	4.9		Endangered	Endangered	False	157
1194_Low_TEC	37.5	37.5	3		Endangered	Endangered	False	57
							Subtotal	634
<i>Tyto novaehollandiae / Masked Owl (Fauna)</i>								
1194_Moderate_TEC	64.6	64.6	0.3		Vulnerable	Not Listed	False	10
1194_Low_TEC	37.5	37.5	0.08		Vulnerable	Not Listed	False	1
							Subtotal	11
<i>Vespadelus troughtoni / Eastern Cave Bat (Fauna)</i>								
931_High	44.4	44.4	0.28		Vulnerable	Not Listed	True	9

BAM Credit Summary Report

931_Moderate	55.1	55.1	1.4			Vulnerable	Not Listed	True	59
934_Moderate	99.6	99.6	0.13			Vulnerable	Not Listed	True	10
934_Low	99.6	99.6	0.4			Vulnerable	Not Listed	True	30
934_DNG	21.9	21.9	7.5			Vulnerable	Not Listed	True	123
1194_High_TEC	72.6	72.6	2.7			Vulnerable	Not Listed	True	148
1194_Moderate_TEC	64.6	64.6	4			Vulnerable	Not Listed	True	193
1194_Low_TEC	37.5	37.5	1.9			Vulnerable	Not Listed	True	53
1194_DNG_TEC	7.4	7.4	1.1			Vulnerable	Not Listed	True	6
								Subtotal	631

Proposal Details

Assessment Id	Proposal Name	BAM data last updated *
00020779/BAAS18138/20/00021863	Hills of Gold Wind Farm	19/12/2022
Assessor Name	Report Created	BAM Data version *
Callan Wharfe	23/01/2023	56
Assessor Number	BAM Case Status	Date Finalised
BAAS18138	Finalised	23/01/2023
Assessment Revision	Assessment Type	
5	Major Projects	

* Disclaimer: BAM data last updated may indicate either complete or partial update of the BAM calculator database. BAM calculator database may not be completely aligned with Bionet.

Ecosystem credits for plant communities types (PCT), ecological communities & threatened species habitat

Zone	Vegetation zone name	TEC name	Current Vegetation integrity score	Change in Vegetation integrity (loss / gain)	Area (ha)	Sensitivity to loss (Justification)	Species sensitivity to gain class	BC Act Listing status	EPBC Act listing status	Biodiversity risk weighting	Potential SAI	Ecosystem credits
Narrow-leaved Ironbark - Grey Box - Spotted Gum shrub - grass woodland of the central and lower Hunter												
1	1604_Low	Not a TEC	99.7	99.7	0.02	PCT Cleared - 71%	High Sensitivity to Gain			2.00		1
											Subtotal	1

BAM Credit Summary Report

Narrow-leaved Ironbark - Grey Box grassy woodland of the central and upper Hunter											
2	1691_Low	Not a TEC	99.7	99.7	0.04	PCT Cleared - 77%	High Sensitivity to Gain			2.00	2
										Subtotal	2
										Total	3

Species credits for threatened species

Vegetation zone name	Habitat condition (Vegetation Integrity)	Change in habitat condition	Area (ha)/Count (no. individuals)	Sensitivity to loss (Justification)	Sensitivity to gain (Justification)	BC Act Listing status	EPBC Act listing status	Potential SAI	Species credits
Myotis macropus / Southern Myotis (Fauna)									
1604_Low	99.7	99.7	0.02			Vulnerable	Not Listed	False	1
								Subtotal	1
Phascolarctos cinereus / Koala (Fauna)									
1604_Low	99.7	99.7	0.02			Endangered	Endangered	False	1
1691_Low	99.7	99.7	0.02			Endangered	Endangered	False	1
								Subtotal	2

Appendix H BAM plot survey data

BDAR plot	Field ID	PCT	Cond.	easting	northing	bearing	Comp Tree	Comp Shrub	Comp Grass	Comp Forbs	Comp Ferns	Comp Other	Struc Tree	Struc Shrub	Struc Grass
1	HoG-Mar-13	931	Low	327308.9662	6509746.216	352	1	0	2	4	1	0	55.0	0.0	6.0
2	HoG_Mar_14	1194	DNG	318788.1415	6501027.491	112	0	0	3	15	1	1	0.0	0.0	14.1
3	HoG_Mar_15	541	High	319104.1107	6501473.217	282	2	4	4	25	1	3	70.0	50.6	2.7
4	HoG_Mar_16	540	High	319185.5396	6501621.687	180	1	4	9	27	2	2	75.0	27.4	9.4
5	HoG_TLC_Mar_01	1194	Low	324070.4841	6499595.519	300	1	2	9	27	1	5	5.0	0.2	7.9
6	HoG_TLC_Mar_02	540	High	321676.0275	6500657.758	0	1	7	17	27	1	4	60.0	12.7	33.5
7	HoG_TLC_Mar_03	540	Moderate	320039.0753	6501800.129	320	3	6	15	18	4	6	15.2	65.2	74.1
8	HoG_TLC_Mar_04	540	DNG	320626.6472	6501288.621	30	1	6	9	25	0	1	0.1	1.7	112.1
9	HoG_TLC_Mar_05	492	Moderate	320460.6042	6501462.687	190	4	7	13	26	2	5	54.3	5.5	28.1
10	HoG-Mar-01	934	High	326223.9671	6500317.283	340	2	6	5	18	1	3	40.0	9.7	85.3
11	HoG-Mar-02	1194	Low	326447.0812	6500884.121	345	1	2	1	9	1	1	5.0	70.0	20.0
12	HoG-Mar-03	1194	High	325970.2763	6503456.045	136	2	5	3	13	1	1	20.0	5.3	82.5
13	HoG-Mar-04	1194	Moderate	325897.2991	6504202.339	10	3	3	2	14	1	5	25.0	10.3	3.0
14	HoG-Mar-05	1194	DNG	325809.2251	6498707.069	176	0	0	2	6	1	0	0.0	0.0	52.0
15	HoG-Mar-06	931	High	325416.449	6498917.793	294	1	2	2	9	1	2	45.0	6.0	6.0
16	HoG-Mar-07	934	DNG	323571.1749	6499222.691	168	0	0	4	11	1	1	0.0	0.0	21.1
17	HoG-Mar-08	1194	Moderate	323694.2016	6499507.521	116	2	1	2	11	1	1	25.0	8.0	31.0
18	HoG-Mar-09	931	Moderate	317618.9057	6502049.383	56	1	3	2	9	1	3	40.0	3.6	3.1
19	HoG-Mar-10	507	Moderate	317222.7521	6502995.643	304	2	2	3	17	1	3	33.0	2.5	11.1
20	HoG-Mar-11	1194	High	316955.454	6502914.754	58	2	4	6	20	1	1	15.5	15.1	77.8
21	HoG-Mar-12	931	Moderate	327223.7577	6509755.951	0	2	2	3	12	0	2	23.0	1.0	74.0
22	HoG_TLC_Mar_07	492	Low	309500.9931	6506810.496	282	1	2	6	12	0	1	5.0	8.0	37.0
23	HoG_TLC_Mar_06	541	High	307959.7084	6507113.884	126	2	4	5	13	0	2	35.0	35.2	14.0
24	HoG_TLC_Mar_08	492	Moderate	308602.1704	6507013.704	81	2	1	7	12	1	3	30.0	50.0	65.0
25	NWF_02	934	DNG	324407.3537	6498745.128	112	0	0	2	4	1	0	0.0	0.0	0.6
26	NWF_01	1194	High	324838.1716	6497967.382	100	1	2	5	20	1	0	10.0	0.1	85.0
27	34784_01	540	Moderate	323407.929	6500249.198	150	3	5	9	20	0	5	55.1	6.3	69.7
28	34784_02	540	Moderate	324879.2154	6500128.176	55	3	8	8	25	1	5	20.1	2.5	49.9
29	34784_04	541	Moderate	319336.5404	6502344.313	345	2	9	11	16	1	6	15.0	19.0	79.8
30	34784_05	1194	Moderate	329445.6977	6512527.375	225	3	2	6	16	0	1	22.0	0.6	76.5
31	34784_07	1194	High	325201.2616	6499090.96	270	2	0	3	11	1	1	35.1	0.0	30.2
32	34784_13	540	Low	319781.5162	6501896.426	155	2	5	6	18	0	3	85.0	3.2	21.2
33	34784_16	540	Moderate	323184.6919	6500162.409	330	4	2	7	21	2	2	30.2	0.3	23.7
34	34784_17	540	High	324824.4673	6500104.34	40	2	6	14	29	1	4	20.1	0.7	26.9
35	34784_20	492	High	326388.8792	6516251.367	100	6	6	13	22	3	5	22.5	4.4	72.9
36	34784_28	586	Low	324631.2	6500025.962	90	0	0	7	5	0	0	0.0	0.0	11.5
37	34784_34	934	High	325935.6234	6500220.052	60	4	7	5	17	2	5	32.0	80.6	85.7
38	34784_42	931	Moderate	324169.2682	6499236.13	200	3	0	2	11	1	0	0.3	0.0	60.1
39	34784_46	541	Low	314828.1934	6504683.359	130	0	1	12	13	0	1	0.0	0.1	78.7

BDAR plot	Field ID	PCT	Cond.	Struc Forbs	Struc Ferns	Struc Other	Fun Large Trees	Fun Hollow Trees	Fun Litter Cover	Fun Len Fallen Logs	Fun Tree Stem 5to9	Fun Tree Stem 10to19	Fun Tree Stem 20to29
1	HoG-Mar-13	931	Low	2.6	0.1	0.0	4	0	4.0	0.0	0	0	0
2	HoG_Mar_14	1194	DNG	91.4	4.0	0.1	0	0	2.0	0.0	0	0	0
3	HoG_Mar_15	541	High	13.0	0.1	0.8	1	0	22.0	0.0	1	0	1
4	HoG_Mar_16	540	High	9.1	0.2	0.6	1	1	62.0	35.0	0	1	1
5	HoG_TLC_Mar_01	1194	Low	4.6	70.0	0.6	0	1	7.0	60.0	0	0	1
6	HoG_TLC_Mar_02	540	High	15.5	0.2	3.9	2	0	45.0	5.0	0	0	1
7	HoG_TLC_Mar_03	540	Moderate	10.0	2.3	5.9	1	0	21.0	53.0	0	0	1
8	HoG_TLC_Mar_04	540	DNG	15.0	0.0	0.1	0	0	2.0	0.0	0	0	0
9	HoG_TLC_Mar_05	492	Moderate	7.6	1.1	2.3	5	0	56.0	356.0	0	1	1
10	HoG-Mar-01	934	High	12.6	5.0	2.5	0	2	62.0	64.0	0	1	1
11	HoG-Mar-02	1194	Low	27.4	1.0	5.0	0	0	29.0	19.0	0	1	1
12	HoG-Mar-03	1194	High	6.4	0.2	0.1	2	2	73.0	24.0	0	1	1
13	HoG-Mar-04	1194	Moderate	10.3	1.0	2.3	2	0	70.0	22.0	0	1	1
14	HoG-Mar-05	1194	DNG	11.9	95.0	0.0	0	0	8.0	0.0	0	0	0
15	HoG-Mar-06	931	High	74.9	0.5	4.0	0	0	72.0	56.0	0	1	1
16	HoG-Mar-07	934	DNG	14.0	75.0	0.1	0	0	11.0	0.0	0	0	0
17	HoG-Mar-08	1194	Moderate	14.0	3.0	0.5	8	6	94.0	203.0	0	0	1
18	HoG-Mar-09	931	Moderate	28.7	0.1	0.8	8	7	51.0	28.0	0	0	0
19	HoG-Mar-10	507	Moderate	20.6	5.0	0.3	2	4	52.0	72.0	0	1	1
20	HoG-Mar-11	1194	High	7.7	0.1	0.1	2	4	57.0	44.0	0	0	1
21	HoG-Mar-12	931	Moderate	16.1	0.0	0.6	2	2	18.0	86.0	0	1	1
22	HoG_TLC_Mar_07	492	Low	46.4	0.0	0.1	1	0	45.0	73.0	0	0	1
23	HoG_TLC_Mar_06	541	High	14.7	0.0	2.0	2	1	37.0	64.0	0	0	0
24	HoG_TLC_Mar_08	492	Moderate	47.7	1.0	1.6	2	2	33.0	36.0	0	0	1
25	NWF_02	934	DNG	0.1	25	0.0	0	0	9.0	0.0	0	0	0
26	NWF_01	1194	High	0.1	15.0	0.0	4	0	62.0	70.0	0	1	1
27	34784_01	540	Moderate	19.8	0.0	3.3	3	1	30.0	66.0	0	0	1
28	34784_02	540	Moderate	13.5	4.0	1.1	2	0	41.0	82.0	1	0	1
29	34784_04	541	Moderate	12.2	0.2	6.7	1	1	25.0	66.0	1	1	1
30	34784_05	1194	Moderate	13.5	0.0	0.1	4	2	21.0	55.0	1	1	1
31	34784_07	1194	High	16.3	0.1	0.1	4	0	50.0	19.0	0	0	0
32	34784_13	540	Low	16.5	0.0	1.6	2	1	54.0	34.0	1	1	1
33	34784_16	540	Moderate	34.9	0.2	6.0	3	2	12.0	77.0	1	1	1
34	34784_17	540	High	3.8	0.1	0.4	12	0	32.0	73.0	0	1	1
35	34784_20	492	High	8.4	0.3	0.6	10	1	29.0	45.0	0	1	1
36	34784_28	586	Low	0.7	0.0	0.0	0	0	11.0	0.0	0	0	0
37	34784_34	934	High	2.5	0.8	0.7	1	3	83.0	202.0	1	1	1
38	34784_42	931	Moderate	1.6	30.0	0.0	3	3	4.0	17.0	0	0	0
39	34784_46	541	Low	5.6	0.0	0.1	0	0	56.0	0.0	0	1	1

BDAR plot	Field ID	PCT	Cond.	Fun Tree Stem 30to49	Fun Tree Stem 50to79	Fun Tree Regen	Fun High Threat Exotic
1	HoG-Mar-13	931	Low	0	0	0	0.0
2	HoG_Mar_14	1194	DNG	0	0	0	0.0
3	HoG_Mar_15	541	High	1	1	0	0.1
4	HoG_Mar_16	540	High	1	0	0	0.0
5	HoG_TLC_Mar_01	1194	Low	1	1	1	0.0
6	HoG_TLC_Mar_02	540	High	1	0	0	0.0
7	HoG_TLC_Mar_03	540	Moderate	0	0	0	0.7
8	HoG_TLC_Mar_04	540	DNG	0	0	0	0.6
9	HoG_TLC_Mar_05	492	Moderate	1	0	0	0.2
10	HoG-Mar-01	934	High	1	1	0	4.0
11	HoG-Mar-02	1194	Low	0	0	0	1.0
12	HoG-Mar-03	1194	High	1	1	0	0.2
13	HoG-Mar-04	1194	Moderate	1	1	0	10.0
14	HoG-Mar-05	1194	DNG	0	0	0	0.0
15	HoG-Mar-06	931	High	1	0	0	0.0
16	HoG-Mar-07	934	DNG	0	0	0	0.0
17	HoG-Mar-08	1194	Moderate	1	1	0	1.0
18	HoG-Mar-09	931	Moderate	0	1	0	1.0
19	HoG-Mar-10	507	Moderate	1	0	0	1.0
20	HoG-Mar-11	1194	High	1	1	0	0.0
21	HoG-Mar-12	931	Moderate	1	1	0	0.0
22	HoG_TLC_Mar_07	492	Low	1	0	0	0.0
23	HoG_TLC_Mar_06	541	High	1	0	0	0.0
24	HoG_TLC_Mar_08	492	Moderate	1	0	0	0.0
25	NWF_02	934	DNG	0	0	0	0.0
26	NWF_01	1194	High	1	0	0	0.0
27	34784_01	540	Moderate	1	0	0	0.0
28	34784_02	540	Moderate	1	0	1	0.0
29	34784_04	541	Moderate	1	0	0	0.0
30	34784_05	1194	Moderate	1	1	1	0.0
31	34784_07	1194	High	0	1	0	0.0
32	34784_13	540	Low	1	0	0	0.0
33	34784_16	540	Moderate	1	0	0	0.0
34	34784_17	540	High	1	1	0	0.0
35	34784_20	492	High	1	1	1	0.0
36	34784_28	586	Low	0	0	0	0.7
37	34784_34	934	High	1	1	1	0.0
38	34784_42	931	Moderate	1	0	0	0.2
39	34784_46	541	Low	1	0	1	0.0

BDAR plot	Field ID	PCT	Cond.	easting	northing	bearing	Comp Tree	Comp Shrub	Comp Grass	Comp Forbs	Comp Ferns	Comp Other	Struc Tree	Struc Shrub	Struc Grass
40	34784_47	492	High	326274.4454	6516190.311	90	6	7	12	17	0	7	12.6	16.2	24.9
41	34784_08	1194	Low	323360.9042	6499626.655	295	1	2	1	16	1	1	40.0	0.3	1.0
42	34784_09	1194	High	326494.6185	6508705.079	45	1	3	8	16	1	5	1.0	3.0	8.0
43	34784_11	1194	High	317363.9959	6502097.867	310	3	4	5	26	2	4	51.2	5.6	65.1
44	34784_21	492	High	326599.2775	6516285.376	20	3	7	12	19	0	6	18.5	9.9	70.5
45	34784_23	541	High	314558.1255	6504520.94	105	4	6	11	18	2	6	16.2	6.8	97.7
46	34784_27	586	Low	325034.6554	6500182.867	50	0	0	6	10	0	0	0.0	0.0	33.1
47	34784_29	586	Low	325349.1568	6500365.69	232	0	0	9	9	0	0	0.0	0.0	85.2
48	34784_32	541	Moderate	314437.0618	6504741.527	110	2	8	10	20	2	8	40.2	8.6	34.6
49	34784_35	541	High	318203.0365	6501353.959	130	4	8	6	25	1	6	40.0	13.6	52.1
50	34784_45	541	Low	314275.2971	6504581.488	340	3	3	5	20	0	2	50.1	0.3	50.2

BDAR plot	Field ID	PCT	Cond.	Struc Forbs	Struc Ferns	Struc Other	Fun Large Trees	Fun Hollow Trees	Fun Litter Cover	Fun Len Fallen Logs	Fun Tree Stem 5to9	Fun Tree Stem 10to19	Fun Tree Stem 20to29
40	34784_47	492	High	2.8	0.0	1.2	4	1	69.0	15.0	0	1	1
41	34784_08	1194	Low	16.4	0.5	0.1	1	2	27.0	17.0	0	0	1
42	34784_09	1194	High	16.0	1.0	5.0	9	0	52.0	65.0	0	1	1
43	34784_11	1194	High	11.9	2.1	2.6	3	3	58.0	69.0	1	1	1
44	34784_21	492	High	7.3	0.0	2.5	10	0	9.0	47.0	1	1	1
45	34784_23	541	High	9.9	0.2	6.5	0	1	71.0	181.0	1	1	1
46	34784_27	586	Low	7.8	0.0	0.0	0	0	4.0	0.0	0	0	0
47	34784_29	586	Low	6.8	0.0	0.0	0	0	4.0	0.0	0	0	0
48	34784_32	541	Moderate	11.4	1.0	7.2	2	0	70.0	16.0	1	1	1
49	34784_35	541	High	20.7	3.0	8.7	4	0	43.0	139.0	1	1	1
50	34784_45	541	Low	18.0	0.0	1.5	4	0	32.0	58.0	1	1	1

BDAR plot	Field ID	PCT	Cond.	Fun Tree Stem 30to49	Fun Tree Stem 50to79	Fun Tree Regen	Fun High Threat Exotic
40	34784_47	492	High	1	1	1	0.1
41	34784_08	1194	Low	1	1	0	0.0
42	34784_09	1194	High	1	1	0	0.0
43	34784_11	1194	High	1	1	0	0.0
44	34784_21	492	High	1	1	1	0.0
45	34784_23	541	High	1	0	1	0.0
46	34784_27	586	Low	0	0	0	0.0
47	34784_29	586	Low	0	0	0	0.0
48	34784_32	541	Moderate	1	0	1	0.0
49	34784_35	541	High	1	1	0	0.0
50	34784_45	541	Low	1	0	1	0.0

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Family	Scientific name	Common name	Plot 01 (HoG_Mar_13)		Plot 02 (HoG_Mar_14)		Plot 03 (HoG_Mar_15)		Plot 04 (HoG_Mar_16)		Plot 05 (HoG_TLC_Mar_01)		Plot 6 (HoG_TLC_Mar_02)		Plot 07 (HoG_TLC_Mar_03)		Plot 08 (HoG_TLC_Mar_04)		Plot 09 (HoG_TLC_Mar_05)		Plot 10 (HoG-Mar-01)		Plot 11 (HoG-Mar-02)		Plot 12 (HoG-Mar-03)		Plot 13 (HoG-Mar-04)		Plot 14 (HoG-Mar-05)		Plot 15 (HoG-Mar-06)		Plot 16 (HoG-Mar-07)			
			Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.	Cov %	Abnd.		
Asteraceae	Lagenophora stipitata	Common Lagenophora																																		
Asteraceae	Leontodon spp.												0.1	1					0.1	5																
Asteraceae	Olearia alpicola	Alpine Daisy-bush											0.1	1																						
Asteraceae	Olearia elliptica	Sticky Daisy-bush																																		
Asteraceae	Olearia elliptica subsp. Elliptica																																			
Asteraceae	Olearia sp.																																			
Asteraceae	Olearia spp.																																			
Asteraceae	Podolepis sp.																																			
Asteraceae	Senecio diaschides																																			
Asteraceae	Senecio gunnii																	0.1	10																	
Asteraceae	Senecio hispidulus	Hill Fireweed						0.2	4		0.1	1								1	40															
Asteraceae	Senecio linearifolius	Fireweed Groundsel																																		
Asteraceae	Senecio minimus																																			
Asteraceae	Senecio prenanthoides					0.1	1		0.1	7		0.1	10						0.2	20		1	20		0.1	10		0.1	1		0.5	30				
Asteraceae	Senecio quadridentatus	Cotton Fireweed				0.6	10		0.2	5				0.4	10					0.1	1															
Asteraceae	Senecio sp.																																			
Asteraceae	Senecio spp.	Groundsel, Fireweed																																		
Asteraceae	Senecio tenuiflorus											0.3	20																							
Asteraceae	Sigesbeckia orientalis subsp. Orientalis	Indian Weed			5	100		5	200		5	100		0.1	1			1	50		0.3	10		0.6	60											
Asteraceae	Solenogyne gunnii	Solengyne								0.1	1				0.2	20		0.1	2																	
Asteraceae	Vittadinia cervicularis													0.1	1																					
Asteraceae	Vittadinia cuneata	A Fuzzweed																																		
Asteraceae	Vittadinia cuneata var. cuneata	A Fuzzweed												0.1	2				0.3	50																
Asteraceae	Vittadinia hispidula																		0.1	10																
Asteraceae	Vittadinia sp.																																			
Bignoniaceae	Pandorea pandorana	Wonga Wonga Vine													0.1	1																				
Blechnaceae	Blechnum sp.																																			
Boraginaceae	Cynoglossum australe																																			
Boraginaceae	Hackelia latifoli						0.1	2																												
Boraginaceae	Hackelia latifolia				0.4	5														0.1	1															
Campanulaceae	Lobelia concolor	Poison Pratia					0.1	3		0.1	5		0.1	2		0.2	50		0.1	10		0.1	5			0.1	50		0.1	10		2	100		0.1	30
Campanulaceae	Lobelia pedunculata												0.1	10																						
Campanulaceae	Lobelia purpurascens	whiteroot																									0.1	50								
Campanulaceae	Lobelia spp.																										0.5	30								
Campanulaceae	Wahlenbergia gracilis	Sprawling Bluebell			0.1	6		0.1	4										0.1	10					0.1	5										
Campanulaceae	Wahlenbergia stricta	Tall Bluebell																																		
Campanulaceae	Wahlenbergia stricta subsp. Stricta	Tall Bluebell								0.2	6					0.4	30		0.1	10		0.1	3													
Caryophyllaceae	Scleranthus biflorus	Two-flowered Knawel																	0.1	2																
Caryophyllaceae	Stellaria pungens				0.1	1		0.1	5		0.1	3		0.2	10			0.1	10				0.2	30			0.1	10		0.2	50					
Casuarinaceae	Allocasuarina torulosa	Forest Oak																0.1	1																	
Chenopodiaceae	Dysphania pumilio	Small Crumbweed	0.8	20	55	200	2	40		0.1	7		0.3	5			0.5	10		0.4	20		0.1	1			3	100			0.2	100		1	100	
Chenopodiaceae	Einadia hastata	Berry Saltbush	0.3	7																																

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Family	Scientific name	Common name	Plot 35 (34784_20)		Plot 36 (34784_28)		Plot 37 (34784_34)		Plot 38 (34784_42)		Plot 39 (34784_46)		Plot 40 (34784_47)		Plot 41 (34784_08)		Plot 42 (34784_09)		Plot 43 (34784_11)		Plot 44 (34784_21)		Plot 45 (34784_23)		Plot 46 (34784_27)		Plot 47 (34784_29)		Plot 48 (34784_32)		Plot 49 (34784_35)		Plot 50 (34784_45)	
			Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.
Asteraceae	Lagenophora stipitata	Common Lagenophora					0.1	30																										
Asteraceae	Leontodon spp.																																	
Asteraceae	Olearia alpicola	Alpine Daisy-bush																																
Asteraceae	Olearia elliptica	Sticky Daisy-bush	0.1	5									0.5	30																				
Asteraceae	Olearia elliptica subsp. Elliptica																												1					
Asteraceae	Olearia sp.																		1															
Asteraceae	Olearia spp.																																	
Asteraceae	Podolepis sp.											0.1				0.1		0.2		0.1									0.5					
Asteraceae	Senecio diaschides																		1			0.2						0.1		2			2	
Asteraceae	Senecio gunnii																																	
Asteraceae	Senecio hispidulus	Hill Fireweed	0.5	500			0.1	1	0.1	10			0.5	200																				
Asteraceae	Senecio linearifolius	Fireweed Groundsel					0.1	1																										
Asteraceae	Senecio minimus																																	
Asteraceae	Senecio prenanthoides																																	
Asteraceae	Senecio quadridentatus	Cotton Fireweed																																
Asteraceae	Senecio sp.														2		0.1				0.1								0.1		0.5		0.1	
Asteraceae	Senecio spp.	Groundsel, Fireweed	0.1	100									0.5	200																				
Asteraceae	Senecio tenuiflorus																																	
Asteraceae	Sigesbeckia orientalis subsp. Orientalis	Indian Weed	0.1	10															1		0.1		0.1								0.5		0.1	
Asteraceae	Solenogyne gunnii	Solengyne																	0.1		1							0.1		0.1				
Asteraceae	Vittadinia cervicularis																																	
Asteraceae	Vittadinia cuneata	A Fuzzweed	0.1	1									0.1	2																				
Asteraceae	Vittadinia cuneata var. cuneata	A Fuzzweed																																
Asteraceae	Vittadinia hispida																																	
Asteraceae	Vittadinia sp.												0.1							2		0.2							0.1					
Bignoniaceae	Pandorea pandorana	Wonga Wonga Vine					0.2	50																										
Blechnaceae	Blechnum sp.																													0.5				
Boraginaceae	Cynoglossum australe		0.6	10			0.2	50																										
Boraginaceae	Hackelia latifoli																																	
Boraginaceae	Hackelia latifolia														3				1											0.5		0.1		
Campanulaceae	Lobelia concolor	Poison Pratia																																
Campanulaceae	Lobelia pedunculata				0.2	2000																												
Campanulaceae	Lobelia purpurascens	whiteroot					0.1	100			2								0.5			1		0.2		0.2						0.5		
Campanulaceae	Lobelia spp.																																	
Campanulaceae	Wahlenbergia gracilis	Sprawling Bluebell															1					0.2				0.1		0.1		0.1		0.1		
Campanulaceae	Wahlenbergia stricta	Tall Bluebell	0.1	100					0.2	1000			0.1	1000																				
Campanulaceae	Wahlenbergia stricta subsp. Stricta	Tall Bluebell									0.1				0.1		2		0.1		2		0.2					0.2		0.1		0.5		
Caryophyllaceae	Scleranthus biflorus	Two-flowered Knawel																																
Caryophyllaceae	Stellaria pungens		2	100																														
Casuarinaceae	Allocasuarina torulosa	Forest Oak																																
Chenopodiaceae	Dysphania pumilio	Small Crumbweed												2																				
Chenopodiaceae	Einadia hastata	Berry Saltbush																																

[illegible]

Family	Scientific name	Common name	Plot 01 (HoG_Mar_13)		Plot 02 (HoG_Mar_14)		Plot 03 (HoG_Mar_15)		Plot 04 (HoG_Mar_16)		Plot 05 (HoG_TLC_Mar_01)		Plot 6 (HoG_TLC_Mar_02)		Plot 07 (HoG_TLC_Mar_03)		Plot 08 (HoG_TLC_Mar_04)		Plot 09 (HoG_TLC_Mar_05)		Plot 10 (HoG_Mar-01)		Plot 11 (HoG_Mar-02)		Plot 12 (HoG_Mar-03)		Plot 13 (HoG_Mar-04)		Plot 14 (HoG_Mar-05)		Plot 15 (HoG_Mar-06)		Plot 16 (HoG_Mar-07)	
			Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.
Fabaceae	Desmodium varians	Slender Tick-trefoil			0.1	1	0.4	30	0.3	30	0.1	20	3	100	0.5	30			0.3	20	0.5	100						0.5	100					
Fabaceae	Glycine clandestina	Twining glycine											0.1	1											0.1	30	1	100						
Fabaceae	Glycine microphylla	Small-leaf Glycine					0.1	1	0.3	40	0.1	5	0.5	100	0.1	10	0.1	10	0.1	10														
Fabaceae	Glycine tabacina										0.2	50			3	100					1	200												
Fabaceae	Hardenbergia violacea										0.1	1							0.8	10														
Fabaceae	Indigofera australis	Australian Indigo																0.7	20															
Fabaceae	Swainsona galegifolia										0.5	30	0.3	10							0.1	5												
Geraniaceae	Geranium homeanum																																	
Geraniaceae	Geranium potentilloides																						0.2	30	0.1	20	0.2	50	0.5	100	0.5	100	0.1	50
Geraniaceae	Geranium solanderi	Native Geranium																																
Geraniaceae	Geranium solanderi var. solanderi				0.4	30	0.1	6	0.4	30	0.5	50	0.5	20	0.2	20	5	200	0.5	60														
Geraniaceae	Geranium sp.																																	
Geraniaceae	Geranium spp.																				0.2	30												
Haloragaceae	Gonocarpus micranthus										0.1	20																						
Haloragaceae	Gonocarpus sp.																																	
Haloragaceae	Gonocarpus tetragynus	Poverty Raspwort																																
Haloragaceae	Gonocarpus teucrioides	Germander Raspwort																			0.5	50										0.1	40	
Haloragaceae	Haloragis heterophylla	Variable Raspwort																																
Juncaceae	Juncus flavidus										1	40																						
Juncaceae	Juncus subsecundus																																	
Juncaceae	Luzula flaccida	Woodrush																																
Lamiaceae	Ajuga australis	Austral Bugle						0.7	30				0.5	20	1	30	0.1	4																
Lamiaceae	Mentha diemenica	Slender Mint			0.2	10	0.1	1									0.2	10																
Lamiaceae	Mentha satureioides																																	
Lamiaceae	Plectranthus parviflorus																																	
Lamiaceae	Scutellaria humilis	Dwarf Skullcap				0.1	10	0.1	2				1	200	0.1	30			0.5	60														
Lamiaceae	Scutellaria mollis	Soft Skullcap																																
Lamiaceae	Scutellaria sp.																																	
Lomandraceae	Lomandra bracteata	Mat-rush																	0.1	1														
Lomandraceae	Lomandra confertifolia	Matrush											0.1	3																				
Lomandraceae	Lomandra filiformis	Wattle Matt-rush																			0.2	20												
Lomandraceae	Lomandra filiformis subsp. filiformis												0.1	1																				
Lomandraceae	Lomandra filiformis subsp. Flavior							0.5	30	0.1	4	0.1	10				0.1	1																
Lomandraceae	Lomandra glauca	Pale Mat-rush													1	70																		
Lomandraceae	Lomandra longifolia									4	80			40	100				0.2	2	50	300												
Lomandraceae	Lomandra multiflora subsp. Multiflora	Many-flowered Mat-rush																	0.2	2														
Lomandraceae	Lomandra sp.																																	
Loranthaceae	Amyema sp.																																	
Luzuriagaceae	Eustrephus latifolius	Wombat Berry				0.3	7				0.1	2	0.3	5	2	20			1	20														
Luzuriagaceae	Geitonoplesium cymosum	Scrambling Lily																						0.1	3									
Myrsinaceae	Myrsine variabilis																																	
Myrtaceae	Angophora floribunda	Rough-barked Apple																	40	5														
Myrtaceae	Callistemon spp.																																	
Myrtaceae	Eucalyptus blakelyi				5	2																												

Family	Scientific name	Common name	Plot 17 (HoG-Mar-08)		Plot 18 (HoG-Mar-09)		Plot 19 (HoG-Mar-10)		Plot 20 (HoG-Mar-11)		Plot 21 (HoG-Mar-12)		Plot 22 (HoG_TLC_Mar_07)		Plot 23 (HoG_TLC_Mar_06)		Plot 24 (HoG_TLC_Mar_08)		Plot 27 (34784_01)		Plot 28 (34784_02)		Plot 29 (34784_04)		Plot 30 (34784_05)		Plot 31 (34784_07)		Plot 32 (34784_13)		Plot 33 (34784_16)		Plot 34 (34784_17)	
			Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.
Fabaceae	Desmodium varians	Slender Tick-trefoil												1	50							3		0.1										
Fabaceae	Glycine clandestina	Twining glycine					0.1	30	0.1	20			0.1	20			0.1	20	0.5		0.2											0.1	10	
Fabaceae	Glycine microphylla	Small-leaf Glycine																	2		0.2	1							5		0.1	10		
Fabaceae	Glycine tabacina				0.1	10					0.5	20			1	100	0.5	10																
Fabaceae	Hardenbergia violacea						0.1	1											0.5			2							1					
Fabaceae	Indigofera australis	Australian Indigo																	0.2													0.1	1	
Fabaceae	Swainsona galegifolia								0.1	1																								
Geraniaceae	Geranium homeanum																									5	1000							
Geraniaceae	Geranium potentilloides		0.5	50	5	300	1	200	1	30			0.2	10	0.5	10	0.1	10					0.5											
Geraniaceae	Geranium solanderi	Native Geranium																	10		5	5				2		5			0.5	1000		
Geraniaceae	Geranium solanderi var. solanderi																																	
Geraniaceae	Geranium sp.																		2			0.5												
Geraniaceae	Geranium spp.																																	
Haloragaceae	Gonocarpus micranthus																																	
Haloragaceae	Gonocarpus sp.																		0.1											0.1				
Haloragaceae	Gonocarpus tetragynus	Poverty Raspwort																					2											
Haloragaceae	Gonocarpus teucrioides	Germander Raspwort							0.5	20																						0.1	10	
Haloragaceae	Haloragis heterophylla	Variable Raspwort																																
Juncaceae	Juncus flavidus																																	
Juncaceae	Juncus subsecundus																																	
Juncaceae	Luzula flaccida	Woodrush																													0.1	1		
Lamiaceae	Ajuga australis	Austral Bugle																	0.1		0.2	0.1							0.5					
Lamiaceae	Mentha diemenica	Slender Mint																																
Lamiaceae	Mentha satuireioides																		0.1			0.1												
Lamiaceae	Plectranthus parviflorus																					0.1												
Lamiaceae	Scutellaria humilis	Dwarf Skullcap																					0.1											
Lamiaceae	Scutellaria mollis	Soft Skullcap																													0.1	1		
Lamiaceae	Scutellaria sp.																		0.1								0.1							
Lomandraceae	Lomandra bracteata	Mat-rush																																
Lomandraceae	Lomandra confertifolia	Matrush																																
Lomandraceae	Lomandra filiformis	Wattle Matt-rush																	0.1										0.1					
Lomandraceae	Lomandra filiformis subsp. filiformis																														0.1	2		
Lomandraceae	Lomandra filiformis subsp. Flavior																																	
Lomandraceae	Lomandra glauca	Pale Mat-rush																																
Lomandraceae	Lomandra longifolia					0.1	5	0.5	1									3		3			0.1		0.1		0.5			0.1	1			
Lomandraceae	Lomandra multiflora subsp. Multiflora	Many-flowered Mat-rush																	0.1		0.1	0.2			0.1		0.1		0.1					
Lomandraceae	Lomandra sp.																																	
Loranthaceae	Amyema sp.																				0.5													
Luzuriagaceae	Eustrephus latifolius	Wombat Berry																	0.2			0.1									0.1	20		
Luzuriagaceae	Geitonoplesium cymosum	Scrambling Lily									0.1	2					1	5			0.1						0.5							
Myrsinaceae	Myrsine variabilis																				0.1													
Myrtaceae	Angophora floribunda	Rough-barked Apple													20	2						5						50						
Myrtaceae	Callistemon spp.																															0.1	3	
Myrtaceae	Eucalyptus blakelyi																																	

Family	Scientific name	Common name	Plot 35 (34784_20)		Plot 36 (34784_28)		Plot 37 (34784_34)		Plot 38 (34784_42)		Plot 39 (34784_46)		Plot 40 (34784_47)		Plot 41 (34784_08)		Plot 42 (34784_09)		Plot 43 (34784_11)		Plot 44 (34784_21)		Plot 45 (34784_23)		Plot 46 (34784_27)		Plot 47 (34784_29)		Plot 48 (34784_32)		Plot 49 (34784_35)		Plot 50 (34784_45)	
			Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.
Fabaceae	Desmodium varians	Slender Tick-trefoil									0.1						3				1		5							0.5			1	
Fabaceae	Glycine clandestina	Twining glycine	0.2	200									0.1	100						0.1		0.2							0.5					
Fabaceae	Glycine microphylla	Small-leaf Glycine					0.1	10									1		1		1		0.5						1		2			
Fabaceae	Glycine tabacina																																	
Fabaceae	Hardenbergia violacea												0.5	100							0.2								3		1			
Fabaceae	Indigofera australis	Australian Indigo																			0.2								3		0.1			
Fabaceae	Swainsona galegifolia		0.3	20			0.1	1					0.2	200																				
Geraniaceae	Geranium homeanum						0.2	1000	0.3	1000			0.2	1000																				
Geraniaceae	Geranium potentilloides																																	
Geraniaceae	Geranium solanderi	Native Geranium	0.2	1000			0.1	500			0.1				2		3		2		0.1					5		5		3		5		
Geraniaceae	Geranium solanderi var. solanderi																																	
Geraniaceae	Geranium sp.																		0.1		0.1									0.1		1		
Geraniaceae	Geranium spp.																																	
Haloragaceae	Gonocarpus micranthus																																	
Haloragaceae	Gonocarpus sp.																			0.1							0.1				0.1			
Haloragaceae	Gonocarpus tetragynus	Poverty Raspwort							0.1	1																								
Haloragaceae	Gonocarpus teucrioides	Germaner Raspwort																																
Haloragaceae	Haloragis heterophylla	Variable Raspwort			0.2	5000																												
Juncaceae	Juncus flavidus																																	
Juncaceae	Juncus subsecundus																					0.1		20		2								
Juncaceae	Luzula flaccida	Woodrush	0.1	1																														
Lamiaceae	Ajuga australis	Austral Bugle												0.1								0.1					0.1		0.1					
Lamiaceae	Mentha diemenica	Slender Mint									0.1									0.1														
Lamiaceae	Mentha satuireioides																			0.5														
Lamiaceae	Plectranthus parviflorus																					0.1												
Lamiaceae	Scutellaria humilis	Dwarf Skullcap							0.1	3																								
Lamiaceae	Scutellaria mollis	Soft Skullcap																																
Lamiaceae	Scutellaria sp.																					0.1			0.1									
Lomandraceae	Lomandra bracteata	Mat-rush																																
Lomandraceae	Lomandra confertifolia	Matrush																																
Lomandraceae	Lomandra filiformis	Wattle Matt-rush																															0.1	
Lomandraceae	Lomandra filiformis subsp. filiformis																																	
Lomandraceae	Lomandra filiformis subsp. Flavior																																	
Lomandraceae	Lomandra glauca	Pale Mat-rush																																
Lomandraceae	Lomandra longifolia		0.5	20			85	2000												5							1							
Lomandraceae	Lomandra multiflora subsp. Multiflora	Many-flowered Mat-rush	2	100									1	200			0.1			3		0.1					0.5		0.1			0.1		
Lomandraceae	Lomandra sp.																			0.1														
Loranthaceae	Amyema sp.																														0.1			
Luzuriagaceae	Eustrephus latifolius	Wombat Berry	0.1	10			0.1	10					0.2	50			0.1		1		0.1		0.2					1		5		0.5		
Luzuriagaceae	Geitonoplesium cymosum	Scrambling Lily	0.1	10									0.1	20																				
Myrsinaceae	Myrsine variabilis																																	
Myrtaceae	Angophora floribunda	Rough-barked Apple	0.3	2									0.1	1								10						40						
Myrtaceae	Callistemon spp.						40	200																										
Myrtaceae	Eucalyptus blakelyi																																	

Family	Scientific name	Common name	Plot 01 (HoG_Mar_13)		Plot 02 (HoG_Mar_14)		Plot 03 (HoG_Mar_15)		Plot 04 (HoG_Mar_16)		Plot 05 (HoG_TLC_Mar_01)		Plot 6 (HoG_TLC_Mar_02)		Plot 7 (HoG_TLC_Mar_03)		Plot 8 (HoG_TLC_Mar_04)		Plot 9 (HoG_TLC_Mar_05)		Plot 10 (HoG-Mar-01)		Plot 11 (HoG-Mar-02)		Plot 12 (HoG-Mar-03)		Plot 13 (HoG-Mar-04)		Plot 14 (HoG-Mar-05)		Plot 15 (HoG-Mar-06)		Plot 16 (HoG-Mar-07)		
			Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	
			Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %
Poaceae	Anthosachne scabra	Wheatgrass, Common Wheatgrass											0.1	1																					
Poaceae	Aristida lignosa															5	100																		
Poaceae	Aristida ramosa	Purple Wiregrass							0.1	2				0.1	10	4	200																		
Poaceae	Aristida sp.																																		
Poaceae	Aristida vagans	Threeawn Speargrass												0.1	1																				
Poaceae	Austrostipa scabra	Speargrass																																	
Poaceae	Austrostipa verticillata	Slender Bamboo Grass																																	
Poaceae	Bothriochloa decipiens var. decipiens	Pitted Bluegrass							0.1	3				20	500	15	300	75	700	1	50														
Poaceae	Bothriochloa macra	Red Grass																																	
Poaceae	Chloris truncata	Windmill Grass																																	
Poaceae	Cymbopogon refractus	Barbed Wire Grass												0.5	30	4	100																		
Poaceae	Dichelachne crinita	Longhair Plumegrass																		0.3	50														
Poaceae	Dichelachne micrantha	Shorthair Plumegrass												0.1	1																				
Poaceae	Echinopogon cheelii	Long-flowered Hedgehog Grass																																	
Poaceae	Echinopogon mckiei						0.1	2	5	200	0.2	20	0.3	30	0.1	1				3	300														
Poaceae	Echinopogon ovatus	Forest Hedgehog Grass																							2	50	2	100							
Poaceae	Enneapogon nigricans	Niggerheads																	0.2	10															
Poaceae	Entolasia marginata										1	70																							
Poaceae	Eragrostis leptostachya	Paddock Lovegrass																0.5	50																
Poaceae	Eragrostis sp.																																		
Poaceae	Eriochloa pseudoacrotricha	Early Spring Grass																														10	200		
Poaceae	Imperata cylindrica	Blady Grass																	6	200															
Poaceae	Lachnagrostis aemula																																		
Poaceae	Lachnagrostis filiformis																																		
Poaceae	Microlaena stipoides	Weeping Grass	5	100	10	200	2	80	0.3	20	1	50	5	300	2	100	25	1000	5	400	5	300	20	1000	0.5	30	1	100	50	2000	5	500	10	500	
Poaceae	Microlaena stipoides var. stipoides																																		
Poaceae	Panicum effusum																																		
Poaceae	Panicum simile																																		
Poaceae	Paspalum distichum	Water Couch																																	
Poaceae	Pennisetum alopecuroides																																		
Poaceae	Poa labillardierei var. labillardierei	Tussock	1	30	4	40	0.5	20	2	50	0.3	30	0.5	30	1	30	4	50	10	400															
Poaceae	Poa sieberiana	Snowgrass											5	100	0.5	20																			
Poaceae	Poa sieberiana var. sieberiana	Snowgrass																			30	100			80	500				1	20	1	20		
Poaceae	Rytidosperma carphoides	Short Wallaby Grass																			0.1	50													
Poaceae	Rytidosperma laeve	Wallaby Grass					0.1	10					0.1	10			3	200																	
Poaceae	Rytidosperma penicillatum	Slender Wallaby Grass							0.3	20			1	100	0.2	20	0.5	50	0.3	40															

[illegible]

[illegible]

Family	Scientific name	Common name	Plot 17 (HoG-Mar-08)		Plot 18 (HoG-Mar-09)		Plot 19 (HoG-Mar-10)		Plot 20 (HoG-Mar-11)		Plot 21 (HoG-Mar-12)		Plot 22 (HoG_TLC_Mar_07)		Plot 23 (HoG_TLC_Mar_06)		Plot 24 (HoG_TLC_Mar_08)		Plot 27 (34784_01)		Plot 28 (34784_02)		Plot 29 (34784_04)		Plot 30 (34784_05)		Plot 31 (34784_07)		Plot 32 (34784_13)		Plot 33 (34784_16)		Plot 34 (34784_17)	
			Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.
Poaceae	Rytidosperma pilosum	Smooth-flowered Wallaby Grass													1		50																	
Poaceae	Rytidosperma racemosum	Wallaby Grass																														0.1	10	
Poaceae	Rytidosperma racemosum var. racemosum	Wallaby Grass																		0.1					0.1									
Poaceae	Rytidosperma sp.																	0.5					0.5											
Poaceae	Rytidosperma spp.																																	
Poaceae	Sorghum leiocladum	Wild Sorghum																																
Poaceae	Sporobolus creber																																	
Poaceae	Themeda triandra																						20											
Polygalaceae	Polygala japonica	Dwarf Milkwort																																
Polygonaceae	Persicaria decipiens	Slender Knotweed																																
Polygonaceae	Persicaria hydropiper	Water Pepper																														0.2	30	
Polygonaceae	Rumex brownii	Swamp Dock											0.5	5	1	20	1	20																
Portulacaceae	Portulaca oleracea	Pigweed																																
Proteaceae	Lomatia arborescens	Tree Lomatia							2	10																								
Pteridaceae	Cheilanthes sieberi																							0.2										
Pteridaceae	Cheilanthes sieberi subsp. Sieberi	Rock Fern																																
Ranunculaceae	Clematis aristata	Old Man's Beard																		0.1		0.5						1						
Ranunculaceae	Clematis glycinoides	Headache Vine																																
Ranunculaceae	Ranunculus lappaceus																																	
Ranunculaceae	Ranunculus sessiliflorus	Small-flowered Buttercup																																
Rosaceae	Acaena novae-zelandiae	Bidgee-widgee	0.1	10			0.5	20	0.5	10	0.5	20	0.5	10	0.1	5				0.5				1				0.2		0.2	50			
Rosaceae	Acaena ovina	Acaena																																
Rosaceae	Rubus parvifolius	Native Raspberry									0.5	5			0.2	5			2	0.2	0.5	0.5				1				0.1	5			
Rubiaceae	Asperula conferta	Common Woodruff					0.5	100	0.5	40			0.1	10					0.1	0.1	0.1	0.1		0.1	10	1		0.1						
Rubiaceae	Asperula scoparia																																	
Rubiaceae	Coprosma quadrifida	Prickly Currant Bush																		0.5											0.1	1		
Rubiaceae	Galium binifolium																									0.1	2							
Rubiaceae	Galium binifolium subsp. Binifolium																														0.1	2		
Rubiaceae	Galium ciliare																																	
Rubiaceae	Galium gaudichaudii	Rough Bedstraw									1	30																						
Rubiaceae	Galium gaudichaudii subsp. gaudichaudii																														0.1	20		
Rubiaceae	Galium leiocarpum						2	100																										
Rubiaceae	Galium leptogonium																														0.1	10		
Rubiaceae	Galium sp. 2																																	
Rubiaceae	Morinda jasminoides																					0.1				0.1								
Santalaceae	Exocarpos cupressiformis	Cherry Ballart																	0.1			0.1												
Sapindaceae	Dodonaea viscosa subsp. Spatulata	Broad-leaf Hopbush													10	20																		
Smilacaceae	Smilax australis	Lawyer Vine	0.5	10	0.5	4	0.1	5											0.1						0.1	1					0.1	5		
Solanaceae	Solanum aviculare	Kangaroo Apple			0.5	5																					0.1		0.2					

Family	Scientific name	Common name	Plot 35 (34784_20)		Plot 36 (34784_28)		Plot 37 (34784_34)		Plot 38 (34784_42)		Plot 39 (34784_46)		Plot 40 (34784_47)		Plot 41 (34784_08)		Plot 42 (34784_09)		Plot 43 (34784_11)		Plot 44 (34784_21)		Plot 45 (34784_23)		Plot 46 (34784_27)		Plot 47 (34784_29)		Plot 48 (34784_32)		Plot 49 (34784_35)		Plot 50 (34784_45)	
			Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.
Poaceae	Rytidosperma pilosum	Smooth-flowered Wallaby Grass																																
Poaceae	Rytidosperma racemosum	Wallaby Grass	5	500									5	500																				
Poaceae	Rytidosperma racemosum var. racemosum	Wallaby Grass									0.2									0.1														
Poaceae	Rytidosperma sp.										10						0.5				5							1		2				
Poaceae	Rytidosperma spp.				0.1	20																												
Poaceae	Sorghum leiocladum	Wild Sorghum																																
Poaceae	Sporobolus creber										1																							
Poaceae	Themeda triandra		0.2	100									0.5	30							15		35						10					
Polygalaceae	Polygala japonica	Dwarf Milkwort																																
Polygonaceae	Persicaria decipiens	Slender Knotweed			0.1	30																		3										
Polygonaceae	Persicaria hydropiper	Water Pepper																					2			0.2								
Polygonaceae	Rumex brownii	Swamp Dock			0.1	10	0.1	1						3				0.1				2			0.5									
Portulacaceae	Portulaca oleracea	Pigweed																																
Proteaceae	Lomatia arborescens	Tree Lomatia																																
Pteridaceae	Cheilanthes sieberi		0.1	10																														
Pteridaceae	Cheilanthes sieberi subsp. Sieberi	Rock Fern																																
Ranunculaceae	Clematis aristata	Old Man's Beard											0.1	2		0.1				0.1		0.5						1		0.5				
Ranunculaceae	Clematis glycinoides	Headache Vine	0.1	20									0.1	30																				
Ranunculaceae	Ranunculus lappaceus																						0.1											
Ranunculaceae	Ranunculus sessiliflorus	Small-flowered Buttercup																																
Rosaceae	Acaena novae-zelandiae	Bidgee-widgee												0.5		0.5		1		0.1										0.5		0.5		
Rosaceae	Acaena ovina	Acaena																																
Rosaceae	Rubus parvifolius	Native Raspberry					0.1	10								0.2		0.5				0.2				0.5		0.1						
Rubiaceae	Asperula conferta	Common Woodruff							0.2	500	0.1					0.1				0.2		0.1						0.1			1			
Rubiaceae	Asperula scoparia																																	
Rubiaceae	Coprosma quadrifida	Prickly Currant Bush					20	500																										
Rubiaceae	Galium binifolium		0.2	200																														
Rubiaceae	Galium binifolium subsp. Binifolium						0.1	30					0.1	50																				
Rubiaceae	Galium ciliare		0.1	2																														
Rubiaceae	Galium gaudichaudii	Rough Bedstraw							0.1	10																								
Rubiaceae	Galium gaudichaudii subsp. gaudichaudii																																	
Rubiaceae	Galium leiocarpum																																	
Rubiaceae	Galium leptogonium						0.1	20																										
Rubiaceae	Galium sp. 2																																	
Rubiaceae	Morinda jasminoides																											0.1		0.1				
Santalaceae	Exocarpos cupressiformis	Cherry Ballart	1	1									0.1	1						0.1		0.1					0.5					0.1		
Sapindaceae	Dodonaea viscosa subsp. Spatulata	Broad-leaf Hopbush																																
Smilacaceae	Smilax australis	Lawyer Vine					0.2	20							0.1		0.5		0.5			0.1									0.1			
Solanaceae	Solanum aviculare	Kangaroo Apple																	0.1												1			

Family	Scientific name	Common name	Plot 01 (HoG_Mar_13)		Plot 02 (HoG_Mar_14)		Plot 03 (HoG_Mar_15)		Plot 04 (HoG_Mar_16)		Plot 05 (HoG_TLC_Mar_01)		Plot 6 (HoG_TLC_Mar_02)		Plot 7 (HoG_TLC_Mar_03)		Plot 8 (HoG_TLC_Mar_04)		Plot 9 (HoG_TLC_Mar_05)		Plot 10 (HoG_Mar-01)		Plot 11 (HoG_Mar-02)		Plot 12 (HoG_Mar-03)		Plot 13 (HoG_Mar-04)		Plot 14 (HoG_Mar-05)		Plot 15 (HoG_Mar-06)		Plot 16 (HoG_Mar-07)			
			Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.		
Solanaceae	Solanum linearifolium	Mountain Kangaroo Apple																																		
Solanaceae	Solanum papaverifolium																																			
Solanaceae	Solanum prinophyllum	Forest Nightshade																			1	10	0.1	5					0.1	20	2	30				
Solanaceae	Solanum pungetium	Eastern Nightshade																																		
Solanaceae	Solanum sp.	A Nightshade																																		
Sterculiaceae	Brachychiton populneus subsp. Populneus																																			
Thymelaeaceae	Pimelea curviflora var. divergens											0.2	20				0.9	70	0.1	1																
Thymelaeaceae	Pimelea curviflora var. gracilis											0.7	60				0.2	40																		
Thymelaeaceae	Pimelea linifolia																																			
Urticaceae	Urtica incisa	Stinging Nettle	0.5	10			0.5	40				0.2	10								0.1	3	20	300	0.5	10	1	20	10	200	70	500	2	50		
Violaceae	Hybanthus monopetalus	Slender Violet-bush																																		
Violaceae	Melicytus dentatus	Tree Violet					20	40				0.1	1						0.7	1	1	5	60	50	3	7	10	10			1	3				
Violaceae	Viola betonicifolia								0.1	3		0.1	4	0.5	50	0.1	10		0.1	3	1	50			1	100	0.5	30								
Violaceae	Viola hederacea	Ivy-leaved Violet																																		
Violaceae	Viola sp.																																			
Zamiaceae	Macrozamia concinna	Rainforest species																																		
Zamiaceae	Rainforest sp.																																			
Introduced Species																																				
Amaranthaceae	Amaranthus viridis	Green Amaranth	0.1	1																																
Amygdalaceae	Prunus avium																		0.1	1																
Apiaceae	Cyclospermum leptophyllum	Slender Celery					0.1	5									0.1	10	0.2	20																
Apocynaceae	Gomphocarpus fruticosus	Narrow-leaved Cotton Bush																																		
Asteraceae	Ambrosia artemisifolia	Annual Ragweed																																		
Asteraceae	Bidens pilosa	Cobbler's Pegs						1	50			0.1	10	20	200				0.1	10																
Asteraceae	Bidens sp.																																			
Asteraceae	Bidens subalternans	Greater Beggar's Ticks					1	50	0.1	1																										
Asteraceae	Bidens tripartita						3	100																												
Asteraceae	Carthamus lanatus	Saffron Thistle					0.1	2									0.1	3																		
Asteraceae	Centaurea spp.	Thistle																																		
Asteraceae	Cirsium spp.																																			
Asteraceae	Cirsium vulgare	Spear Thistle	2	40	2	40	0.3	10	0.2	30	5	50			0.1	10	2	50	0.2	30							0.1	5	1	100	0.1	10	1	20		
Asteraceae	Conyza bonariensis	Flaxleaf Fleabane			2	30											1	100	0.1	10																
Asteraceae	Conyza parva	Fleabane															0.1	10																		
Asteraceae	Conyza sp.																																			
Asteraceae	Gamochaeta purpurea	Purple Cudweed															0.1	6																		
Asteraceae	Hypochaeris radicata	Catsear	0.2	10	2	40			0.1	1	0.1	1	0.1	4	0.1	2	1	80	0.2	20	0.5	20	0.1	10	0.2	10	0.1	10				1	20			
Asteraceae	Senecio madagascariensis	Fireweed				0.1	1																													
Asteraceae	Sonchus asper											0.1	10																							
Asteraceae	Sonchus oleraceus	Common Sowthistle					0.1	2						0.1	1				0.3	40																
Asteraceae	Sonchus sp.																																			

Family	Scientific name	Common name	Plot 17 (HoG-Mar-08)		Plot 18 (HoG-Mar-09)		Plot 19 (HoG-Mar-10)		Plot 20 (HoG-Mar-11)		Plot 21 (HoG-Mar-12)		Plot 22 (HoG_TLC_Mar_07)		Plot 23 (HoG_TLC_Mar_06)		Plot 24 (HoG_TLC_Mar_08)		Plot 27 (34784_01)		Plot 28 (34784_02)		Plot 29 (34784_04)		Plot 30 (34784_05)		Plot 31 (34784_07)		Plot 32 (34784_13)		Plot 33 (34784_16)		Plot 34 (34784_17)	
			Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.	Cur %	Abnd.
Solanaceae	Solanum linearifolium	Mountain Kangaroo Apple																																
Solanaceae	Solanum papaverifolium													0.5	10	0.5	5																	
Solanaceae	Solanum prinophyllum	Forest Nightshade			0.5	10	0.5	10																										
Solanaceae	Solanum pungetium	Eastern Nightshade																																
Solanaceae	Solanum sp.	A Nightshade																0.5		0.1			0.1				2		0.2					
Sterculiaceae	Brachychiton populneus subsp. Populneus																																	
Thymelaeaceae	Pimelea curviflora var. divergens																																	
Thymelaeaceae	Pimelea curviflora var. gracilis																																	
Thymelaeaceae	Pimelea linifolia																																	
Urticaceae	Urtica incisa	Stinging Nettle	10	50	0.5	10					5	20	1	6	1	10	30	100		0.1		0.1		0.1				0.1		0.1	0.2			
Violaceae	Hybanthus monopetalus	Slender Violet-bush																							0.1	1								
Violaceae	Melicytus dentatus	Tree Violet	8	7	3	20					0.5	10	3	2	10	50	50	40		0.3		0.1					1					0.1	5	
Violaceae	Viola betonicifolia						0.5	20	0.5	30	1	30							0.5		0.1		0.1									0.1	2	
Violaceae	Viola hederacea	Ivy-leaved Violet	0.5	20			0.1	20	0.1	20																								
Violaceae	Viola sp.																																	
Zamiaceae	Macrozamia concinna																																	
	Rainforest species																																	
	Rainforest sp.																																	
Introduced Species																																		
Amaranthaceae	Amaranthus viridis	Green Amaranth																																
Amygdalaceae	Prunus avium																																	
Apiaceae	Cyclospermum leptophyllum	Slender Celery																														0.1	2	
Apocynaceae	Gomphocarpus fruticosus	Narrow-leaved Cotton Bush																																
Asteraceae	Ambrosia artemisifolia	Annual Ragweed																																
Asteraceae	Bidens pilosa	Cobbler's Pegs			0.2	20	0.1	20	0.1	5						0.1	5																	
Asteraceae	Bidens sp.																					0.2						5		0.5				
Asteraceae	Bidens subalternans	Greater Beggar's Ticks																																
Asteraceae	Bidens tripartita																																	
Asteraceae	Carthamus lanatus	Saffron Thistle																																
Asteraceae	Centaurea spp.	Thistle																																
Asteraceae	Cirsium spp.																																	
Asteraceae	Cirsium vulgare	Spear Thistle	1	10	1	30	1	50	0.1	10	1	30			1	10	1	10	0.5		0.1		0.1			0.1	5	0.5		0.5		0.1	10	
Asteraceae	Conyza bonariensis	Flaxleaf Fleabane					0.5	10																		0.1	1					0.1	1	
Asteraceae	Conyza parva	Fleabane																																
Asteraceae	Conyza sp.																		1		0.2							1		1				
Asteraceae	Gamochaeta purpurea	Purple Cudweed																																
Asteraceae	Hypochaeris radicata	Catsear							1	10					1	30			0.5		0.2		0.1		0.2		0.1	1	0.5		0.1		0.1	10
Asteraceae	Senecio madagascariensis	Fireweed																																
Asteraceae	Sonchus asper																																	
Asteraceae	Sonchus oleraceus	Common Sowthistle																																
Asteraceae	Sonchus sp.																																	

[illegible]

[illegible]

[illegible]

[illegible]

Family	Scientific name	Common name	Plot 35 (34784_20)		Plot 36 (34784_28)		Plot 37 (34784_34)		Plot 38 (34784_42)		Plot 39 (34784_46)		Plot 40 (34784_47)		Plot 41 (34784_08)		Plot 42 (34784_09)		Plot 43 (34784_11)		Plot 44 (34784_21)		Plot 45 (34784_23)		Plot 46 (34784_27)		Plot 47 (34784_29)		Plot 48 (34784_32)		Plot 49 (34784_35)		Plot 50 (34784_45)		
			Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	Cvr %	Abnd.	
Plantaginaceae	Plantago lanceolata	Lamb's Tongues									0.5																							1	
Poaceae	Anthoxanthum odoratum	Sweet Vernal Grass							20		6000							5						1			0.2		1		5		1		
Poaceae	Avena sp.																					0.5					0.1								
Poaceae	Axonopus compressus	Broad-leaved Carpet Grass																																	
Poaceae	Bromus catharticus	Praire Grass																																	
Poaceae	Bromus sp.																											2		2					
Poaceae	Cenchrus clandestinus	Kikuyu Grass			0.5		100								1														2		2				
Poaceae	Dactylis glomerata	Cocksfoot			30		1000	0.2		100																									
Poaceae	Echinochloa crusgalli	Barneyard Grass			0.1		10																												
Poaceae	Ehrharta calycina	Perennial Veldtgrass																																	
Poaceae	Ehrharta erecta																																	1	
Poaceae	Eleusine tristachya																						1												
Poaceae	Holcus lanatus	Yorkshire Fog																																	
Poaceae	Lolium multiflorum	Italian Ryegrass																																	
Poaceae	Lolium perenne	Perennial Ryegrass			0.1		10																												
Poaceae	Lolium sp.																																		
Poaceae	Lolium spp.	A Ryegrass																					20			0.1				1					
Poaceae	Panicum gilvum				0.1		30																												
Poaceae	Paspalum dilatatum	Paspalum			0.1		10																												
Poaceae	Paspalum urvillei	Vasey Grass									6													25		5		0.5						1	
Poaceae	Poa annua																								2										
Poaceae	Poaceae indeterminate	Grasses, reeds and bamboos																																	
Poaceae	Setaria parviflora																																		
Poaceae	Sporobolus africanus																																		
Poaceae	Vulpia bromoides	Squirrel Tail Fesque																																	
Polygonaceae	Acetosella vulgaris	Sheep Sorrel								0.2		1000																							
Polygonaceae	Polygonum aviculare	Wireweed			0.2		1000																												
Polygonaceae	Rumex obtusifolius	Broadleaf Dock																																	
Primulaceae	Lysimachia arvensis	Scarlet Pimpernel																																	
Rosaceae	Rosa rubiginosa	Sweet Briar																						0.1		0.2					0.1				
Rosaceae	Rubus fruticosus sp. agg.	Blackberry complex			0.1		1		0.1		2			10		0.1		1					1					60		0.1			0.5		
Rubiaceae	Galium aparine	Goosegrass																																	
Rubiaceae	Galium divaricatum	Slender Bedstraw																																	
Rubiaceae	Galium spurium																																		
Rutaceae	Rutaceae indeterminate																																		
Salicaceae	Populus sp.																																	0.2	
Scrophulariaceae	Parentucellia latifolia	Red Bartsia																																	
Scrophulariaceae	Verbascum blattaria																																		
Scrophulariaceae	Verbascum thapsus subsp. Thapsus	Great Mullein	0.1	2								0.1			0.1	2	0.1					0.1					0.1		0.1						
Scrophulariaceae	Verbascum virgatum	Twiggy Mullein										0.1																							
Solanaceae	Solanum chenopodioides																																		

[illegible]

[illegible]

[illegible]

Appendix I AECOM 330 kV Overhead Line vegetation Clearance Report

Hills of Gold Wind Farm

330 kV Overhead Line Vegetation Clearance Report

13-Aug-2021

Hills of Gold Wind Farm Grid Connection

Hills of Gold Wind Farm

330 kV Overhead Line Vegetation Clearance Report

Client: Hills of Gold Wind Farm Pty Limited

ABN: 28 145 173 324

Prepared by

AECOM Australia Pty Ltd

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13-Aug-2021

Job No.: 60633842

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Quality Information

Document Hills of Gold Wind Farm
60633842

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Date 13-Aug-2021

Prepared by Hew McLachlan

Reviewed by Rita Bu, Caitie McClelland, Rajesh Arora

Revision History


Rev	Revision Date	Details	Authorised	
			Name/Position	Signature
0	13-Aug-2021	For Issue	Rajesh Arora Technical Director - Power and Industrial	

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1.0 Overview

Hills of Gold Wind Farm Pty Limited is considering the development of Hills of Gold Wind Farm located 60 km south-east of Tamworth New South Wales, Australia. The wind farm has a proposed capacity of 400 MW and is proposed to connect to the TransGrid network.

The project is currently in the early stages of development for the 330 kV connection interface between a new TransGrid 330 kV switching station and a new 330 kV transmission line (20 km in length) to the Hills of Gold new collector 330/33 kV substation located within the boundaries of the proposed wind farm development.

This report discusses the vegetation clearing that may be required based on an indicative transmission line concept design and route.

2.0 Scope of Works – 330 kV Concept Vegetation Clearing

The agreed scope is as follows:

- Update the transmission line alignment
- Review the Lidar Survey to incorporate the vegetation data for upload into PLS-CADD.
- Prepare the PLS-CADD Maximum design temperature bottom conductor profile with a vegetation clearance line – 3.0 metres plus 1.0 metre regrowth (minimum 4.0 metre required)
- Prepare the structure clearance envelope of 60.0 metres around structure locations.
- Prepare the 50 degree C and 500 pa vegetation clearance envelopes for all spans on the Hills of Gold 330 kV transmission line.

3.0 Key Deliverables

The agreed deliverables are as follows:

- Basis of Design summary on vegetation Standards as per AS7000, inclusive of a table of vegetation clearance locations with chainages. (TransGrid reference standard to be included)
- Shapefiles (and kmz) of the 60 meter transmission line corridor with vegetation clearance locations identified
- Updated plan and profile cross sections with vegetation canopy height and calculated clearance envelopes overlaid. Locations where these overlays intersect requiring vegetation clearance will be clearly shown with chainages

3.1 Concept Design

The PLS-CADD transmission line model has been updated based on the concept design prepared last year (2020). The updated model includes the recently received ground and vegetation lidar, as well as the received alignment centreline which includes both minor and major changes for almost the entirety of the route compared to the previous design. As this is a concept design and the terrain is highly challenging, a single indicative strain pole arrangement has been used for the design to determine the vegetation clearing that may be required. The pole height has been adjusted for the terrain and is a maximum of 32 m.

3.2 Vegetation Clearing Criteria

Vegetation clearing has been assessed with the transmission line at 85°C no wind, and 500 Pa at 50°C. Vegetation within 4 m of the conductors under these conditions have been identified as requiring clearing. This 4 m value is comprised of a 3 m clearance zone plus a 1 m regrowth zone. Although there is no vegetation clearance requirement in AS/NZS 7000, SA/SNZ HB 331 notes 3 m clearance is

required up to 33 kV. We are aware that TransGrid has previously used 7.5 m clearance required for 330 kV lines.

3.3 Vegetation Clearing Results

The vegetation clearing identified from this analysis is shown in the plan and profile drawings, the shapefile, the Google Earth KMZ file, and is summarised in the table below. The approximate total area of vegetation clearance is 193,000 m².

Table 1 Vegetation Clearing Results

Site #	Site X (m)	Site Y (m)	Chainage (m)	Approx. Area (m ²)
1	303643	6510443	181	574
2	303709	6510386	269	137
3	303826	6510267	435	2086
4	303906	6510165	565	99
5	304075	6509984	812	1750
6	304297	6509750	1135	178
7	304399	6509643	1283	126
8	304550	6509456	1522	202
9	305242	6508765	2501	31
10	305905	6508233	3351	135
11	305730	6508359	3136	1075
12	305960	6508192	3419	100
13	306007	6508151	3482	148
14	306149	6508048	3658	486
15	306697	6507756	4282	2574
16	307859	6507182	5590	172
17	307958	6507138	5698	262
18	306939	6507680	4536	590
19	307009	6507644	4613	2
20	306830	6507718	4420	7
21	311408	6506370	9286	2686
22	309377	6506829	7151	56
23	309466	6506820	7241	317
24	309275	6506847	7048	249
25	308986	6506886	6756	2766
26	308847	6506911	6614	1109
27	310450	6506707	8232	580
28	310089	6506750	7868	13
29	310179	6506726	7960	537
30	310360	6506705	8142	69
31	311065	6506619	8852	4644
32	310880	6506661	8664	8
33	310801	6506674	8583	24
34	310744	6506680	8526	5

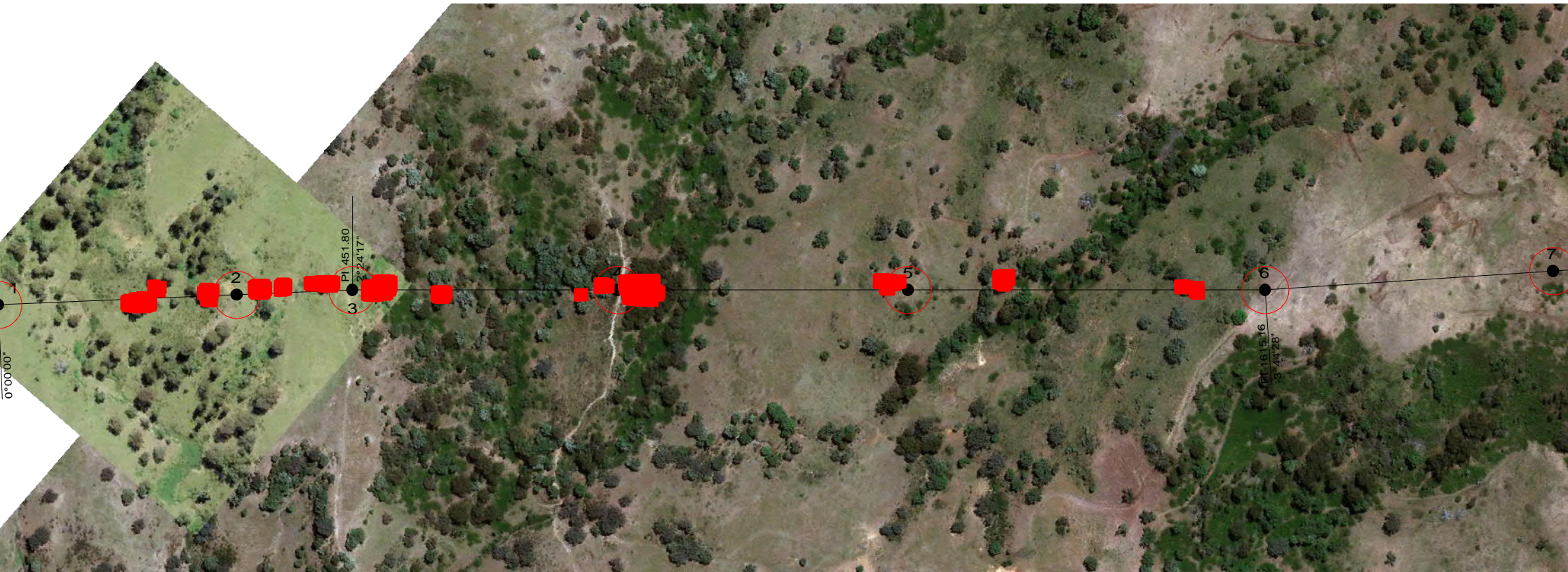
Site #	Site X (m)	Site Y (m)	Chainage (m)	Approx. Area (m ²)
35	310692	6506668	8476	17
36	311833	6505783	10010	618
37	311974	6505566	10269	1459
38	312067	6505446	10421	4358
39	311930	6505654	10171	56
40	312319	6505262	10739	22
41	312473	6505170	10919	2924
42	312997	6505008	11468	2406
43	313372	6504903	11858	49
44	313483	6504861	11976	2665
45	313186	6504938	11669	5971
46	318569	6502287	18095	4563
47	318983	6502368	18517	3087
48	319176	6502420	18716	186
49	319381	6502349	18955	1218
50	319449	6502273	19057	0
51	315456	6504169	14136	3689
52	313898	6504723	12413	4285
53	314119	6504657	12644	3340
54	314746	6504524	13285	26021
55	314272	6504621	12801	1108
56	315589	6503850	14481	3303
57	316011	6503389	15119	23765
58	316553	6503014	15778	2762
59	316745	6502882	16012	2729
60	316969	6502690	16307	796
61	317184	6502493	16599	454
62	317264	6502422	16707	1944
63	318369	6502242	17890	4794
64	317933	6502145	17443	2075
65	317689	6502207	17184	2
66	320967	6500986	21164	3081
67	321173	6500939	21375	598
68	319767	6502003	19473	277
69	319829	6501945	19558	587
70	319984	6501817	19759	2632
71	319506	6502242	19120	0
72	319662	6502106	19327	0
73	320241	6501384	20265	3092
74	320611	6501059	20801	6348
75	320830	6500419	25321	874

Site #	Site X (m)	Site Y (m)	Chainage (m)	Approx. Area (m ²)
76	322145	6500385	22502	23
77	321359	6500873	21574	219
78	321504	6500803	21737	3338
79	321139	6500815	24819	958
80	321075	6500711	24942	2
81	322063	6500443	22402	1092
82	322420	6500145	22870	7574
83	322966	6499819	23537	23167
84	323336	6499611	24027	4401

Appendix A

Plan and Profile Drawing

Appendix A Plan and Profile Drawing



PROJECT
HILLS OF GOLD WIND FARM

CLIENT

CONSULTANT
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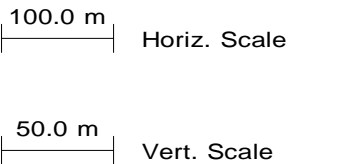
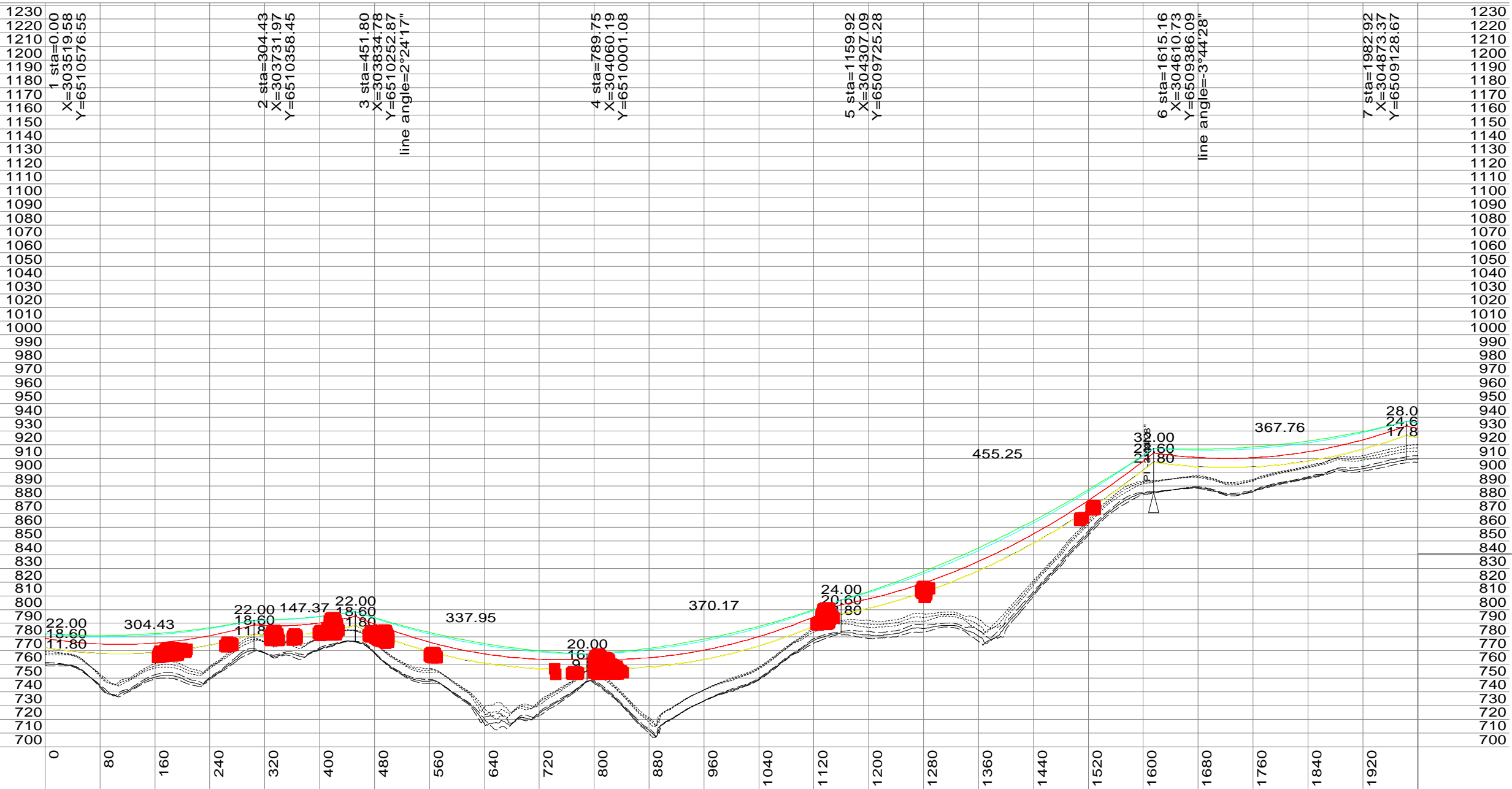
PROJECT NUMBER
60633842

SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 1 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
1. TWIN CONDUCTOR LEMON
ACSR/GZ STRUNG AT 20%
CBL @ 15 °C, DISPLAYED
AT 85 °C.
2. OPGW & OHEW STRUNG
AT 18% CBL @ 15 °C.
DISPLAYED AT 50 °C.
3. GROUND CLEARANCE LINE
DISPLAYED @ 8.0 m.
4. THIS DRAWING SHALL BE
READ IN CONJUNCTION WITH
THE CONCEPT VEGETATION
CLEARING REPORT.
5. VEGETATION CLEARANCE
IS ASSESSED WITH
CONDUCTORS AT 500 Pa @
50 °C, AS WELL AS 85 °C NO
WIND.
6. REQUIRED VEGETATION
CLEARING IS SHOWN AS
RED MARKERS, AS WELL AS
RED CIRCLES AT STRUCTURE
LOCATIONS WITH DIAMETER
60 m.





PROJECT
HILLS OF GOLD WIND FARM

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CONSULTANT
AECOM

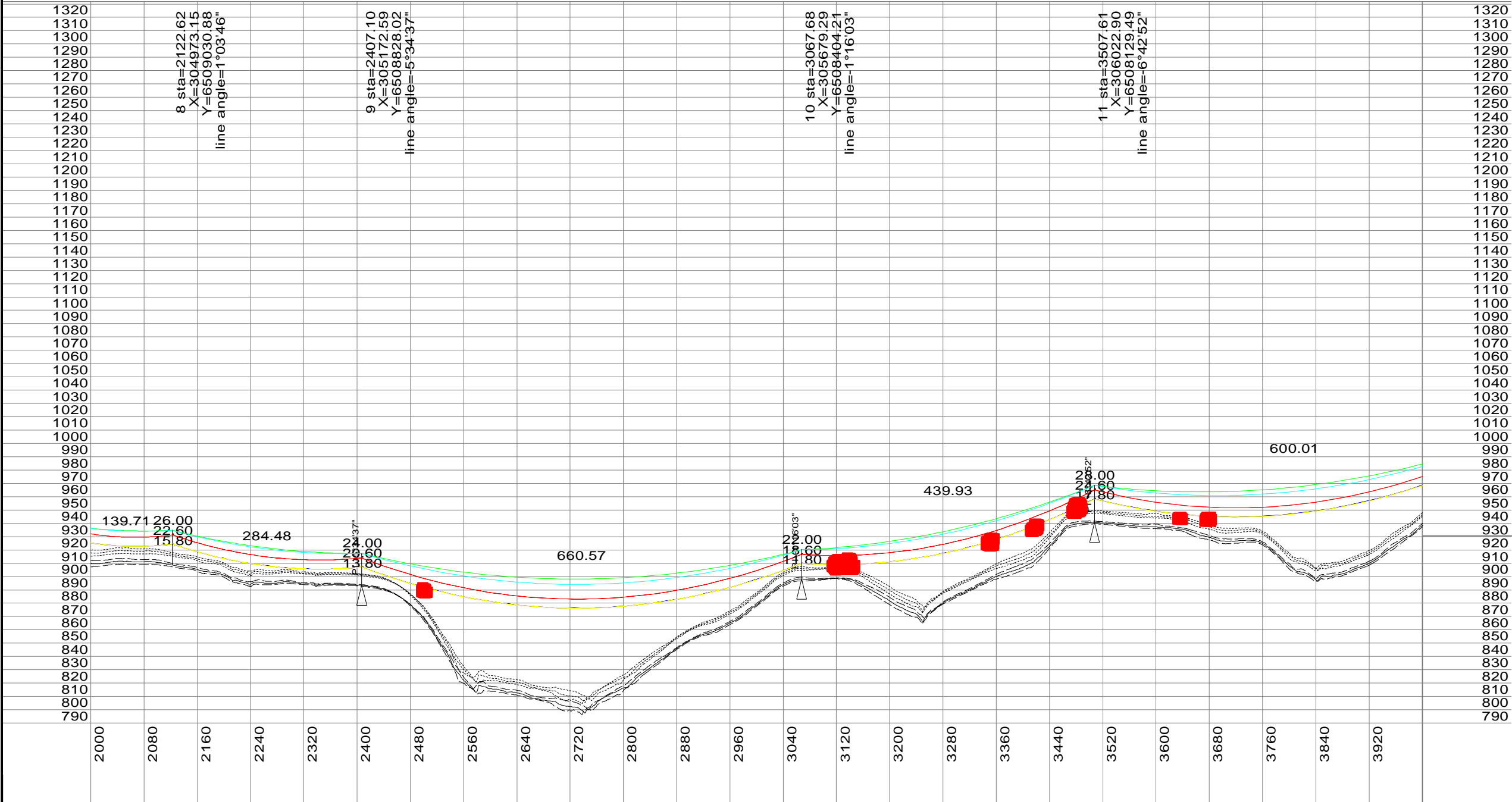
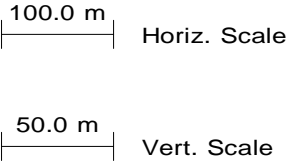
PROJECT NUMBER
60633842

SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 2 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
1. TWIN CONDUCTOR LEMON
ACSR/GZ STRUNG AT 20%
CBL @ 15 °C, DISPLAYED
AT 85 °C.
2. OPGW & OHEW STRUNG
AT 18% CBL @ 15 °C.
DISPLAYED AT 50 °C.
3. GROUND CLEARANCE LINE
DISPLAYED @ 8.0 m.
4. THIS DRAWING SHALL BE
READ IN CONJUNCTION WITH
THE CONCEPT VEGETATION
CLEARING REPORT.
5. VEGETATION CLEARANCE
IS ASSESSED WITH
CONDUCTORS AT 500 Pa @
50 °C, AS WELL AS 85 °C NO
WIND.
6. REQUIRED VEGETATION
CLEARING IS SHOWN AS
RED MARKERS, AS WELL AS
RED CIRCLES AT STRUCTURE
LOCATIONS WITH DIAMETER
60 m.



100.0 m Horiz. Scale

50.0 m Vert. Scale



PROJECT
HILLS OF GOLD WIND FARM

CLIENT

CONSULTANT
AECOM

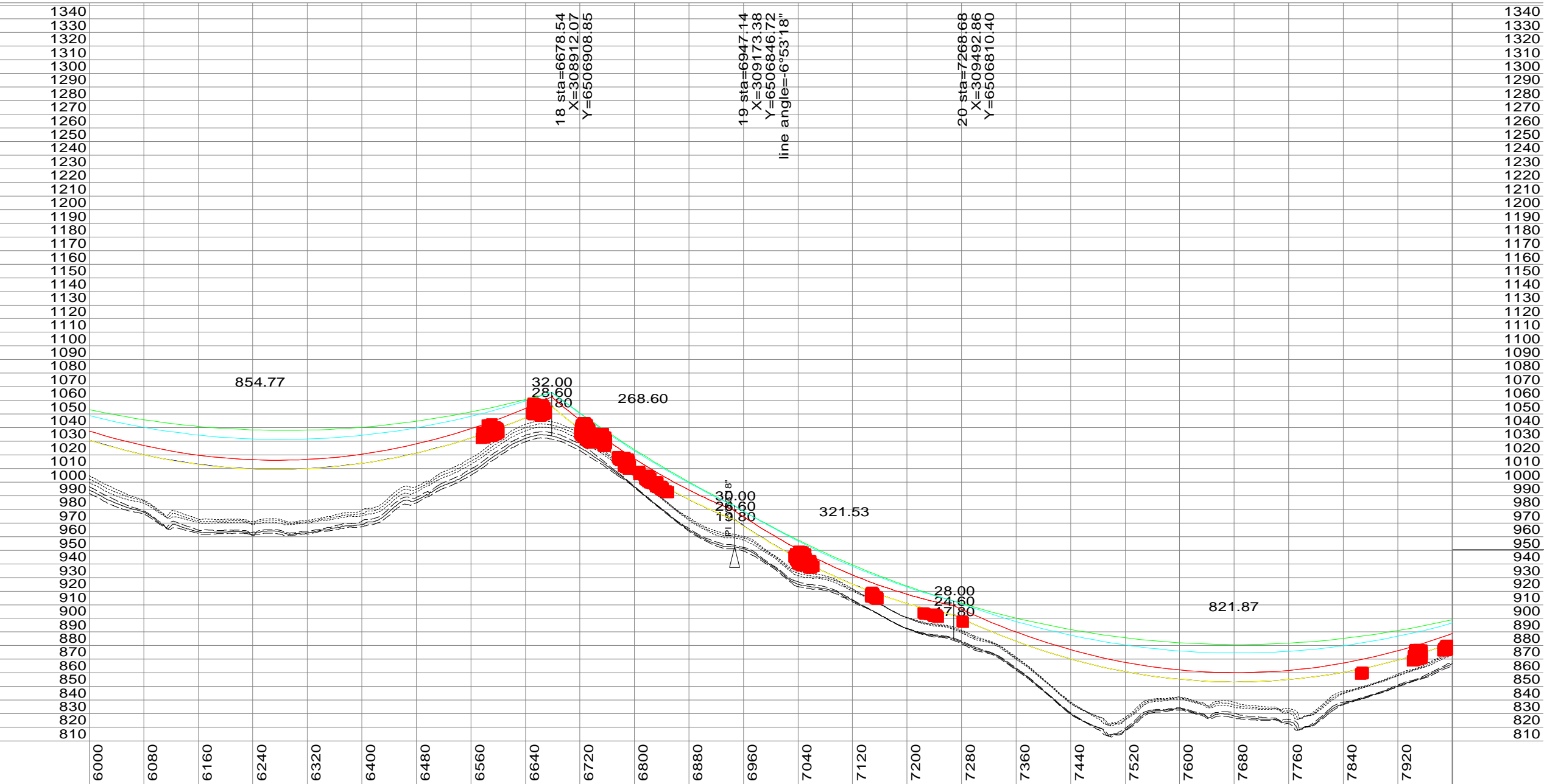
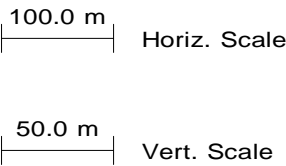
PROJECT NUMBER
60633842

SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 4 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
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LOCATIONS WITH DIAMETER
60 m.





PROJECT
HILLS OF GOLD WIND FARM

CLIENT

CONSULTANT
AECOM

PROJECT NUMBER
60633842

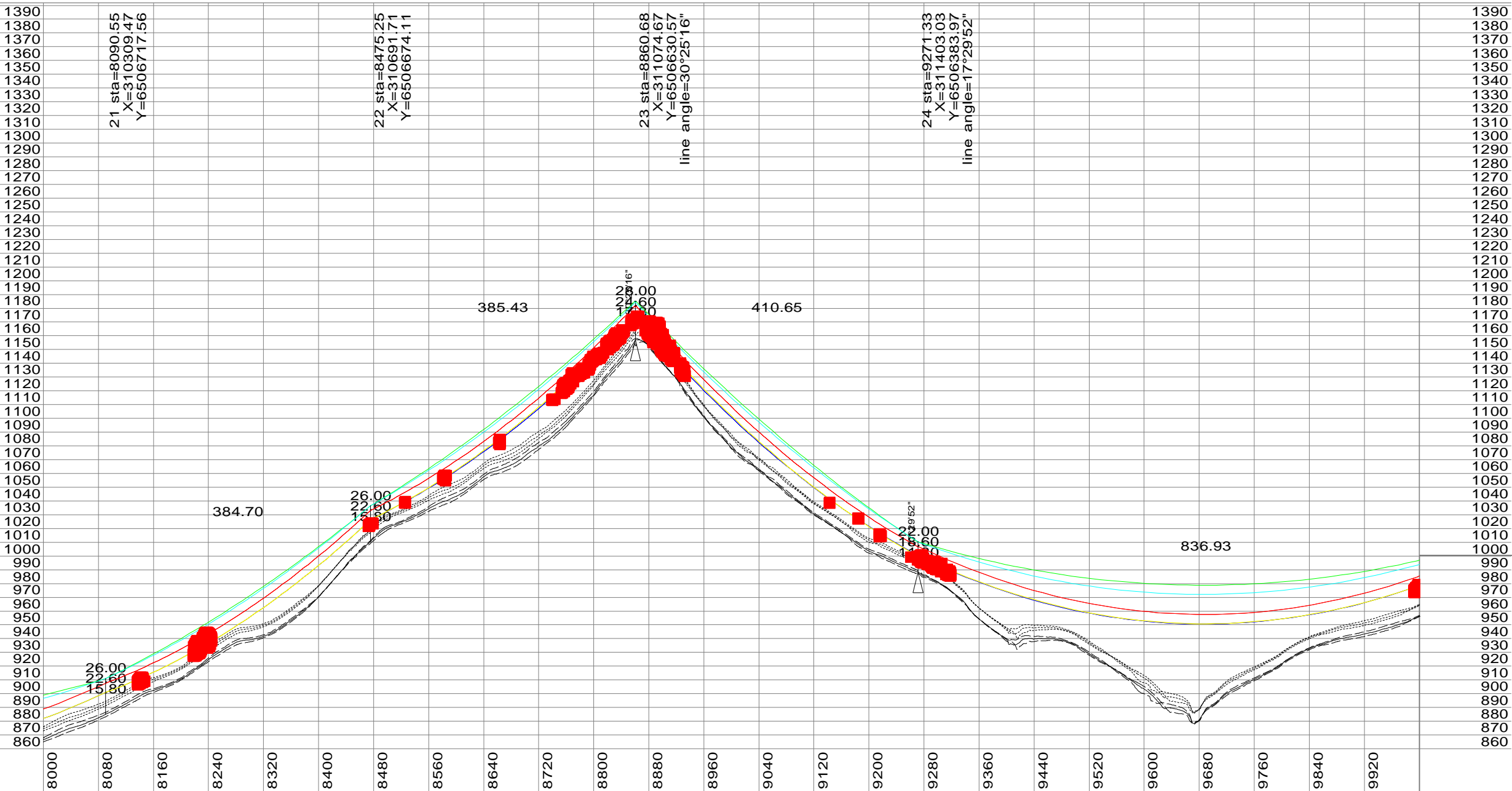
SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 5 OF 13

ISSUE/REVISION

0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

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LOCATIONS WITH DIAMETER
60 m.



100.0 m
Horiz. Scale

50.0 m
Vert. Scale



PROJECT
HILLS OF GOLD WIND FARM

CLIENT

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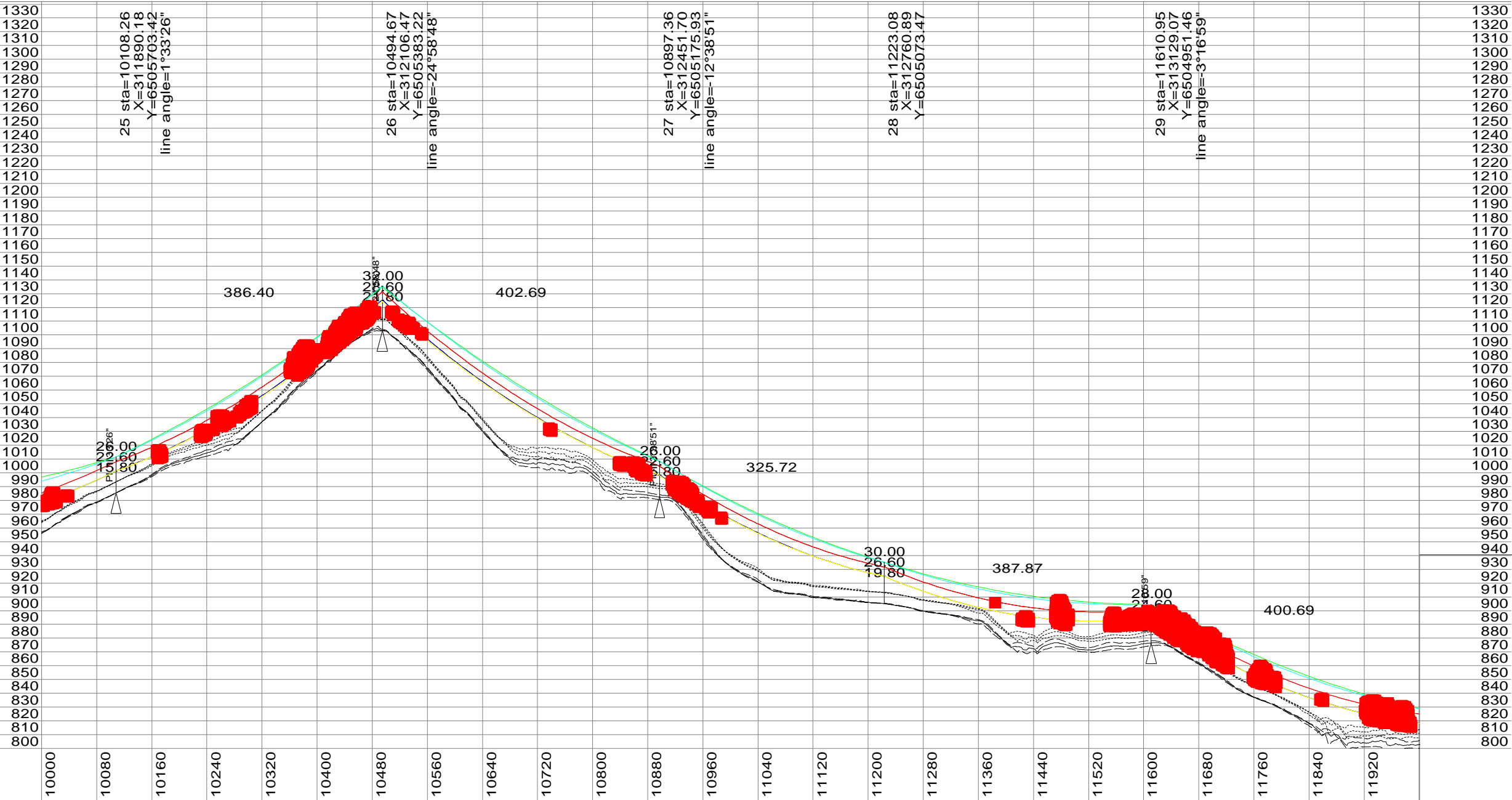
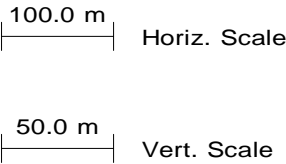
PROJECT NUMBER
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SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 6 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
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PROJECT
HILLS OF GOLD WIND FARM

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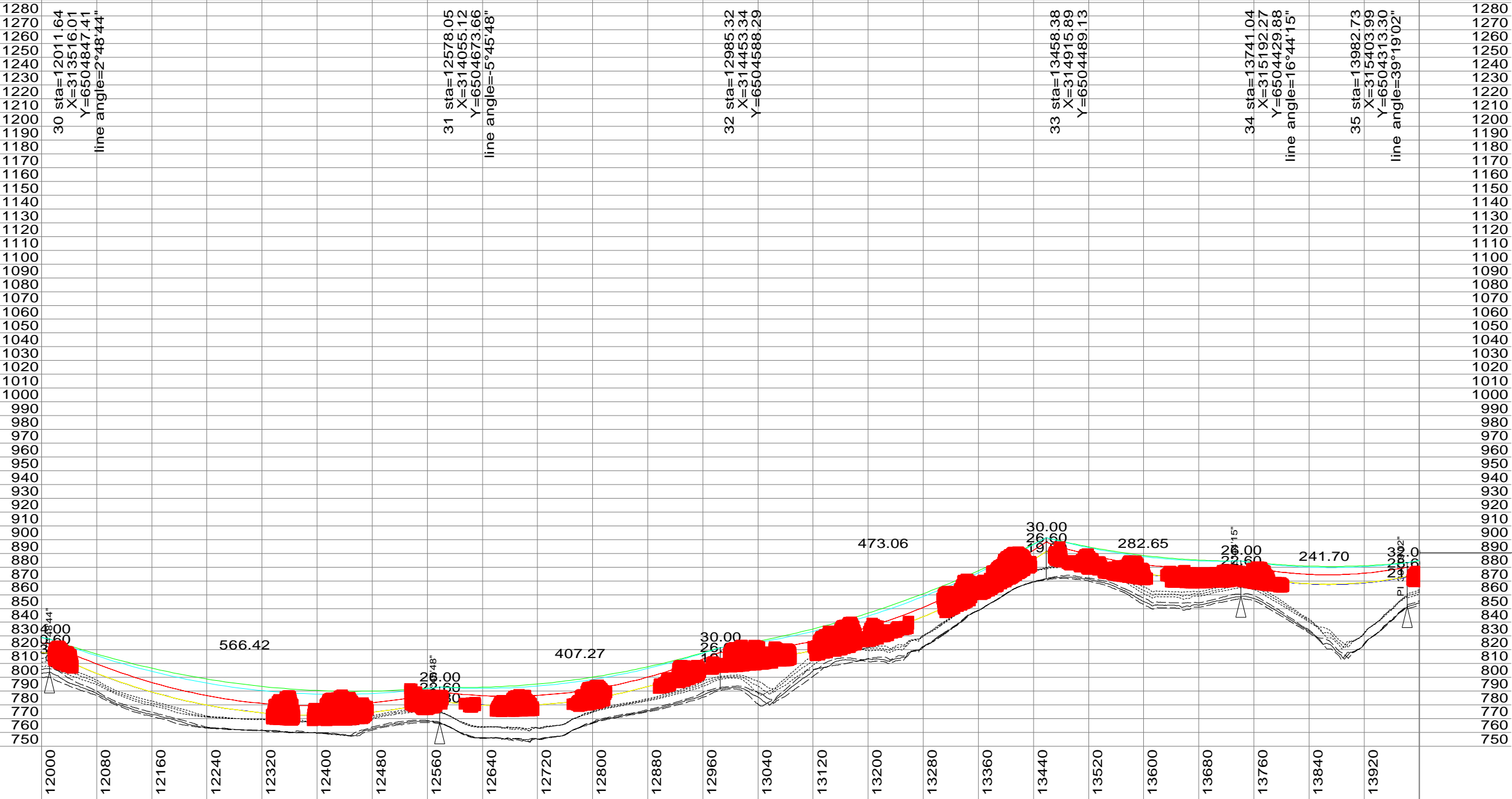
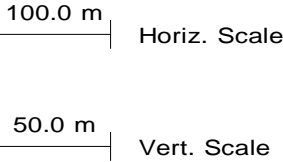
PROJECT NUMBER
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SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 7 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

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PROJECT
HILLS OF GOLD WIND FARM

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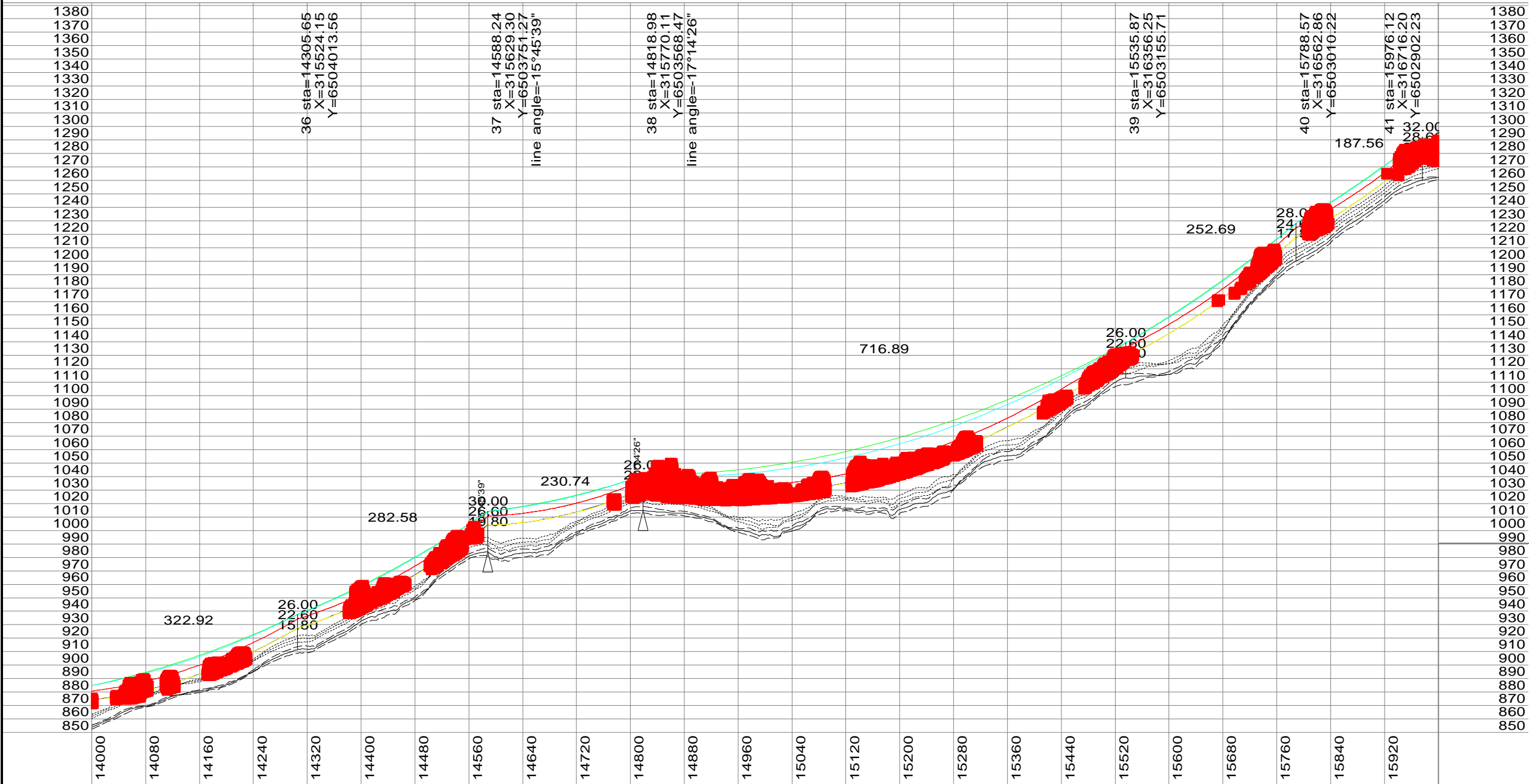
PROJECT NUMBER
60633842

SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 8 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

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100.0 m Horiz. Scale
50.0 m Vert. Scale



PROJECT
HILLS OF GOLD WIND FARM

CLIENT

CONSULTANT
AECOM

PROJECT NUMBER
60633842

SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 9 OF 13

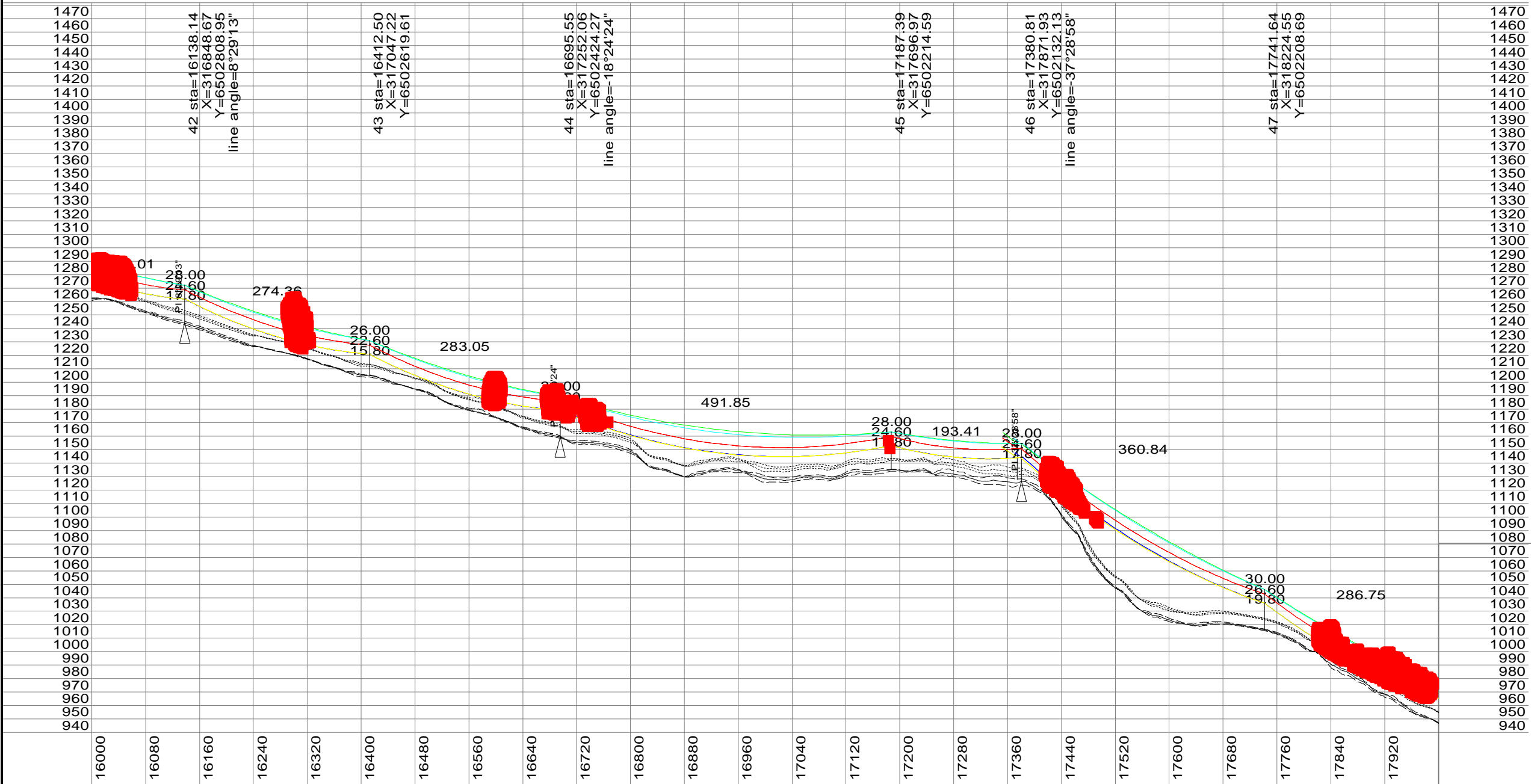
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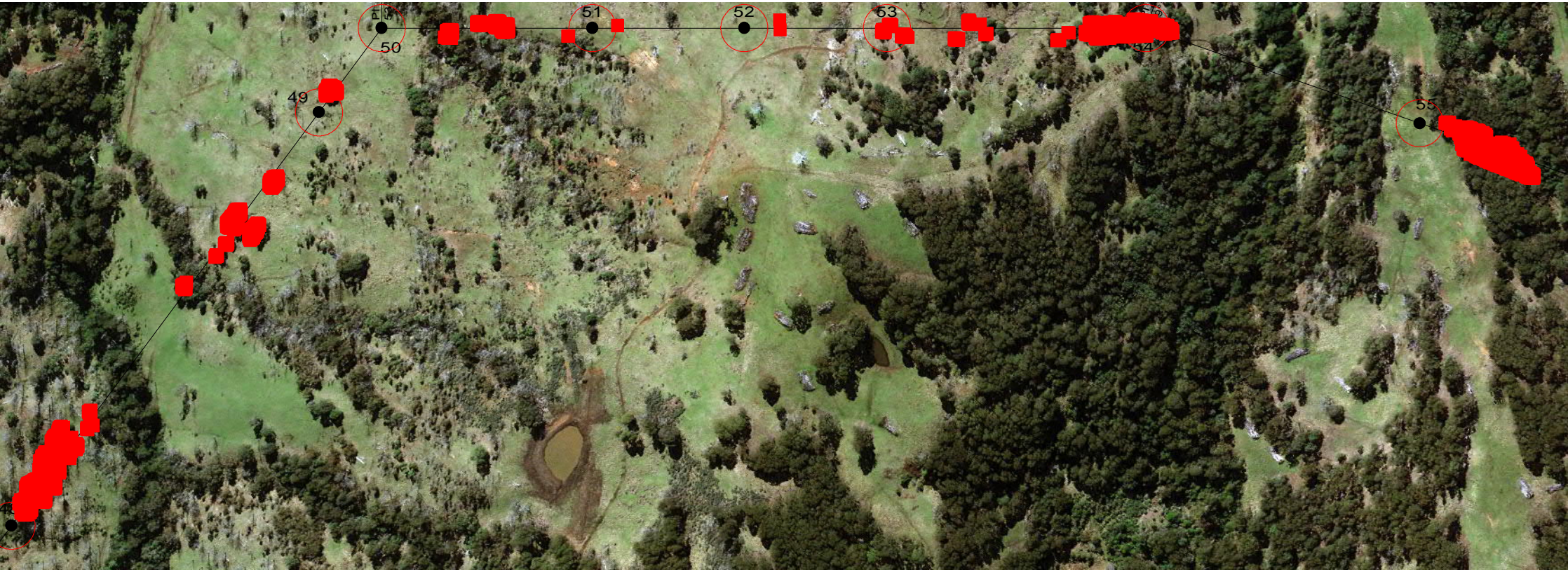
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
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60 m.

100.0 m Horiz. Scale

50.0 m Vert. Scale





PROJECT
HILLS OF GOLD WIND FARM

CLIENT

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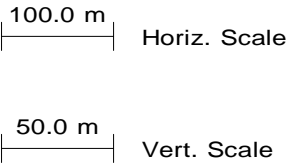
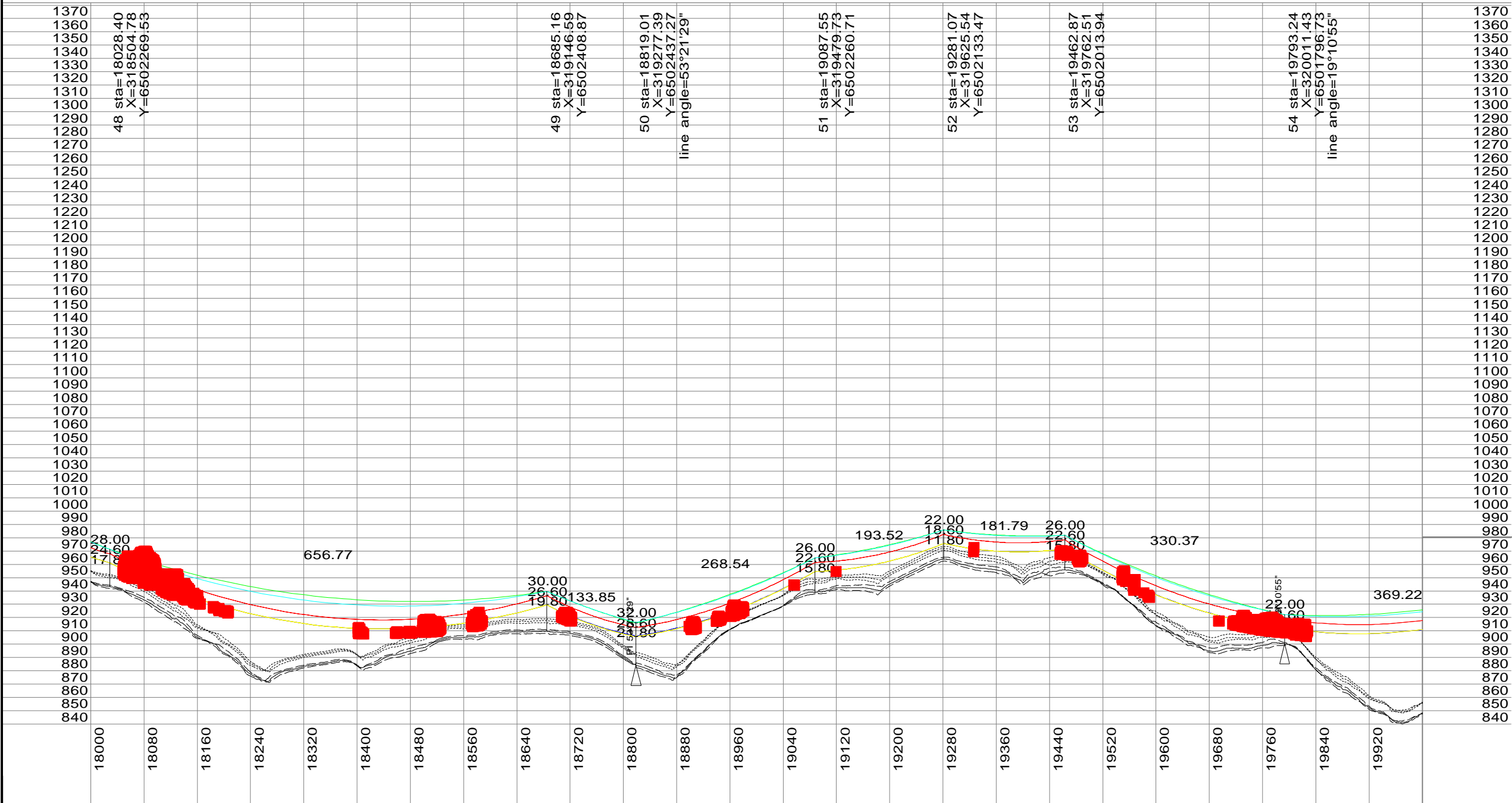
PROJECT NUMBER
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SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 10 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
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LOCATIONS WITH DIAMETER
60 m.





PROJECT
HILLS OF GOLD WIND FARM

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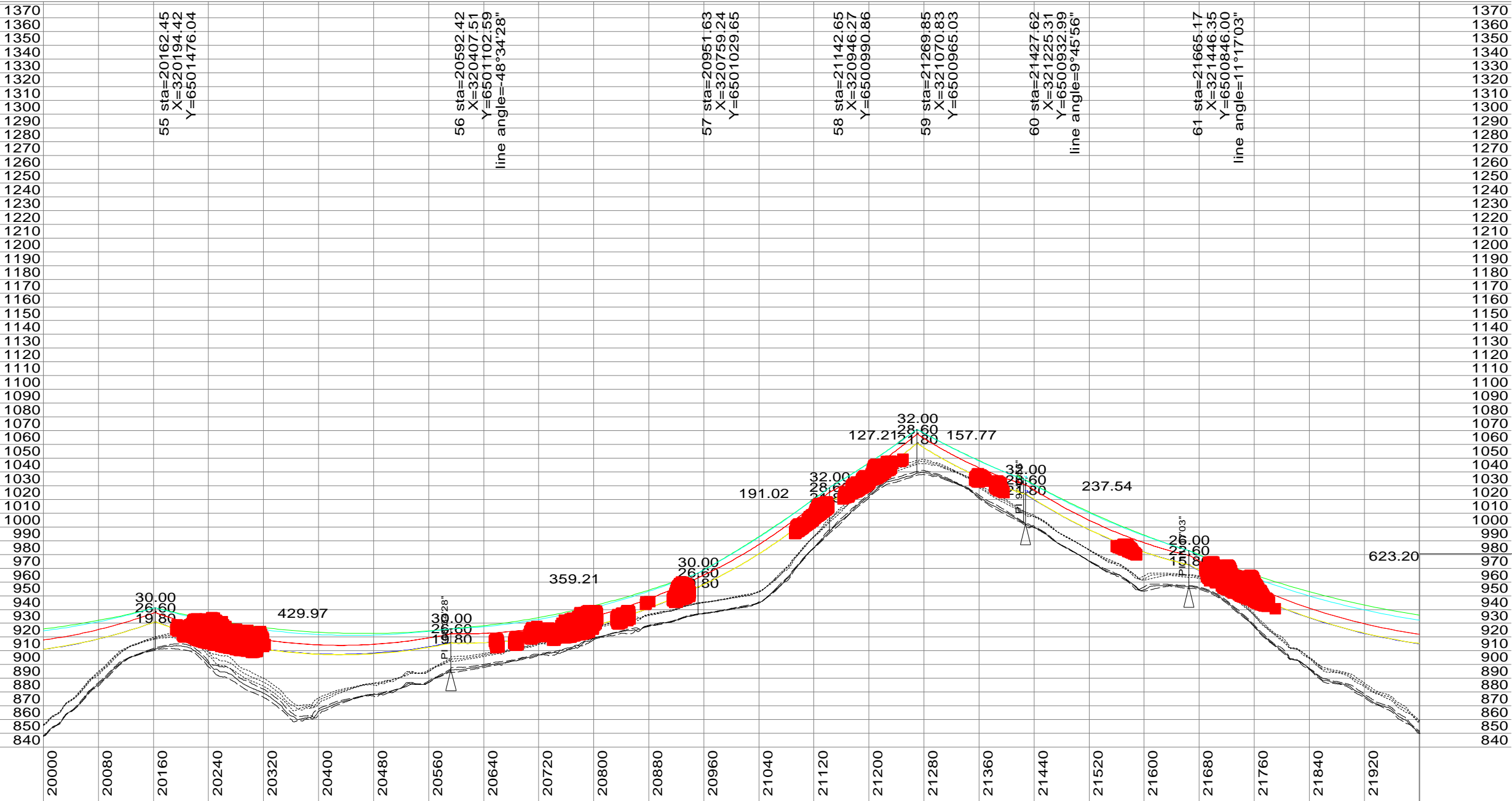
PROJECT NUMBER
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SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 11 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
1. TWIN CONDUCTOR LEMON
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60 m.



100.0 m Horiz. Scale
50.0 m Vert. Scale



PROJECT
HILLS OF GOLD WIND FARM

CLIENT

CONSULTANT
AECOM

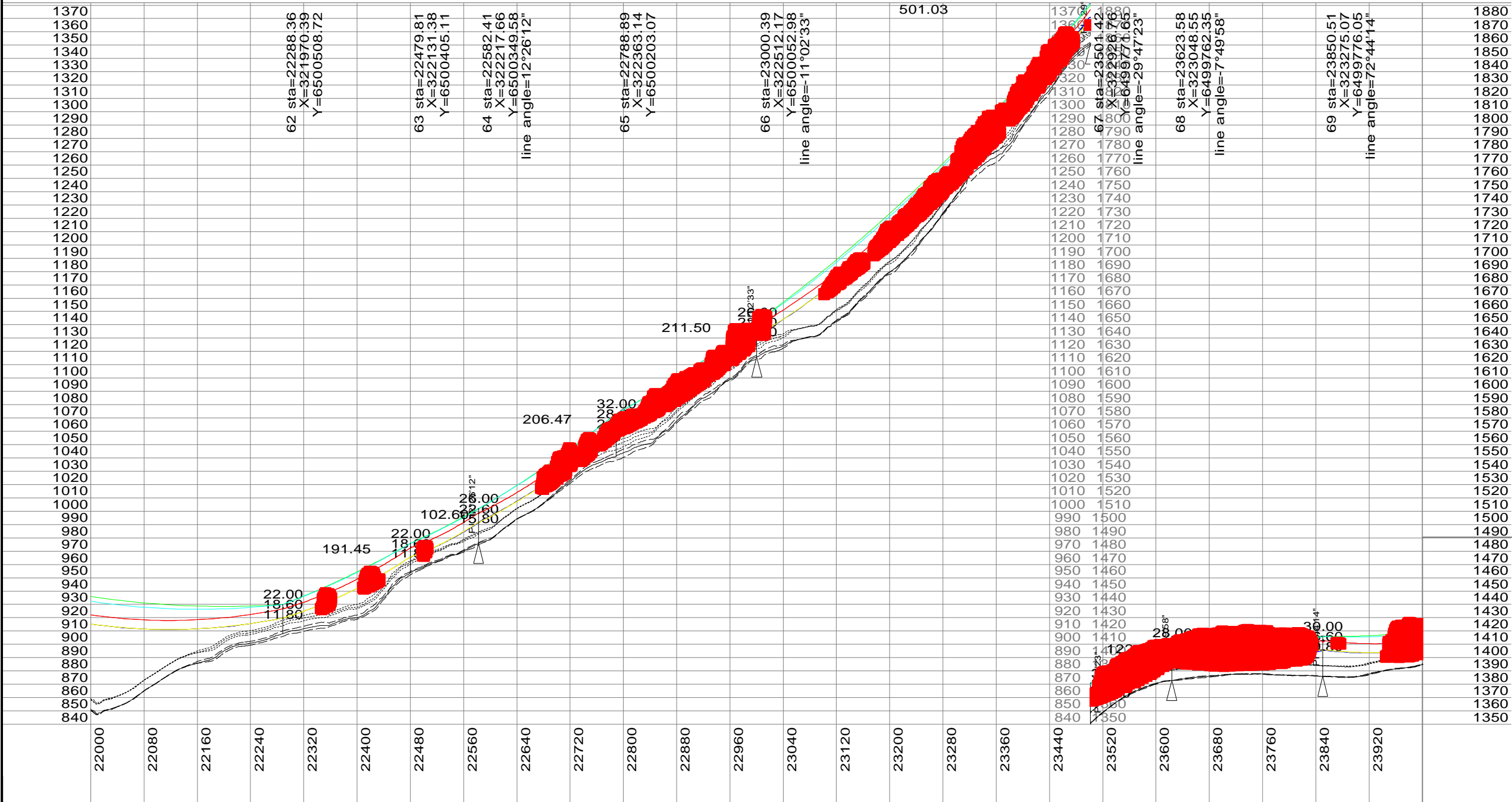
PROJECT NUMBER
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SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 12 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
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LOCATIONS WITH DIAMETER
60 m.



100.0 m Horiz. Scale
50.0 m Vert. Scale



PROJECT
HILLS OF GOLD WIND FARM

CLIENT

CONSULTANT
AECOM

PROJECT NUMBER
60633842

SHEET TITLE
VEGETATION CLEARING
330 kV CONCEPT DESIGN
LINE PROFILE

SHEET NUMBER
SHEET 13 OF 13

ISSUE/REVISION		
0	13/08/2021	CONCEPT VEGETATION CLEARING
I/R	DATE	DESCRIPTION

NOTES:
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